

1 **Scattered accumulation hotspots of different seafloor**  
2 **macro-litter categories: first guidelines for the**  
3 **optimization of mitigation actions**

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34 **Abstract**

35 Marine litter is an ever-increasing problem that demands immediate reduction plans and  
36 mitigation actions that should act synergically to efficiently meet ambitious goals. Since the seafloor  
37 has been recognized as a major sink for marine debris, the study of litter accumulation dynamics  
38 underlies future plans of removals. We analysed a 7 years (2013-2019) standardized data series  
39 collected along Sardinian fishing grounds through MEDiterranean International Trawl Survey, for  
40 which estimates of density and abundance of seafloor macro-litter were calculated over 707 hauls.  
41 Results emphasize the absence of any temporal trend in seafloor macro-litter abundance and density,  
42 but rather a spatial and bathymetric segregation of different litter categories. Our data showed how  
43 different sources and physical features of macro-litter items (*i.e.*, plastic and fishing gears, rubber,  
44 glass, metals and textile) led to spatially segregated accumulation hotspots, often at shallower depths  
45 and closer to coastlines, so that litter removal action could prioritize more accessible hotspots that  
46 could represent a solid benchmark for developing more challenging activities to be performed in  
47 deeper waters. We point here how the identification of seafloor macro-litter hotspots using aggregated  
48 data could be indeed be blurred by the dominance of plastic items and thus hide the presence of less  
49 abundant but yet detrimental litter categories accumulated in the marine environment.

50

51

52 **Keywords:** Marine litter; Single-use Plastic; Mitigation strategies; Seafloor macrolitter  
53 hotspots; litter removal.

54

## 55 1. Introduction

56 Poorly managed waste is contaminating the world's oceans and human-made marine litter is  
57 continuously entering the marine environment, with significant impacts on marine ecosystems (Canals  
58 et al., 2021; Chiba et al., 2018). It is estimated that plastic alone account for ca. 8 million tons of the  
59 solid waste that enters into the marine environment, constituting ca. 60-80% of the whole amount of  
60 marine litter (Jambeck et al., 2015), so that the terms plastic and marine litter are used almost as  
61 synonyms.

62 A global commitment aimed at reducing waste and plastic emissions into the environment is  
63 urgently needed, both from communities, authorities (*e.g.*, the United Nations Environment Assembly  
64 Resolutions Marine Litter and Microplastics and Goal 14.1 of the United Nations Sustainable  
65 Development Goals, among others), non-governmental organizations and several businesses that are  
66 generally promoting more sustainable lifestyles.

67 Activities of beach clean-up are efficient initiatives to raise awareness of the extent of litter  
68 pollution; however, the ocean seafloor is out of sight to most of society. As such, its vulnerability  
69 remains largely unknown to the general public (Canals et al., 2021). This perception does contrast  
70 evidence from recent scientific reports devoted to seafloor litter, which strongly supports the  
71 hypothesis by which the seafloor represents the ultimate sink for the great majority of items that enter  
72 into the marine environment (Cau et al., 2018; Peng et al., 2019; Woodall et al., 2014), especially  
73 plastic items, which include the well-known microplastics (1 $\mu$ m to 5 mm in size) (Frias and Nash,  
74 2019).

75 Within the European Union, Marine litter has been recognized as a major issue in European  
76 seas since the entering into force of the Marine Strategy Framework Directive (MSFD; 2008/56/EC),  
77 within which marine litter quantification is the 10<sup>th</sup> of 11 descriptors proposed for evaluating the  
78 environmental status of European seas (Galgani et al., 2013). The Mediterranean is among the worlds'  
79 oceans exhibiting the most considerable accumulation of marine litter (Eriksen et al., 2014; Pham et  
80 al., 2014; Ramirez-Llodra et al., 2011) and, indeed, thousands of items (up to 20,000 km<sup>-1</sup>;  
81 Pierdomenico et al., 2019) have been estimated to sit on its seafloor. However, these *scenarios* could  
82 be even more prominent in other areas, for which no data exist. The assessment of the abundance,

83 distribution, and effects of marine litter on the ocean floor is primarily challenged by our limited  
84 knowledge of this environment. Indeed, studies carried out so far have investigated benthic litter  
85 abundance, composition, and distribution mainly on continental shelves and slope, whereas data from  
86 deeper habitats such as adjacent bathyal plains (Bergmann and Klages, 2012; Galgani et al., 2000,  
87 1996; Pham et al., 2014; Ramirez-Llodra et al., 2013) are far less abundant, primarily due to the  
88 difficulties and associated costs of deep-sea sampling (Barnes et al., 2009).

89 As unfortunately reported, the concept of seas litter-free has proved to be utopian, while more  
90 realistic targets of reduction are foreseen, only if multiple mitigation measures will act synergically  
91 against the global plastic problem (Borelle et al., 2020; Lau et al., 2020); one of these is to potentially  
92 reduce harm by removing marine litter that is already in the ocean (Rochman, 2016). In the past  
93 decade, a remarkable effort has been put into the identification areas of the sea where floating plastic  
94 and microplastic removal could be more effective (Sherman and Van Sebille, 2016). However, this  
95 sort of effort doesn't have a 'benthic' counterpart, despite the availability of numerous studies dealing  
96 with the identification of seafloor litter hotspots (Franceschini et al., 2019; Garofalo et al., 2020).

97 The distribution, abundance, and composition of benthic litter in the Mediterranean Sea have  
98 patchy documentation, further scattered by the use of different sampling techniques. Over the last  
99 decade, the necessity to standardize protocols over wide geographical scales and assess the baseline of  
100 seafloor litter's abundance and composition has been recognized. These information were identified as  
101 vital to implement litter reduction policies and adequate monitoring schemes. While recognizing a  
102 still limited knowledge on the topic, much effort has been put into assessing these two  
103 abovementioned points (Spedicato et al., 2019).

104 What could be considered the 'second step' of the process is the evaluation of multiple years,  
105 standardized data series, which are pretty rare (Canals et al., 2021; Galgani et al., 2021; Parga  
106 Martínez et al., 2020). Indeed, it has been recently documented how difficult it is to provide a solid  
107 answer to whether seafloor macro-litter amount into marine environments is increasing or not due to  
108 the scatteredness of data and the lack of continuous standardized surveys (Galgani et al., 2021). These

109 kinds of data would allow to identify seafloor macro-litter 'hotspots', which are defined as those areas  
110 where macro-litter accumulates (Tubau et al., 2015) and, pragmatically, where eventual mitigation  
111 actions should be prioritized.

112 In this perspective, this work aims at analyzing the spatial and temporal variability of seafloor  
113 litter in the Food and Agriculture Organization (FAO) Geographical Sub-Area 11 (Sardinian seas,  
114 western Mediterranean), taking advantage of the MEDiterranean International Trawl Survey  
115 (MEDITS) (Spedicato et al., 2020) standardized protocol of surveys conducted for 7 years (from 2013  
116 to 2019).

117 The present work also aims to test whether different litter categories (according to the MEDITS  
118 protocol) may result in geographically segregated seafloor litter hotspots, thus providing useful  
119 guidelines for prioritizing remedial action foreseen in the framework of plastic reduction policies.

120

121

## 122 **2. Materials and methods**

### 123 **2.1. Data collection**

124 Data used in the present study were collected in the MEDITS survey framework conducted from 2013  
125 to 2019 in the FAO GSA 11. Spatial patterns of data collected between 2013 and 2015 have been  
126 described in Alvito et al. (2018), while the present study focuses on the entire temporal range between  
127 2013 and 2019, including the analysis of accumulation hotspots.

128 A total of 707 hauls, pre-positioned following a random stratified sampling design, were carried out,  
129 from a minimum of 100 to a maximum of 102 hauls year<sup>-1</sup>(see Tab. 1), distributed over a depth range  
130 comprised between 10 and 800m, grouped according to five bathymetrical *strata*, namely: A (0-50m);  
131 B (51-100m); C (101-200); D (201-500) and E (501-800) and seven zone, comprised within GSA11  
132 (see Fig. 1).

133 Once onboard, seafloor macro-litter data was classified according to nine major categories (i.e., L1:  
134 Plastic; L2: Rubber; L3: Metal; L4: Glass/Concrete; L5: Clothes; L6: Processed wood; L7: Paper and  
135 cardboard; L8: Other; L9: Unspecified), and relative sub-categories (see Alvito et al., 2018; Spedicato  
136 et al., 2019 for more details). To obtain standardized abundance and density indices, the total weight

137 and the number of items collected per each sub-category were standardized according to the swept area,  
138 expressed in km<sup>2</sup>.

139

## 140 **2.2 Data processing and statistical analysis**

141 To test for significant spatio-temporal differences in seafloor litter density (items km<sup>-2</sup>) and  
142 abundance (kg km<sup>-2</sup>), a non-parametric permutational analysis of variance (PERMANOVA; software  
143 PRIMER 7+) was performed. The statistical routine was based on Euclidean distance-based  
144 resemblance matrixes, built used untransformed data on density and abundance. The PERMANOVA  
145 was ran in Uni- and Multivariate environments using the total density or abundance of seafloor  
146 macro-litter and its composition, respectively. The analysis was conducted using a design based on  
147 three fixed orthogonal factors and their interactions; in detail: i) Year (7 levels); ii) Zone (7 levels);  
148 iii) Depth (5 levels), with a variable number of replicates. Differences in seafloor macro-litter  
149 abundance across years were also tested singularly using the five most representative macro-litter  
150 categories (*i.e.*, from L1 to L5) density and abundance data.

## 151 **2.3 Seafloor macro-litter hotspots**

152 In order to highlight seabed hotspots of litter accumulation, abundance data from the MEDITS  
153 survey were interpolated, and then areas of high densities were selected. For each category,  
154 interpolated values were evaluated by 2D interpolating via Inverse Distance Weighting (IDW)  
155 (Falivene et al., 2010) on a grid of 1x1 km, thereby obtaining an extended map of the potential zones  
156 where plastic litter tends to sink and accumulate according to the historical series. Input variables  
157 taken into account for interpolation were longitude, latitude, and depth, while estimation errors were  
158 assessed using cross-validation techniques (CV) (Wise, 2011). *Hotspots* of recurrent litter  
159 accumulation were evidenced by selecting the 90<sup>th</sup> percentile of the interpolated distribution.

160

## 161 **3. Results**

162 The total swept area covered by the survey was ca. 44 km<sup>2</sup>. Among the 707 hauls conducted between  
163 2013 and 2019, 246 showed the absence of seafloor macro-litter (35%). More in detail, the lowest

164 number of macro-litter free hauls (18) was recorded in 2015, corresponding to 17% of the hauls  
165 conducted that year; on the contrary, the highest (45) was recorded in 2016 (44% of the year' hauls).  
166 Overall, a raw total of 1908 items were collected from 2013 to 2019, corresponding to a total of 531 kg  
167 of seafloor litter (Fig. 3).

168

### 169 **3.1. Spatio-temporal trends**

170 During the study period, the average number of collected items was  $107.3 \pm 30$  items  $\text{km}^{-2}$ ,  
171 with a maximum of  $151.1 \pm 44.8$  items  $\text{km}^{-2}$  (average  $\pm$  standard error) in 2014 and a minimum of  $70.8$   
172  $\pm 26.4$  items  $\text{km}^{-2}$  recorded in 2016. With respect to the amount of seafloor macro-litter, the average  
173 abundance recorded in the study period was  $35.06 \pm 20.8$  kg  $\text{km}^{-2}$ , with a maximum of  $85.4 \pm 66.9$  kg  
174  $\text{km}^{-2}$  in 2019 and a minimum of  $10 \pm 2.6$  kg  $\text{km}^{-2}$  in 2013.

175 The highest macro-litter density was recorded in zones 6 and 7 (Fig. 1), showing an average  
176 density of  $180 \pm 39.5$  and  $140 \pm 23.7$  items  $\text{km}^{-2}$ , respectively, while the lowest density was in zone 1  
177 ( $32 \pm 5.5$  items  $\text{km}^{-2}$ ).

178 According to the PERMANOVA routine, there was a lack of significant differences in  
179 seafloor macro-litter density, abundance and composition across years (Tab. 2). The absence of  
180 significance was also confirmed when each of the 5 major categories of macro-litter was considered  
181 singularly.

182 While the interaction of the three factors considered simultaneously (*i.e.* 'year', '*strata*' and  
183 'zone') was not significant, the interaction of the factors '*strata*' and 'zone' was statistically significant  
184 ( $p < 0.001$ ; Tab. 2), emphasizing a significant variability in the spatial and bathymetric distribution and  
185 composition of seafloor macro-litter.

186

### 187 **3.2. Seafloor macro-litter composition**

188 Concerning the seafloor macro-litter composition in terms of abundance, plastic was the most  
189 abundant component, with an average value of 82% across the investigated period (2013-2019). The  
190 highest relative abundance was measured in 2019 with 94%, followed by 2014 and 2017, where plastic

191 accounted for 90% and 86% of the total amount of seafloor macro-litter, respectively (Fig. 2). Despite  
192 the absence of a temporal trend, the macro-category of plastic that includes Derelict Fishing Gears was  
193 consistently the dominant category in relative abundance, followed by rubber, metal and others, that  
194 showed an average percentage of ca. 4% and clothes (3%) (Fig. 2). The mean relative abundance of  
195 plastic collected from the seven different zones of the Sardinian grounds was 82%, followed by rubber,  
196 metal, and 'other', all accounting for ca. 4%. The highest values were observed in zone 4, where plastic  
197 items accounted for 96% of the total, followed by zone 5 (94%) and zone 6 (72%) (Supplementary Fig.  
198 1). Concerning the relative abundance in weight calculated for all the five bathymetric *strata* (A-E),  
199 plastic occupied the major portion with 82%, followed by metal, glass, and clothes with ca. 4%. The  
200 deepest stratum (E) and stratum (C) generally showed the highest values for weight abundance with  
201 85% of plastic debris followed by the shallowest strata (i.e., A and B), showing values of 84% of plastic  
202 debris each. The lowest value for plastic was observed in stratum (D); which showed a higher relative  
203 abundance of 'Other' category and Clothes (26% and 13%, respectively; Supplementary Fig. 2).

204 With respect to the density of items collected throughout the study period, plastic debris  
205 represented the dominant category, which was on average accounting for ca. 84% of the total number  
206 of items (Fig. 2). The highest record was observed in 2019 (90%), followed by 2017, 2018 and 2016,  
207 where items belonging to the plastic category accounted for 88%, 87%, and 86%, respectively (Fig. 2).  
208 As per the relative abundance in terms of weight, plastic remained the dominant category over metal,  
209 glass & clothes, which showed a mean density percentage of 4%. The spatial analysis revealed how  
210 collected items were numerically dominated by the plastic category, with a maximum value of 93%  
211 recorded for zone 5, followed by zone 6 (88%) and zone 4 (83%). The mean density was 84% across  
212 all zones, followed by metal, glass, and clothes, accounting for 4% (Supplementary Fig. 3). The average  
213 relative density calculated for the study period in all the five bathymetric *strata* (A-E) showed how  
214 plastic constituted 84% of collected material, followed by metal, glass, and clothes, all accounting for  
215 4%. The remaining categories accounted cumulatively for 4%. *Stratum* (B) showed the highest values  
216 of plastic macro-litter (89%); followed by *strata* E and A, where plastic always accounted for more than  
217 80% (i.e., 87% and 82%, respectively); whereas stratum (C) recorded the lowest values of 74% plastic  
218 debris (Supplementary Fig. 4). The second most representative category of macro-litter were Glass and



219 Clothes, with variable percentages through different bathymetric *strata*, ranging between 2 and 12%  
220 and between 3 and 7% respectively (Supplementary Fig. 4).

221

### 222 **3.3. Seafloor macro-litter hotspot**

223 Seafloor macro-litter hotspots for the whole set of categories and segregated for each category are  
224 shown in Fig. 4. Results showed that for the Plastic category, which was split in fishing-related (e.g.,  
225 fishing nets, lines, etc.) and not fishing-related (all other plastic items), major concentrations of items  
226 recur in the south-west area of the island. Hotspots for rubber items showed a reasonably homogeneous  
227 distribution around the Sardinia coast, with the highest peaks in the south-eastern area. Similar values  
228 were found for the Metal debris. Glass category showed a consistent hotspot in the north-western area  
229 of the island, while the highest values of fiber and clothes are more likely to occur in the south-eastern  
230 area (Fig. 4).

231

## 232 **4. Discussions**

233 Environmental recovery (*i.e.*, clean-up actions and restoration of highly contaminated sites) is  
234 recognized as one of the mitigation strategies that should act synergically with other measures to  
235 reduce seafloor litter and plastic contamination in the future (Borelle et al., 2020). Removal of litter  
236 contamination is desirable, but in many cases, it lacks a detailed consideration of the possible  
237 effectiveness, costs, and benefits of such solutions and without acknowledging a basic lack of data  
238 (Falk-Andersson et al., 2020). Such strategies will surely be most effective when guided by best-  
239 available science (Rochman, 2016). What is the best science on the topic is a still-opened debate that  
240 will unlike find a proper answer in the near future; however, trying to identify the *best-available*  
241 science is a more affordable task. In this perspective, standardized data that take advantage of well-  
242 established protocols currently represent the most reliable option to study seafloor macro-litter's  
243 temporal trends (Spedicato et al., 2019), while acknowledging the intrinsic drawback that campaigns  
244 designed for other purposes might present (Canals et al., 2021).

245 Based on regular surveys conducted over a 7 years' time series, our results confirm the absence  
246 of a significant temporal trend in seafloor macro-litter quantities, regardless of the litter category, as

247 already observed and commented in a recent review proposed by Galgani et al. (2021). The absence of  
248 any trend is likely related to the oscillatory pattern of plastic macro-litter, which represents the most  
249 conspicuous fraction of collected items (*i.e.*, always >60% over the study period, both in terms of  
250 density and abundance per km<sup>2</sup>), regardless of the geographical area or the bathymetric *stratum*  
251 considered.

252         The plastic fraction is of particular concern due to the detrimental effects that its persistence in  
253 the environment can possibly trigger, such as: i) the introduction of artificial substrates that can locally  
254 enhance benthic biodiversity (the so-called 'plastic benefits paradox'; Carugati et al., 2021; Song et al.,  
255 2021); ii) the accidental ingestion and possible fragmentation into smaller particles mediated by *biota*  
256 that can ideally become more and more bioavailable to multiple levels of the trophic web (Cau et al.,  
257 2020, 2019a; Courtene-Jones et al., 2019; Fossi et al., 2018); iii) compromise the biogeochemistry and  
258 alter microbial communities in sediments (Seeley et al., 2020); iv) provide physical damage to  
259 structuring fauna, in case of Derelict Fishing Gears (Angiolillo and Fortibuoni, 2020; Cau et al., 2017;  
260 Galgani et al., 2018). Nonetheless, beside the enormous attention of the scientific community on the  
261 plastic problem, other undocumented yet detrimental effects of other litter categories to the ecosystem  
262 could likely be discovered.

263         The observation of a decreasing trend can be possibly observed for few specific types of plastics  
264 subject to societal reduction measures (e.g., certain categories of single-use plastic), even though the  
265 time required might be within decades (Maes et al., 2019); on the contrary, other sub-categories of  
266 plastic items (e.g., fishing gears) do not follow the same trend. However, the urgency to remove seafloor  
267 macro-litter and consequently plastic already accumulated in the marine environment for long periods  
268 (Cau et al., 2019b) is crucial to reduce the detrimental effects mentioned above.

269         Fishing For Litter (FFL) initiatives are among the most effective strategies for removing  
270 seafloor-litter from oceans, although many difficulties arise in their execution and implementation  
271 (Ronchi et al., 2018); this is due to the wide spatial coverage of trawling activities, often in remote areas  
272 where fishermen are the only operators capable of providing this type of social service, nonetheless  
273 during their regular working routine.

274 Focusing specifically on DFGs, their removal appears very demanding in terms of tools and  
275 skills required to perform such remedial action, which additional difficulties to the low accessibility of  
276 sea-bottom environment below certain depths. Bottom trawling vessels using hooks and ropes are the  
277 cheapest way for removing DFGs, but it strongly impacts mechanically the seabed (Cho, 2011).  
278 Additionally, it could also be ineffective since nets and ropes are often lost due to their entanglement in  
279 small geomorphologies of the seabed or structuring communities (Angiolillo et al., 2015), that would  
280 be likely ripped out. In order to tackle this issue, a synergic, sustainable action that might involve diverse  
281 parties and technologies (such as professional divers and remote sensing) should be adopted (Cho, 2011;  
282 Madricardo et al., 2020; Morishige and McElwee, 2012).

283 The study of seafloor macro-litter as a whole could be blurring if the aim is identifying areas  
284 where mitigation strategies should be prioritized. The accumulation hotspot analysis, performed  
285 separately per each category identified in the MEDITS protocol, showed segregated distribution for  
286 the 5 major categories investigated (Fig. 3). This segregation pattern is ascribable to the macro-litter's  
287 intrinsic features, which, depending on material and dimensions, can float on the sea surface before  
288 sinking due to changes in specific weight enhanced by biofouling, or reach the seafloor directly once  
289 discarded. While the former case is the one that mostly relates to plastic items (Ryan, 2015), the latter  
290 relates the most with heavier objects such as rubber, glass, metal. Heavier objects hotspots were  
291 generally closer to the coastline, mostly in the proximity of the larger ports (Cagliari, Oristano and  
292 Olbia) or, in the case of glass objects, far from the coast but close to the most trafficked maritime  
293 routes such as the one those across the Gulf of Bonifacio or the port of Olbia, which suggests how this  
294 sort of object can possibly result from local dumping activities.

295 As recently reported (Scotti et al., 2021), according to available information extrapolated by  
296 Maritime State Property Concessions (Italian Ministry of Infrastructure and Transports), the hotspot  
297 area for glass the seafloor category could be related to both intense tourism and shipping activity that  
298 occurs most intensely along the north-western coastline. Also, hydrodynamic drivers such as the  
299 Northern Tyrrhenian Cyclone (east of the Strait of Bonifacio), could be invoked as a driver of local  
300 accumulation pattern by predictive model studies (Mansui et al., 2020). Similarly, the highest density  
301 of plastic litter found in the southwestern waters, could be related not only to the inefficient waste

302 management of the coastline, but even to the mesoscale instabilities belonging to the Algerian  
303 currents, with the latter being one of the most evident feature of swift currents (Zambianchi et al.,  
304 2017).

305 Plastic objects can be transported and accumulated far from the sources and over long  
306 distances following marine currents and in proximity of peculiar geomorphologies of the seafloor  
307 (Galgani et al., 2000; Spedicato et al., 2019). Thus, the distribution of plastic hotspots resulted in few  
308 delimited areas, mainly on the northern, western, and southern coasts, while they were absent on the  
309 eastern shore. The western coast, in particular, is one of the most dynamic areas of the Mediterranean  
310 Sea (Olita et al., 2013). It is characterized by strong northwest winds (Mistral and Tramontane) and by  
311 the presence of a southeastward current coming from the Balearic island, that around 40°N separates  
312 into two branches flowing northward and southward when approaching the coast (Hernandez-  
313 Lasheras and Mourre, 2018). The northward flow branch may be one of the factors that enhanced the  
314 accumulation of plastic and rubber categories in the gulf of Alguer (north-western corner), together  
315 with high tourism activities to which this area is subject especially in the summer; while the  
316 southward branch, also known as the Western Sardinian Current (WSC), might be one of the leading  
317 transport pathways contributing to the accumulation of plastic objects, but also of all the other  
318 categories, in the south-western corner. In this area, the presence of two minor Sardinian islands,  
319 Carloforte and Sant'Antioco, represents a physical barrier to the flow of WSC at its maximum  
320 intensity (Olita et al., 2013), favoring the formation of macro-litter hotspots washed up from the  
321 western continental shelf. Additionally, the presence of mesoscale anticyclonic eddies observed in this  
322 area (Knoll et al., 2017) could be another hydrological contributor to the accumulation of seafloor  
323 litter as they could facilitate the collection and sinking of floating macro-litter to the sea bottom  
324 (Spedicato et al., 2019).

325 Coupled with this highly dynamic hydrological *scenario*, the geomorphology of the seafloor  
326 is another critical feature that acts synergically to the spatial distribution of seafloor macro-litter  
327 hotspots (Galgani et al., 2000). Along the western Sardinian coast, the presence of a large continental  
328 shelf, characterized by extensive areas of outcropping and sub-outcropping rocky volcanic substrates  
329 (Conforti et al., 2016; Deiana et al., 2021), could possibly facilitate the entangling and thus the

330 accumulation of drifting objects on the seafloor. In contrast, along the eastern Sardinian coast, a short  
331 continental shelf, incised by numerous submarine canyons with their heads at a short distance from  
332 the shoreline (Harris and Whiteway, 2011), favors the transport of seafloor macro-litter, especially for  
333 lighter objects such as plastic items, towards deeper sea bottoms (Cau et al., 2017; Tubau et al., 2015;  
334 Zhong and Peng, 2021), explaining the absence of plastic hotspots along the eastern continental shelf.

335 Land-sourced macro-litter can be a major fraction of the total litter retrieved, which may flow  
336 from distant inland sources (Meijer et al., 2021; Pierdomenico et al., 2019). Moreover, such input can  
337 be exacerbated by heavy rainfalls, river floods, sewage overflow, or can carry large amounts of debris  
338 to beaches and coastal waters in a matter of a few hours or days, part of which subsequently spreads  
339 seawards and settles to the seafloor (Canals et al., 2021). Four of the largest Sardinian's rivers (i.e.,  
340 Flumini Mannu, Rio Mannu, Coghinas and Tirso; Alvito et al., 2018) discharges their water, together  
341 with all the land-based litter accumulated along their path, in proximity of highlighted hotspots in the  
342 southern (Gulf of Cagliari), northern (Gulf of Asinara) and western (Gulf of Oristano) Sardinian  
343 coasts. All these rivers pass near by the highest populated Sardinian cities, Cagliari, Oristano, and  
344 Sassari, and they outflow in distinct gulfs, which, for their hydrological dynamics, tend to accumulate  
345 litter debris instead of washing them away by currents (Katsanevakis and Katsarou, 2004). A fifth big  
346 Sardinian river, the Flumendosa river, discharge its water on the island's south-eastern coast.  
347 However, the San Lorenzo canyon's presence in front of the Flumendosa rivers mouth might favor the  
348 convey pathway of litter objects from the continental shelf down to the submarine canyons, as already  
349 documented by Tubau et al., (2015) in the north western Mediterranean.

350

## 351 **Conclusions**

352 Despite the almost exponential increase in the number of studies dealing with seafloor macro-litter  
353 over the past decade, our knowledge on the topic remains still limited when coping with the necessity  
354 to develop sound removal actions for the fraction of macro-litter already accumulated in the marine  
355 environment. Data collection is nowadays capable of providing relatively limited information when  
356 considering the deep-sea as the major sink for macro-litter. Indeed, it can be assumed that some major  
357 accumulation areas, have yet to be identified.

358 Numerous surveys have shown that litter can accumulate more likely in some locations rather than in  
359 others, often showing concentrations comparable to those observed in landfills (Pierdomenico et al.,  
360 2019). This warrants reflections on the urgency to schedule, whenever possible, *in situ* removal  
361 activities, which should be carefully assessed.

362 In this perspective, our results are the first to document a scattered distribution of different  
363 macro-litter categories, which likely depend on their sources and mobility through the marine  
364 environment. Such analysis pointed how the identification of macro-litter hotspots performed using  
365 aggregated density data might actually blur the identification of smaller yet considerable macro-litter  
366 hotspots, due to the unequivocal dominance of plastic related categories, both numerically and in  
367 terms of abundance. Moreover, our analyses also highlighted how in GSA11, hotspots for numerically  
368 minor but yet detrimental for the environment macro-litter categories dwell closer to the coastlines  
369 compared to other litter categories. Considering this, and also considering how removal activities are  
370 still in their infancy, prioritizing such closer and more accessible areas could represent a solid  
371 benchmark for developing more challenging remediation activities, to be performed in the deep-sea.

372

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378 ecosistemi marini e dei regimi di compensazione nell’ ambito di attività di pesca sostenibili - raccolta,  
379 da parte di pescatori, di rifiuti dal mare (attrezzi da pesca perduti e dei rifiuti marini) art. 40, par. 1, lett.  
380 a) del reg. (UE)”

381

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**Table 1.** Summary of the MEDITS survey conducted in Sardinian sea, reporting the number of hauls and the swept area per year, geographical sub-area within GSA11, and bathymetric *stratum*. Also, the average density and abundance of items per km<sup>2</sup> ( $\pm$  st. error) are reported per category.

| Years        | Hauls      | Swept Area (km <sup>2</sup> ) | n. items km <sup>-2</sup> |           | Kg km <sup>-2</sup> |           |
|--------------|------------|-------------------------------|---------------------------|-----------|---------------------|-----------|
|              |            |                               | Average                   | St. Error | Average             | St. Error |
| 2013         | 101        | 6.2                           | 75.5                      | 20.8      | 10.0                | 2.6       |
| 2014         | 102        | 5.7                           | 151.1                     | 44.8      | 44.6                | 23.7      |
| 2015         | 101        | 6.4                           | 141.3                     | 29.7      | 41.0                | 14.0      |
| 2016         | 100        | 6.2                           | 70.8                      | 26.4      | 13.5                | 4.3       |
| 2017         | 100        | 5.8                           | 106.4                     | 38.3      | 37.4                | 14.6      |
| 2018         | 102        | 6.4                           | 84.1                      | 22.7      | 13.6                | 5.6       |
| 2019         | 101        | 7.2                           | 121.3                     | 26.2      | 85.4                | 66.9      |
| <b>TOTAL</b> | <b>707</b> | <b>43.9</b>                   |                           |           |                     |           |

| Zones | Hauls | n. items km <sup>-2</sup> |           | Kg km <sup>-2</sup> |           |
|-------|-------|---------------------------|-----------|---------------------|-----------|
|       |       | Average                   | St. Error | Average             | St. Error |
| 1     | 63    | 31.8                      | 5.5       | 10.4                | 4.5       |
| 2     | 126   | 44.7                      | 6.2       | 8.2                 | 2.5       |
| 3     | 70    | 90.9                      | 19.7      | 26.0                | 7.9       |
| 4     | 68    | 81.1                      | 20.1      | 140.9               | 104.3     |
| 5     | 80    | 87.7                      | 16.4      | 39.3                | 16.3      |
| 6     | 184   | 180.1                     | 39.5      | 26.7                | 6.7       |
| 7     | 116   | 139.8                     | 23.7      | 31.3                | 12.1      |

| Strata | Hauls | n. items km <sup>-2</sup> |           | Kg km <sup>-2</sup> |           |
|--------|-------|---------------------------|-----------|---------------------|-----------|
|        |       | Average                   | St. Error | Average             | St. Error |
| A      | 142   | 101.5                     | 19.1      | 32.3                | 11.7      |
| B      | 131   | 221.5                     | 55.2      | 48.9                | 18.4      |
| C      | 168   | 82.3                      | 10.5      | 64.9                | 40.8      |
| D      | 148   | 73.5                      | 8.6       | 10.4                | 2.8       |
| E      | 118   | 65.6                      | 12.4      | 11.4                | 5.4       |

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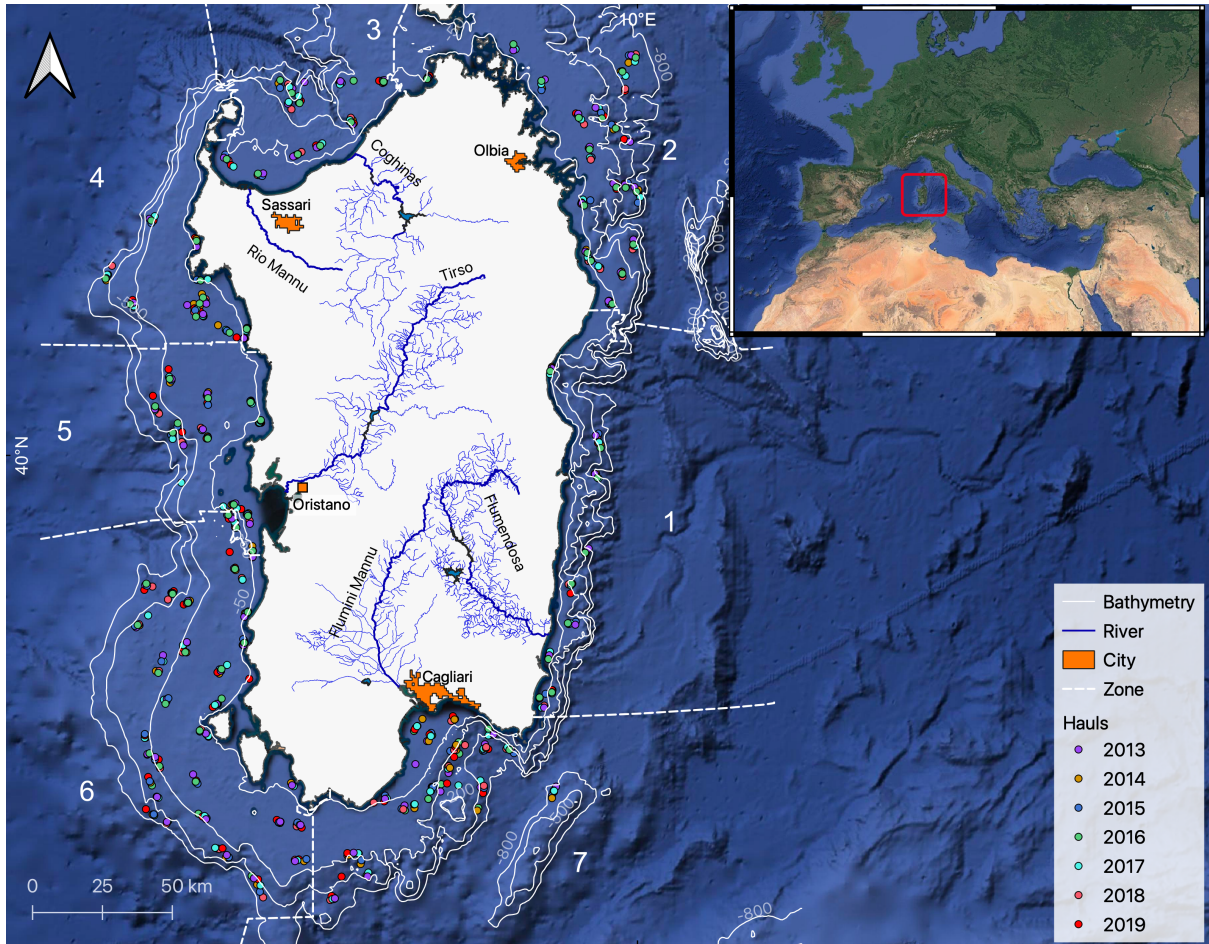
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**Table 2.** Output from the PERMANOVA analysis (main test). Significant Monte Carlo procedure p-values [P(MC)] are reported in bold.

| <b>UNIVARIATE density (n. items km<sup>-1</sup>)</b>   |           |           |                 |              |
|--|-----------|-----------|-----------------|--------------|
| <b>Source</b>  | <b>df</b> | <b>MS</b> | <b>Pseudo-F</b> | <b>P(MC)</b> |
| Year   | 6         | 1153      | 0.691           | 0.762        |
| Zone   | 6         | 4094      | 2.4541          | <b>0.003</b> |
| Stratum  | 4         | 4136      | 2.4791          | <b>0.009</b> |
| Year*Zone  | 36        | 1485      | 0.8899          | 0.730        |
| Year*Stratum   | 24        | 1275      | 0.7639          | 0.885        |
| Zone*Stratum   | 24        | 2550      | 1.5282          | <b>0.013</b> |
| Year*Zone*Stratum                                      | 116       | 1267      | 0.7596          | 0.996        |
| Res  | 247       | 1668      |                 |              |
| Total  | 463       |           |                 |              |
| <b>MULTIVARIATE density (n. items km<sup>-1</sup>)</b> |           |           |                 |              |
| <b>Source</b>  | <b>df</b> | <b>MS</b> | <b>Pseudo-F</b> | <b>P(MC)</b> |
| Year   | 6         | 3462.8    | 1.328           | 0.098        |
| Zone   | 6         | 5967.1    | 2.288           | <b>0.001</b> |
| Stratum  | 4         | 5883.2    | 2.256           | <b>0.004</b> |
| Year*Zone  | 36        | 2608      | 1               | 0.485        |
| Year*Stratum   | 24        | 2732.5    | 1.048           | 0.335        |
| Zone*Stratum   | 24        | 3925.8    | 1.505           | <b>0.002</b> |
| Year*Zone*Stratum                                      | 116       | 2464      | 0.945           | 0.796        |
| Res  | 247       | 2607.3    |                 |              |
| Total  | 463       |           |                 |              |
| <b>UNIVARIATE abundance (Kg km<sup>-1</sup>)</b>       |           |           |                 |              |
| <b>Source</b>  | <b>df</b> | <b>MS</b> | <b>Pseudo-F</b> | <b>P(MC)</b> |
| Year   | 6         | 2251.8    | 0.8915          | 0.601        |
| Zone   | 6         | 3914.2    | 1.5497          | <b>0.050</b> |
| Stratum  | 4         | 7820.5    | 3.0963          | <b>0.001</b> |
| Year*Zone  | 36        | 2082.7    | 0.8246          | 0.898        |
| Year*Stratum   | 24        | 2293.3    | 0.9079          | 0.694        |
| Zone*Stratum   | 24        | 4085.7    | 1.6176          | <b>0.004</b> |
| Year*Zone*Stratum                                      | 116       | 2348.7    | 0.9299          | 0.791        |
| Res  | 247       | 2525.7    |                 |              |
| Total  | 463       |           |                 |              |
| <b>MULTIVARIATE abundance (Kg km<sup>-1</sup>)</b>     |           |           |                 |              |
| <b>Source</b>  | <b>df</b> | <b>MS</b> | <b>Pseudo-F</b> | <b>P(MC)</b> |
| Year   | 6         | 4124.7    | 1.225           | 0.138        |
| Zone   | 6         | 5848.7    | 1.738           | <b>0.007</b> |
| Stratum  | 4         | 8317.9    | 2.471           | <b>0.001</b> |
| Year*Zone  | 36        | 3563.3    | 1.058           | 0.256        |
| Year*Stratum   | 24        | 3533.4    | 1.049           | 0.322        |
| Zone*Stratum   | 24        | 4937.5    | 1.467           | <b>0.001</b> |
| Year*Zone*Stratum                                      | 116       | 3520.8    | 1.046           | 0.189        |
| Res  | 247       | 3365.8    |                 |              |
| Total  | 463       |           |                 |              |

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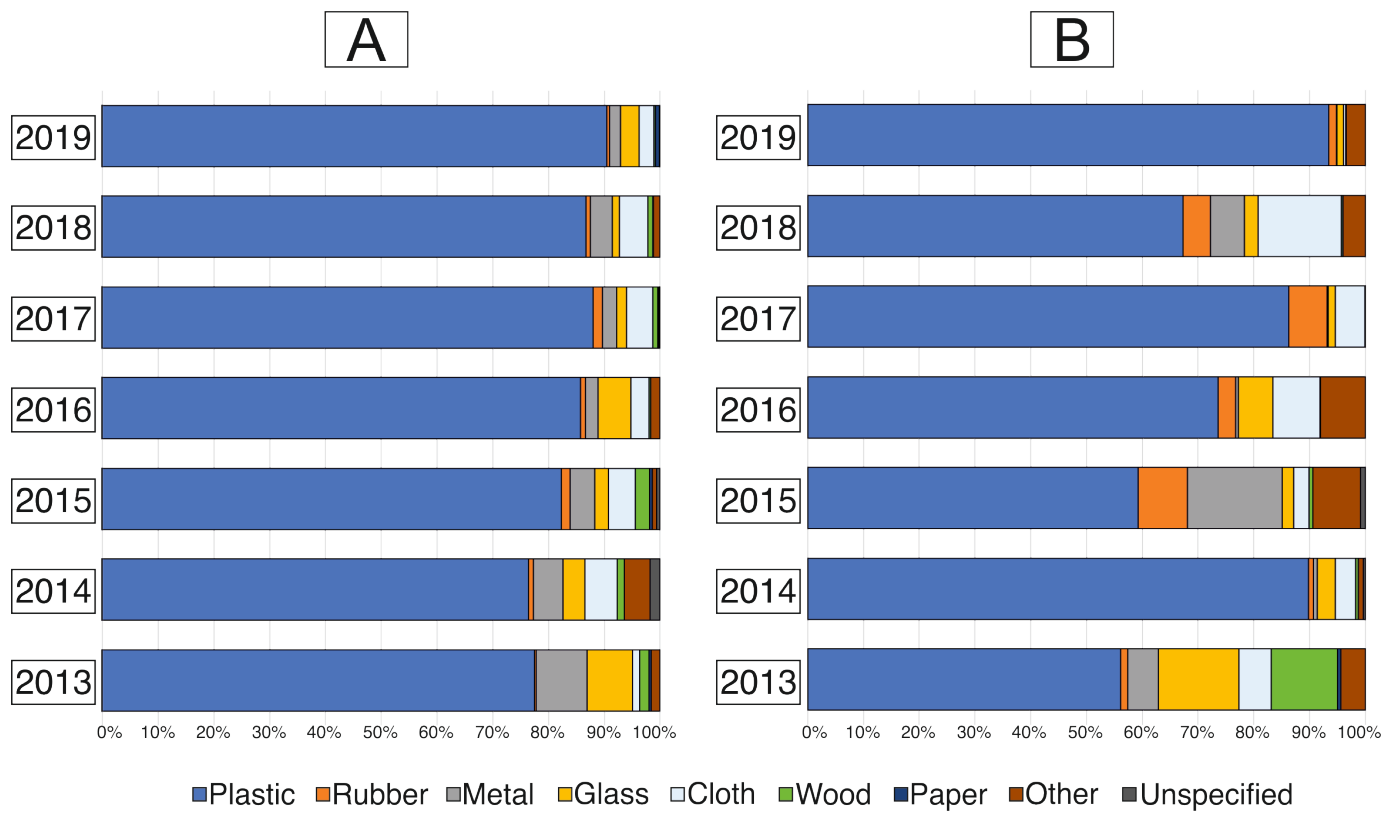
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**Figure 1.** Map of the hauls performed during the study period (2013-2019). The map shows the main rivers in the study area, as well as the major cities in the region.



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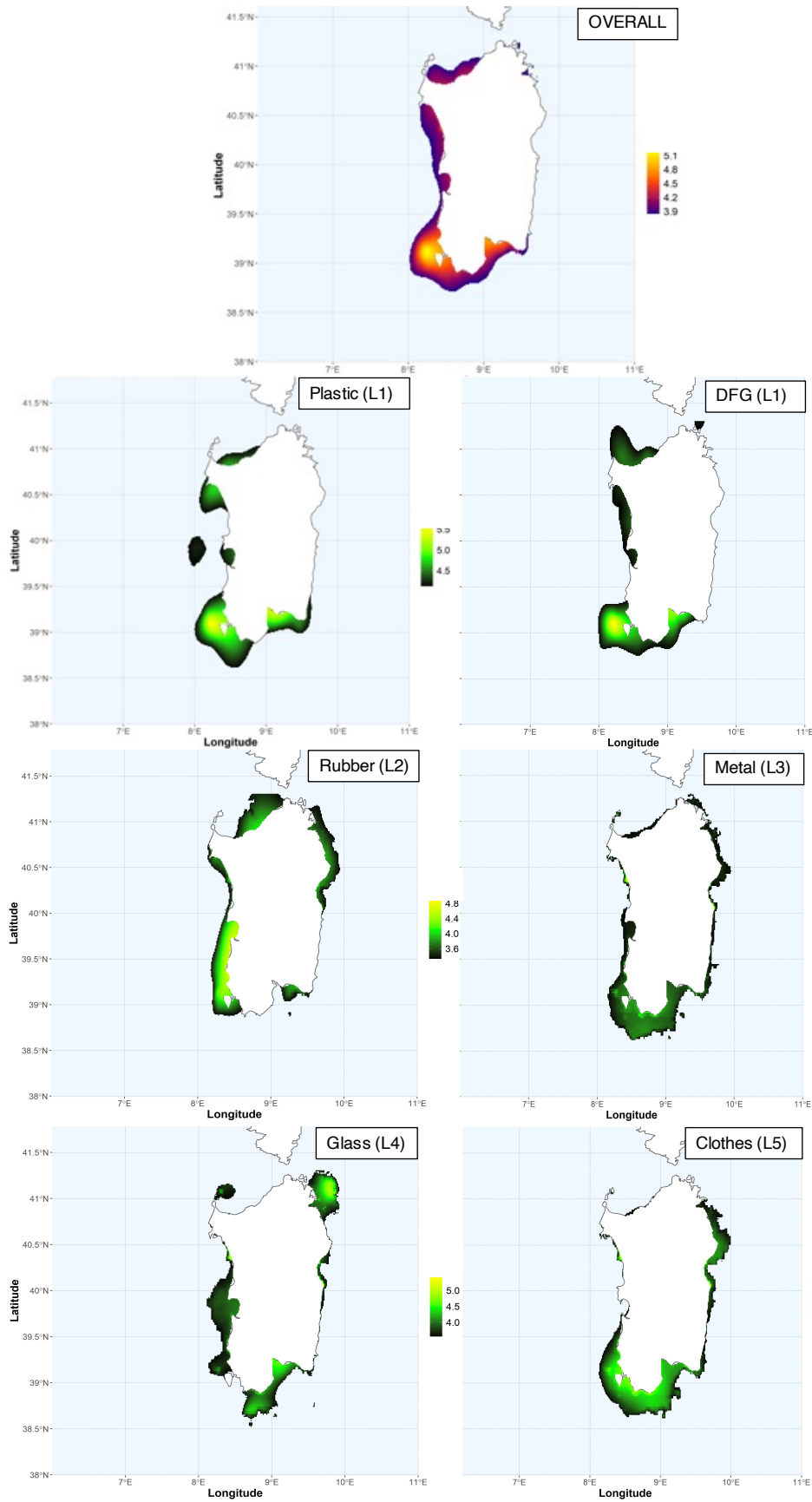
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**Figure 2.** Percentage composition in terms of (A) density (n. items km<sup>-2</sup>) and (B) abundance (Kg km<sup>-2</sup>) of seafloor macro-litter during the study period, comprised between 2013 and 2019.



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**Figure 3.** Multipanel showing different macro-litter categories collected through the MEDITS trawl survey in the study period



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**Figure 4.** Map of sea floor macro-litter hotspot identified in GSA11, represented as results of the five litter categories merged (Overall) and divided according to each of the most representative sea floor macro-litter categories. The scale at the right of the figures is reported as logarithmic weighted n. of items.