1	Integrated environmental evaluation of heavy metals bioaccumulation in invertebrates and
2	seaweeds from different marine coastal areas of Sardinia, Mediterranean Sea.
3 4	Francesco Corrias <sup>a</sup> , Alessandro Atzei <sup>a</sup> , Piero Addis <sup>b</sup> , Marco Secci <sup>b</sup> , Mariateresa Russo <sup>c</sup> , Alberto
5	Angioni <sup>a</sup> *
6	
7	
8	<sup>a</sup> Department of Life and Environmental Science, Food Toxicology Unit, University of Cagliari,
9	University Campus of Monserrato, SS 554, 09042 Cagliari, Italy,
10	Francesco Corrias (francesco.corrias@unica.it), Alessandro Atzei (alessandro.atzei@unica.it),
11	<sup>b</sup> Department of Life and Environmental Science, Biology Section, University of Cagliari, via
12	Tommaso Fiorelli 2, 09126 Cagliari, Italy, Marco Secci (msecci@unica.it), Piero Addis
13	(addisp@unica.it).
14	<sup>c</sup> Department of Agricultural Science, Mediterranean University of Reggio Calabria, Località Feo di
15	Vito - 89122 Reggio Calabria (RC), Mariateresa Russo (mariateresa.russo@unirc.it)
16	
17	Running title: Heavy metals accumulation studies in marine coastal areas of Sardinia,
18	Mediterranean Sea.
19	
20	* Corresponding author:
21	Alberto Angioni
22	Food Toxicology Unit, University of Cagliari, University campus of Monserrato, SS 554, 09042
23	Cagliari, Italy; Phone: +390706758615, Fax: +390706758612; e-mail: aangioni@unica.it
24	
25	
26	

#### 27 Abstract

In this work, three gastropods *Patella vulgata*, *Osilinus turbinata*, and *Tahis clavigera*, one echinoderm *Parancetrotus lividus*, one coelenterate *Anemonia sulcata*, and two seaweed *Padina pavonica*, and *Cystoseira mediterranea* were collected from three different marine areas of Sardinia in the Mediterranean sea and studied for heavy metals and metalloid content and accumulation trends.

33 Inductively coupled plasma optical emission spectrometry (ICP-OES) was used for the 34 determination of Al, AS, B, Ba, Be, Cd, Co, Cr, Cu, Fe, Hg, Mn, Mo, Ni, Pb, Sb, Se, Sn, Sr, Te, Ti, 35 V and Zn in the selected samples. The results showed that gastropods were capable of accumulating 36 Al, Ba, Cu, Fe, Sr, and Zn; seaweeds can better concentrate Al, Fe, and Zn than all other species. At 37 the same time, echinoderms and coelenterate had limited ability to store specific metals, showing a much more homogeneous distribution. PCA analysis allowed us to discriminate among the sites and 38 39 the species. Cala Zafferano was the area with the higher values of accumulation of all metals in all 40 species as expected, considering its proximity to industrial sites. The results of the analysis showed 41 clearly that heavy metal and metalloid accumulation was different for each species studied. 42 Therefore, for a correct environmental assessment of a given area, a comprehensive approach is 43 strongly recommended by exploiting the different properties of both accumulation and 44 concentration of the metals by different aquatic species.

45

46 Summary

47 Different species of aquatic living organisms have different heavy metals and metalloid
48 accumulation trends, even if they belong to the same family and the same water column or have
49 similar living habits.

50

51 Keywords: Heavy metals; accumulation; gastropods; Paracentrotus, Anemonia, seaweed, PCA

#### 53 Introduction

Metals are usually found in the environment worldwide, their presence caused by natural events or because of anthropogenic activities, such as industrial sites and agricultural practice (Kamaruzzaman et al., 2011). The increase in environmental metal concentration in these areas can reach 100-1000 times the normal levels of Earth's crust (Carral et al., 1995). The hydrological cycle of water between the land, open water surface and sea can bring to significant pollution, exposing the aquatic ecosystems directly to a great variety of these metals causing severe contamination of marine organisms (Farombi et al., 2007).

61 Metals are mutually essential elements necessary to support biological activities and unneeded 62 chemicals with an unidentified role (Roesijadi and Robinson, 1994). Heavy metals are problematic 63 pollutants because, unlike most organic contaminants, they are non-biodegradable and can accumulate in living tissues and food, causing a significant threat to both human health and 64 environmental safety. The most hazardous heavy metals and metalloids are Be, Cr, Hg, Cd, Pb, and 65 As. These metal ions are low soluble in water, can cause toxicity, and have serious side effects on 66 67 human health. After discharged in the water, they are usually accumulated by sediments and 68 transferred to invertebrates and biomagnified in fish populations accumulating in the lipid tissue 69 (Fergusson, 1990; Yi et al., 2011; Jakimska, 2011; Jaishankar et al., 2014; Jitar et al. 2015). An 70 accurate assessment of the degree of metal contamination at a specific location requires an adequate 71 estimation of these elements' natural levels in the biological communities and the natural 72 environment. Mussels, as interdital filter feeders, accumulate pollutants coming from local 73 contamination and distant areas influenced by sea current and thus can lead to inaccurate 74 information on the real environmental situation of restricted territories. Therefore, environmental pollution monitoring studies relying on their ability to concentrate high levels of metal in the soft 75 76 tissue used mussels, invertebrates, and marine vegetables widely (Philips, 1977; Ingston, 1982; 77 Goldberg et al., 1983; Cossa, 1989; Bryan and Langston, 1992; Abisil et al., 1997; Hummel et al.,

Commentato [IC1]: Ho messo la bibliografia dopo widely.

78 <mark>1997)</mark>.

The use of sedentary herbivorous such as gastropods, and echinoderms, seaweeds holdfast, and anemonia has several advantages. These living organisms are available all year, simple to collect, resistant to transportation and conservation, and have enough tissues for the analysis.

82 In the last decades, metals bioaccumulation has been investigated in different biological communities such as invertebrate gastropods (Vinogrjidov et al., 1953; Blackmore, 2000; Perez-83 84 Lopez et al., 2003; Amin et al., 2006; Collado et al., 2006; Yüzereroğlu et al., 2009; Ramirez, 2013; 85 Kelepertzis, 2013; Boucetta et al., 2016; Duysak and Azdural, 2017; Aydin-Onen and Ozturk, 2017; 86 Gawad, 2018), anemone (Mitchelmore et al., 2003; Horwitz et al., 2014), or echinoderms (Salvo et 87 al., 2014, Chiarelli et al., 2019). Moreover, some authors investigated seaweed such as Padina spp 88 and Cystoseira spp (Aydin-Onen and Ozturk, 2017; Rodríguez-Figueroa et al., 2008; Benfares et 89 al., 2015; Ryabushko et al., 2017). However, most articles reported pollution from 4 to 15 metals in

90 one or two species.

91 The study aimed to compare metals accumulation performance of 7 living organisms (invertebrates 92 and vegetables) in three marine areas of Sardinia characterized by the different anthropic impact. 93 Cala Zafferano (CZ) affected by heavy metal contamination, Capo Frasca (CF) and Capo 94 Spartivento (CS), representing potentially clean environments, exposed to NW, SW and SE 95 currents. Three gastropods Patella vulgata, Osilinus turbinata, and Tahis clavigera, one 96 echinoderm Parancetrotus lividus, one coelenterate Anemonia sulcata, and two seaweed Padina 97 pavonica, and Cystoseira mediterranea, were collected. All samples were subjected to 98 mineralization and analysis by ICP-OES to determine an exhaustive pattern of 23 metals and 99 metalloids and accumulation studies.

100 1. Material and methods

101 2.1 Sample collection and processing

Samples were handpicked from three sites of the Sardinia coast, Capo Spartivento (38° 52'41.90 "N;
8° 50'28.77 "E), Cala Zafferano Teulada (38° 53'43.26 "N; 8° 39'42.19 "E), and Capo Frasca (39°
42'54.98 "N; 8° 26'44.14 "E). Twenty samples of the gastropods *Patella vulgata, Osilinus turbinata*,

and *Thais haemastoma*, the echinoderm *Parancetritus lividus*, the coelenterate *Anemonia sulcate*, and the seaweeds *Padina pavonica*, and *Cystoseira mediterranea*, were collected at a maximum depth of 10 meters in August (Table 1). All species were present at all stations and individually processed.

After collection, samples were immediately stored in ice cooler "drain plug open" to let the water run out. This operation avoids deterioration during transport to the lab. Samples brought to the laboratory were weighed, and gastropods soft tissue was separated from the shell. Sea urchin's gonads were collected with a stainless-steel spatula from the shell opened with a stainless-steel cutter for sea urchins. Anemone tentacles and seaweed aerial parts were selected using a stainlesssteel scissor. All samples, except gonads, were rinsed with deionized water to remove accidental impurities, and finally stored at -20°C before analysis.

116 Chemicals

HNO<sub>3</sub> 67-69%, H<sub>2</sub>O<sub>2</sub> solution 30%, and standards of Al, AS, B, Ba, Be, Cd, Co, Cr, Cu, Fe, Hg, Mn, Mo, Ni, Pb, Sb, Se, Sn, Sr, Te, Ti, V, Zn were of ICP grade (Carlo Erba Reagents Milan, Italy), HCl 34-37% super pure quality (Romil Spa Cambridge, England). Double-deionized water with a conductivity less than 18.2 MΩ was obtained with a Milli-Q system (Millipore, Bedford, MA, USA). Stock standard solutions were prepared at 1000 mg/L in HNO<sub>3</sub> 10%. Working standard solutions were prepared daily by diluting the stock solution with MilliQ water.

123 2.2 Moisture and ash

Ten grams of homogenized samples were weighed and dried at 105°C in a thermostatic heater until a constant weight was achieved (~ 24 h). After that, samples were stored in a desiccator to reach ambient temperature and weighted for dried mater assessment. Dehydrated samples were then carbonized at 450 C° in a porcelain crucible for five hours for total ash analysis. Moisture and ash

128 determinations were calculated as a percentage of the fresh product.

129

Commentato [IC2]: Weighed??

About 0.2 g of ashes obtained by carbonization were mineralized in a microwave CEM Mars6 system (CEM Corporation, Milano, Italy), by adding 10 mL of a mixture of HNO3 and HCl (1:3), and 1 mL of H2O2. After the mineralization program, the solution was transferred into a 10 mL flask and brought to volume with MilliQ water with a conductivity less than 18.2 MQ, filtered through 0.45  $\mu$ m nitrocellulose membrane filter (Whatman, Milan, Italy) in another flask and finally subjected to analysis. Control solvent samples were simultaneously prepared to avoid false positives and contamination during the investigation.

138 Hg, As, and Se sample preparation were carried out using Agilent VGA-77 (Agilent, Milan, Italy) 139 according to Agilent application notes (Agilent AN, 2014; Agilent AN, 2012). Analysis were 140 carried out using a Varian 710ES ICP Optical Emission Spectrometer for the simultaneous analysis 141 of 23 metals and metalloids. Operating conditions were as follows: radiofrequency (RF) generator 142 power 1.2, frequency 40 MHz; argon (99.996% purity) was used both for plasma (15.5 L/min), 143 nebulizer (200 Kpa) and optic supply (1.5 L/min). The sample uptake delay was 30 seconds, and the 144 instrument stabilization delay 15 seconds. The spray chamber was a double-pass, glass cyclonic. 145 Each measurement was made in triplicate. The power and pressure applied were 600 W and 100 146 PSI for 13 minutes. The limit of quantitation (LOQ) was calculated as ten times the standard 147 deviation reading of the blank sample signal (Wenzl et al. 2016). Calibration curves were calculated 148 on five points starting from the LOQ value and were considered acceptable for  $R^2 \ge 0.999$ .

149

### 150 2.4 Risk assessment

In this study, we used the target hazard quotient (THQ) for risk assessment. THQ represents the ratio of the estimated daily intake (EDI) / reference dose (RfD) for noncarcinogenic risk assessment methods indicating the risk level associated with xenobiotic exposure (Chien et al., 2002; Zhang et al., 2017). RfDs for oral intake were obtained from the Integrated Risk Information System (IRIS-EPA). When THQ is below 1 no adverse effect during a person's lifetime is expected. EDI was calculated as the concentration of heavy metals (mg/kg wet weight) × average daily consumption of



**Commentato [IC5]:** Lo ripetiamo di nuovo dopo i materiali?

Commentato [IC6]: Ho aggiunto questo.

158 and metalloids from invertebrate consumption, we consider an average weight of 60 Kg. Total THQ 159 for each species was calculated as the sum of single metals THQ, while total THQ of each site was 160 calculated assuming the intake from all the selected species in the same day. 161 2.5 Statistical Analysis 162 Analysis of variance (ANOVA) was carried out with the software XLSTAT (Addinsolf LTD, 163 Version 19.4), Mean comparisons of the effects of treatments were calculated by the Fisher's least 164 significant difference test at  $p \le 0.05$ . 165 Analysis data sets were imported into SIMCA 14 (Umetrics AB, Umea, Sweden) for processing 166 using principal component analysis (PCA). Additionally, R2 (the coefficient of determination) and 167 Q2 (the cross-validated correlation coefficient) were used as measures for the robustness of a

invertebrates in the local area (g/day bw) / body weight (Bw). To evaluate the risk of heavy metals

168 169

157

#### 170 3. Results

pattern recognition model (Atherton et al., 2006).

171 The selected sites had different anthropic conditions. Cala Zafferano (CZ) is in a military area, 172 downstream of an industrialized coastal area, and subjected to main sea NW currents. Capo 173 Spartivento (CS), located in the Gulf of Cagliari, is protected from the main current NW and 174 exposed to SW and SE, in contrast, Capo Frasca (CF) is located in the south of the Gulf of Oristano 175 but totally exposed to NW current. Both sites are far from industrial and agriculture areas (Figure 176 1). All organisms tested were present in each site and were selected with similar size and weight 177 among the species (Table 1). The analysis of water content and ashes showed similar values in the 178 three sites for gastropods, ranging in average from  $74.44 \pm 2.43\%$  (Osilinus, CZ), to  $79.30 \pm 1.05\%$ 179 (Patella, CZ). Ash content ranged on average from  $3.75 \pm 2.41\%$  (Thais, CF), to  $8.52 \pm 25.75\%$ 180 (Osilinus, CZ). However, the higher water levels were detected for Anemonia 181 sulcata, Paracentrotus lividus and Padina pavonica accounting in average for  $90.85 \pm 1.72\%$ , 81.72 Commentato [IC7]: Perché maiuscolo?

Commentato [IC8]: Non deve and are in corsivo?

- 182  $\pm$  2.77%, and 82.45  $\pm$  1.82%, respectively, while the lower levels were for *Cystoseira* 183 *mediterranea* with an average value of 72.52  $\pm$  7.12% (Table 2).
- Seaweeds showed the higher mineral content with ash average values accounting for  $15.35 \pm 17.98\%$  (g/100g  $\pm$  RSD, FW) and  $12.86 \pm 19.98\%$  (g/100g  $\pm$  RSD, FW) in *Padina* CZ and *Cystoseira* CS, respectively, almost three times higher than the analyzed invertebrates. Instead, the lowest ash levels were detected in *Anemonia sulcate* CZ and *Paracentrotus lividus* CZ with average values of  $3.50 \pm 7.89\%$  and  $2.10 \pm 12.60\%$  (g/100g  $\pm$  RSD, FW), respectively.
- Gastropods soft tissue, echinoderms gonads, and coelenterate anemone tentacles were used for theanalysis, while the juvenile areal parts were selected for seaweeds.
- 191 The ICP-OES analytical method allowed the analysis of 23 heavy metals and metalloids. 192 Calibration curves showed  $R^2$  values ranging from 0.9967 (Pb) to 0.9997 (Co), and the limits of 193 quantification of the method were suitable to evaluate overall pollution and risk assessment (Table 194 3) (Wenzl et al., 2016).
- Among the heavy metals and metalloids searched within the methods, eight were missing (Be, Co, Hg, Mo, Sb, Se, Sn, Te), and only 6 (Al, Cu, Fe, Sr, V, Zn) were found in all samples from the three sites. Gastropods and seaweeds showed in all sites the presence of 12 heavy metals and metalloids, *Paracentrotus* 10, and *Anemonia* 9 (Table 4-6).
- 199 3.1 Gastropods
- 200 Patella vulgata showed higher values for seven heavy metals and metalloids (Al, Ba, Fe, Mn, Pb, 201 Sr, Ti) in the samples from CZ, while CS and CF showed even values. B, Cd, Cr, Cu, V, and Zn had 202 similar values in all three sites. Fe, Sr, and Al from CZ had the higher mean values accounting for 203  $34.28 \ \mu g/g$ ,  $7.06 \ \mu g/g$ , and  $6.44 \ \mu g/g$ , respectively, about four times higher than those from CF and 204 CS (Table 3). Osilinus turbinatus had the higher mean values in the samples from CZ for Al, Cr, 205 Cu, Mn, Fe, Ni, Pb, Sr, Ti, and Zn, among these, the highest concentrations were detected for Fe 206 40.22 µg/g, Sr 19.08 µg/g, Al 8.77 µg/g, and Ni 7.93 µg/g. B, Ba, Cd, and V had similar values in 207 the three sites. Also, Thais showed higher values in the samples of CZ. In particular, Al, Ba, Cu, Fe,

- 208 Pb, and V were the most concentrated, with Fe 36.89  $\mu$ g/g, Cu 29.68  $\mu$ g/g, Zn 21.91  $\mu$ g/g, and Ba
- $209 \quad 10.04 \ \mu g/g$  with the higher values (Table 4).
- 210 Comparing the three gastropods in the three sites, Al, B, Cu, Fe, and Ti showed similar values in the
- 211 samples from CZ for Patella and Osilinus, Mn and Sr for Patella and Thais, while Cr had similar
- values in all sites. In CZ, *Thais* showed higher mean values of Ba, Cd, Cu, and Zn, and lower mean
  values for Al, B, and V.
- 214 In CS, B, Fe, and Ti had overlapping values in Patella and Osilinus, while Mn and Sr were similar
- 215 in Osilinus and Thais. Al, Ba, Cd, Pb, and Zn had different values in all sites.
- 216 CF showed more considerable variability among all metal's accumulations in the three species.
- 217 Considering all sites for the single species, the samples of CZ showed higher levels of all metals,
- 218 except for Zn, which showed overlapping values in the three sites for the three species.
- 219 3.2 Seaweeds
- Heavy metal and metalloids distribution followed an entirely different and uneven distribution among the sites and between the two seaweed species (Table 5). *Padina* and *Cystoseira* showed both the higher values in the samples of CZ, Fe 225.46  $\mu$ g/g, Al and Zn both at 104.4  $\mu$ g/g, and Sr 77.46  $\mu$ g/g in *Padina*, and Fe 184.03  $\mu$ g/g, Al 41.85  $\mu$ g/g, Sr 55.46  $\mu$ g/g, Zn 41.84  $\mu$ g/g, Ba 32.08  $\mu$ g/g, B 12.60  $\mu$ g/g, and Mn 6.58  $\mu$ g/g in *Cystoseira*.
- 225 Al, Fe, and Zn in *Padina* showed higher concentration in CZ compared to CS and CF, and
- 226 to *Cystoseira* samples in all sites. *Cystoseira* levels of Al, B, Ba, Cu, Fe, Mn, Sr and Zn were higher
- 227 in CZ samples, while Cr, Pb, Ti and V were higher in CF.
- 228 3.3 Echinoderms and coelenterate
- *Parancetrotus lividus* accumulated only a few metals at moderate concentrations, Al, Sr, and Zn were highly concentrate in CZ. At the same time, the other trace elements showed values in an even range in the three sites. The higher levels were detected for Zn 9.24 mg/g, Al and Sr 5.64  $\mu$ g/g, and Fe 4.25  $\mu$ g/g in CZ. The levels of Fe were similar in CZ and CF, and lower in CS (Table 6).

233 Anemonia sulcata showed the presence in the tentacles of some minerals at low levels, among the 234 heavy metals detected, Fe 4.07  $\mu$ g/g, and Al 2.21 mg/g were the most abundant in CZ. Also, Sr and 235 Zn showed values slightly higher in CZ than in CS and CF (Table 5).

236 Accumulation results from all samples showed an overall content of metals higher in Cala 237 Zafferano, especially for Fe, Al, and Zn in seaweeds. These findings confirmed an increase in the 238 environmental pollution in this area which suffered the anthropic impact of the industrial area due 239

to the exposition to NW current (Figure 1).

240 Among the heavy metals and metalloids considered most dangerous, Hg was never detected at a 241 value above the LOQ of the method (0.05  $\mu$ g/g), which was ten times below the MRL set for fish 242 (CR-EC No 1881/2006). Arsenic was below the LOQ in all samples except in seaweed, and 243 Cystoseira showed higher values in all sites. Pb was not present in Paracentrotus and Anemonia, 244 and the samples of Osilinus from CS and CF. In the other samples, its values were all below the 245 MRL (0.3  $\mu$ g/g), except in the samples of CZ accounting for 0.50  $\mu$ g/g in *Patella* and 0.46  $\mu$ g/g 246 in Thais. Cd was not detected in Anemonia, while showed values below the MRL (0.05 mg/g) 247 in Paracentrotus and Padina from CS and CF. All other samples showed values above the MRL, 248 with Patella showing a value four times higher of the MRL, and Thais 20 times higher. No 249 differences were detected among each species between CS and CF sites.

250 3.4 Multivariate analysis

251 Multivariate analysis statistics could help in better discriminate among the different species, due to 252 the high number of samples and variables from the analysis. In this paper, we have used the 253 principal component analysis (PCA) approach do describe if the various species studied had 254 different heavy metals accumulation capacity and which were the most discerning heavy metals. 255 PCA biplot of score and loadings for CZ and CS showed a separate group of datasets for each 256 species, allowing to identify 12 clustering groups and to discriminate metals along PC1 and PC2 257 axis (Figure 3).

Commentato [IC9]: Tabella 6

Commentato [IC10]: Ho aggiunto metalloids visto che poi parli di mercurio, ecc.

258 PCA emphasize the differences in accumulation of the seaweed Cystoseira and Padina in CZ 259 samples to the other species studied in the same area and to all the other samples collected in CS. 260 Among the gastropod, Thais was the most discriminating species. CZ and CS samples were 261 separated from the other gastropod's samples, invertebrate and seaweed, moreover, the samples 262 from CZ were split from CS along the PCA1 axis stressing a more influence of the residues of the 263 most discriminating metals for this axis. The samples of Patella, Osilinus and Paracentrotus fall in 264 a restricted zone of the PCA plot, despite this they were well separated between CZ and CS along 265 the PC1 axis. The metals Pb, Ba, Fe, Mn, Al, Cr, and B were the most discriminating along the PC1 266 axis, while Cu, Zn, Mo, and Cd were the most discriminating in PC2 axis (Figure 3).

Risk assessment showed total THQ for each species below 1, indicating no adverse effects from the consumption of the selected organisms. In all sites, *Thais* showed higher values, while *Anemonia* showed lower values in all sites (Table 7). Moreover, total THQ for the site obtained from the sum of total THQ of the various species were below 1, confirming that no adverse effects were reached also consuming the different species in the same day.

272 The values of total THQ for the site confirmed that CZ had higher harmful metals concentration

than CS and CF, and the amounts consumed should be taken into consideration (Table 7).

## 274 4. Discussion

275 Biomonitoring studies have the scope to assess the environmental occurrence of pollutants 276 potentially toxic for human and other living organisms. Different approaches can be used, such as 277 bioconcentration, that studies the direct partition form the environment to target organisms, and 278 biomagnification that is strictly related to the diet leading to increase xenobiotic concentration in the 279 feeder than in the diet. In contrast, bioaccumulation describes the accumulation and enrichment of 280 contaminants in a selected living organism. This process is species-dependent since different 281 organisms have different feeding habits, mechanisms of detoxification and metabolism (Jasminska 282 et al. 2011). Metals and metalloids are commonly present in the environment and can be essential

**Commentato [IC11]:** Ripetizione. Lo hai già messo a inizio frase.

- 283 elements supporting biological activities or have adverse effects decreasing aquatic species or being
- 284 dangerous for human being trough the food chain.
- 285 Gastropods, echinoderm, and anemonia species are usually eaten in Sardinia. THQ showed that the
- intake of single metals from the selected areas trough the diet did not create an alarm for the
- 287 population if food quantities remains in the local typical FDI (food daily intake).
- 288 Metals do not degrade and along the trophic pyramid tend to increase, and anthropic activity or
- 289 geochemical conditions can influence their concentration, the natural levels can grow significantly,
- 290 becoming harmful to human safety. Literature data reported xenobiotic accumulation on suitable
- 291 species used as indicators. Although most studies have addressed local issues, their impact is
- 292 important to assess the different bioaccumulation trends of the organisms used.
- 293 Perez-Lopez et al. (2003) studied the contamination of Patella vulgata from Zn, Cd, Pb and Cu with
- levels ranging in the soft tissue from 91 to 429 µg/g for Zn, 5.6 µg/g for Cd, 1-3 µg/g for Pb, and 210 µg/g for Cu.
- 296 Kelepertzis (2013), studying the heavy metal concentration in *Patella*, postulated that many 297 mineralized areas give rise to higher levels of Pb (96 vs 8  $\mu$ g/g), Zn (196 vs 48  $\mu$ g/g), Mn (36 vs 10
- 298  $\mu g/g$ ), and Cu (28 vs 6  $\mu g/g$ ) bioconcentration. In contrast, Ni (9  $\mu g/g$ ), and Cr (8  $\mu g/g$ ) were even,
- 299 the levels of metals decreased in the line Zn > Pb > Mn > Cu.
- 300 Duysak and Azdural (2017) studied the accumulation from 8 sites in the different tissues
  301 of *Patella* of Fe, Zn, Cd, Cu, Co, Ni, Mn, Pb, Cr and Al, with the higher concentration in all sites
- 302 for Fe, Zn, Cd, Co, Pb, and Al.
- Collado et al. (2006) found that mean total concentrations of Cd, Cu, Pb, and Zn in *P. rustica* were 0.37  $\pm$  0.05, 1.77  $\pm$  0.09, 1.27  $\pm$  0.07 and 8.84  $\pm$  0.71 µg/g DW (mean  $\pm$  S.E.) respectively; whereas in *P. candei crenata* were 0.71  $\pm$  0.10, 2.94  $\pm$  0.11, 0.09  $\pm$  0.01 and 33.74  $\pm$  1.15 µg/g DW (mean  $\pm$
- 306 S.E.).

Commentato [IC12]: Through?

Gadaw (2018) studied the accumulation in the gastropod *Lanistes carinatus* of Mn, Fe, Co, Ni, Cu, Zn, Cd, and Pb. With values ranging from 240 to 567  $\mu$ g/g dw for Fe, 5.12 – 32.8  $\mu$ g/g dw for Mn, Zn 5.35-28.15  $\mu$ g/g dw, 0.69-3.25  $\mu$ g/g dw for Cu, 4.42-13.87  $\mu$ g/g dw for Ni, 0-9.24  $\mu$ g/g dw for Co, 0.03-0.04  $\mu$ g/g dw for Cd, and 0.24-0.40  $\mu$ g/g dw for Pb.

Jakimska et al. (2011) reported the levels of Zn and Fe in gastropods. In these molluscs these two metals have essential biochemical functions, Zn is a constituent of haemocyanin, Fe is a constituent of goethite ( $\alpha$ -FeOOH) responsible for the proper functioning of the radula. This fact can explain the high levels of Zn and Fe in gastropods. Moreover, the levels of Zn are related to Cu levels through the metabolic processes and homeostasis, metallothionein mediated.

316 Amin et al. (2006) studied the accumulation of Cu, Zn, Fe, Pb, Cd in the soft tissues of *Thais* 317 *aculeata* with values of 150.3  $\pm$  58.0 µg/g DW, 158.4  $\pm$  27.9 µg/g DW, 316.7  $\pm$  125.3 µg/g DW, 318 11.3  $\pm$  5.9 µg/g DW and 0.6  $\pm$  0.3 µg/g DW, respectively.

Ramirez et al. (2013), studied the levels of Cd (0.69-15.81  $\mu$ g/g), Cu (10.82-55.68  $\mu$ g/g), Pb (0.21-1.36  $\mu$ g/g), and Zn (15.78-32.09  $\mu$ g/g) in *Osilinus*. Later Boucetta et al. (2016) reported the determination in *Osilinus* of Cd, Cr, Cu, Ni, Pb, Zn in two stations, recording different values in the two stations, accounting for 318.10±38.20  $\mu$ g/g and 6.51±1.47  $\mu$ g/g for Zn and Pb in the first station. And 3.63±1.14  $\mu$ g/g for Cd, 32.70±2.90  $\mu$ g/g for Cu, and 24.80±20  $\mu$ g/g for Ni in the station 2.

Salvo et al. (2014) studied the accumulation of Hg, As, Cr, Ni, Cu, V, Cd and Pb in Paracentrotus *lividus*, their findings suggested the use of *Paracentrotus* as a bioindicator for pollution in defined
areas. Similar findings were presented by Angioni et al. (2012, 2014) analyzing polycyclic aromatic
hydrocarbon (PAH) pollution in different Sardinian coastal areas.

Horovits et al. (2014) analyzed symbiotic algae and the tentacles of *Anemonia* for the bioaccumulation of 12 trace elements. *Anemonia viridis* regulates its internal trace element concentrations by compartmentalization and excretion owing a high resilience to environmental contamination from high levels of metals. Pb, Fe, and Cu were mostly associated with symbiotic 13 algae, while Cd, As, and Zn were associated with the *Anemonia*. 85% of Fe was in the alga fraction,
Pb and Cu were evenly distributed, while Zn, As, and Cd were mostly distributed in the *Anemonia*.
Mitchelmore et al. (2003) studying the accumulation of heavy metals in anemone pointed out that
symbiotic anemones accumulate Cd, Ni e Zn higher than aposymbiotic anemones, and the levels
were depleted when recovered in seawater.

338 Data on the contamination of Cystoseria have been reported by Ryabushko et al. (2017) on Pb (35-

339 50  $\mu$ g/g), Cd (2-3.5  $\mu$ g/g), Zn (40-65  $\mu$ g/g), and Cu (13-24  $\mu$ g/g), and by Benfares et al. (2015) on

340 Cd, Pb (0-5.1  $\mu$ g/g), Cr (0.1-1.9  $\mu$ g/g), and Hg.

341 The same specimen with different size can give rise to different levels of xenobiotics; therefore, to 342 avoid individual variation, we selected specimens from the same species with similar size in each 343 site and among sites.

Literature most studied metals were Cd, Cr, Cu, Pb, and Zn, by using one or two organisms, mainly comparing polluted vs clean areas. The merged data from Sardinia obtained in the present study showed values for all living organisms in the lower range compare to literature data for Cd, Cr and Cu, while Pb and Zinc levels in gastropods were higher than the levels reported in the Canary Islands (Table 8).

349 Mercury detected in mussels collected along the coastline of Sardinia from two sampling campaigns 350 showed values ranging from 0.035 to 0.11  $\mu$ g/g and 0.048 to 0.83  $\mu$ g/g (Ipolyi et al. 2004). 351 Moreover, data reported on carnivorous and herbivorous or omnivorous fish showed that the 352 seconds accumulate till 10 times less the concentration of the former (Levitan et al., 1974; Liu et al., 353 2014). Therefore, values of Hg below the LOQ in the samples of the present study were not 354 surprising since most of the living organisms studied are mostly herbivorous or seaweed (Table 1). 355 The results from the studies examined showed that local pollution cannot be used for global 356 discussions but take pictures of a scenario in restricted areas. The Mediterranean basin, which is an enclosed sea, has different contamination characteristics related to the anthropic effects on the 357 358 coasts and on the sea of the single surrounding countries.

Commentato [IC13]: Forse bisogna mettere Hg?

Therefore, the real value of these studies is related to the global information of the accumulation trends of the different living organisms.

361 The present study elucidated that the species of gastropods, Paracentrotus, Anemonia and seaweed

362 living in the same water column (upper littoral zone) had different uptake capability on metals.

363 These data confirm that accumulation is species and site-dependent, being related to pollution364 magnitude, feeding habitude and geochemical morphology of the site.

365 MVA analysis emphasizes the different accumulation capacity of the seven species collected at the 366 different upper littoral areas of Sardinia, allowing to distinguish areas with different pollution 367 contaminations in a similar contest.

368

# 369 5. Conclusions

370 The data reported showed different capabilities of the species involved in accumulating different 371 metals. Gastropods are more capable of accumulating Al, Ba, Cu, Fe, Sr, and Zn, but in different 372 amount. Thais levels of Cu and Zn are much higher than Patella and Osilinus, seaweeds can 373 concentrate Al, Fe, and Zn (Padina more than Cystoseira) in large extent in respect to all other 374 species. At the same time, Paracentrotus and Anemonia have limited ability to concentrate on 375 specific metals but have a much more homogeneous distribution. THQ showed that Thais 376 contributed for almost 50% to total THQ of the site and that among the heavy metals Cd was the 377 most impacting on the overall health concerns for invertebrate eating.

All these species have been referred to be model organisms for biomonitoring environmentalquality and pollution.

380 Therefore, for a correct environmental assessment of a given area, a diversified approach is strongly 381 recommended by exploiting the different properties of both accumulation and concentration of the 382 metals by different aquatic species.

383

## 384 **References**

- Abisil, M., Bemtssen, M., Gerringa, L. 1996. The influence of sediment, food and organic ligands
  on the uptake of copper by sediment-dwelling bivalves, Aquatic Toxicology 134, 13-29.
- 387 Amin, N.M., Noor, J.M., Shazili, A.M. 2006. Analysis of Heavy Metals in Soft Tissues of Thais
- 388 Aculeata, a Gastropod Taken from Chendering Beach, Terengganu as an Attempt to Search for
- 389 Indicator of Heavy Metal Pollution in the Aquatic Environment. Proceedings of the 1st International
- 390 Conference on Natural Resources Engineering & Technology, Putrajaya, Malaysia, 54-59.
- 391 Angioni, A., Porcu, L., Secci, M., Addis, P. 2012. QuEChERS Method for the Determination of
- 392 PAH Compounds in Sardinia Sea Urchin (Paracentrotus lividus) Roe, Using Gas Chromatography
- 393 ITMS-MS Analysis. Food Anal. Methods, 5, 1131-1136. DOI 10.1007/s12161-011-9353-7
- 394 Angioni, A., Cau, A., Secci, A., Addis, P. 2014. GC–ITMS analysis of PAH contamination levels in
- the marine sea urchin Paracentrotus lividus in Sardinia. Marine Pollution Bulletin 82, 201-207.
- 396 Atherton, H.J.; Bailey, N.J.; Zhang, W.; Taylor, J. Major, H.; Shockcor, J. Clarke, K.; Griffin, J.L.
- 397 2006. A combined 1H NMR spectroscopy and mass spectrometry-based metabolomic study of the
- 398 PPAR- $\alpha$  null mutant mouse defines profound systemic changes in metabolism linked to the 399 metabolic syndrome. Physiological Genomics. 27, 178-186.
- 400 doi:10.1152/physiolgenomics.00060.2006
- Aydin-Onen, S., Ozturk, M., 2017. Investigation of heavy metal pollution in eastern Aegean Sea
  coastal waters by using *Cystoseira barbata, Patella caerulea*, and *Liza aurata* as biological
  indicators. Environmental Science and Pollution Research 24, 8, 7310-7334. DOI: 10.1007/s11356016-8226-4
- Benfares, R., Seridi, H., Belkacem, Y., Inal, A. 2015. Heavy Metal Bioaccumulation in Brown
  Algae Cystoseira compressa in Algerian Coasts, Mediterranean Sea. Environ. Process. 2, 429–439.
  DOI 10.1007/s40710-015-0075-5
- Blackmore, G. 2000. Field Evidence of Metal Transfer from Invertebrate Prey to an Intertidal
  Predator, *Thais clavigera* (Gastropoda: Muricidae). Estuarine, Coastal and Shelf Science. 51, 2,
  127-139.

- 411 Boucetta, S., Beldi, H., Draredja, B. 2016. Seasonal Variation of Heavy Metals in Phorcus
- 412 (Osilinus) turbinatus (Gastropod, Trochidae) in the Eastern Algerian Coast. Global Veterinaria 17,
- 413 1, 25-41. DOI: 10.5829/idosi.gv.2016.17.01.104129
- Bryan, G.W., Langston, W.J. 1992. Bioavailability, accumulation and effects of heavy metals in
  sediments with special reference to United Kingdom estuaries, a review, Environ Poll 76, 89-102.
- 416 Carral, E., Puente, X., Villares, R., Carballeira, A. 1995. Background heavy metal levels in
- 417 estuarine sediments and organisms in Galicia (northwest Spain) as determined by modal analysis.
- 418 Sci. Total. Environ. 172, 175–188.
- 419 Chiarelli, R., Martino, C., Roccheri, M.C. 2019. Cadmium stress effects indicating marine pollution
- 420 in different species of sea urchin employed as environmental bioindicators. Cell Stress and
- 421 Chaperones 24, 675–687. <u>https://doi.org/10.1007/s12192-019-01010-1</u>
- 422 Collado, C., Ramirez, R., Bergasa, O., Hernandez-Brito, J.J., Gelado-Caballero, M.D., Haroun, R.J.
- 423 2006. Heavy metals (Cd, Cu, Pb, and Zn) in two species of limpets (Patella rustica, Patella candei
- 424 *crenata*) in the canary island. WIT Transactions on Ecology and the Environment DOI:425 10.2495/WP060051
- 426 COMMISSION REGULATION (EC) No 1881/2006 of December 19 2006, setting maximum
- 427 levels for certain contaminants in foodstuff. Official Journal of the European Union. 20.12.2006, L428 364/5.
- 429 Cossa, D. 1989. A review of the use of *Mytilus* spp as quantitative indicator of cadmium and 430 mercury) contamination in coastal waters, Oceanologica Acta 12, 417-432.
- 431 Duysak, Ö., Azdural, K. 2017. Evaluation of Heavy Metal and Aluminium Accumulation in a
- 432 Gastropod, Patella caerulea L., 1758 in Iskenderun Bay, Turkey. Pakistan J. Zool., 49, 2, 629-637.
- 433 DOI: <u>http://dx.doi.org/10.17582/journal.pjz/2017.49.2.629.637</u>
- 434 Farombi, E.O., Adelowo, O.A., Ajimoko, Y.R. 2007. Biomarkers of oxidative stress and heavy
- 435 metal levels as indicators of environmental pollution in African catfish (Clarias gariepinus) from
- 436 Nigeria Ogun River. Int. J. Environ. Res. Public Health, 4, 158-165.

- Fergusson, J. 1990. The heavy elements: chemistry, environmental impact and health effects,Perganlon Press, Oxford.
- 439 Gawad, S.S.A. 2018. Concentrations of heavy metals in water, sediment and mollusc gastropod,
- 440 Lanistes carinatus from Lake Manzala, Egypt. Egyptian Journal of Aquatic Research 44, 77–82.
  441 https://doi.org/10.1016/j.ejar.2018.05.001
- 442 Goldberg, E., Bowen, V., Hodge, Y., Flegal, A., Martin, J. 1983. U.S. mussel watch: 1977-1978.
- 443 Results of trace metals and radionuclides, Estuar. Coast Shelf Sci. 16, 69-93.
- 444 Horwitz, R., Borell, E.M., Fine, M., Shaked, Y. 2014. Trace element profiles of the sea anemone
- 445 Anemonia viridis living nearby a natural CO2 vent. PeerJ 2:e538. DOI 10.7717/peerj.538
- 446 Hummel, H., Moderman, R., Amirad-Triquet, C., Rainglet, F., Duijn, Y., Herssevoost, M., de Jong,
- 447 J., Bogaards, R., Bachelet, G., Desprez, M., Marehand, J., Sylvand, B., Amirad, J.C., Rybarczyk,
- 448 H., de Wolf, L. 1997. A comparative study on the relation between copper and condition in marine
- bivalves and the relation with copper in the sediment, Aquatic Toxicology 38, 165 181.
- 450 Ingston, W.J. 1982. The distribution of mercury in British estuarine sediments and its availability to
- 451 deposit feeding bivalves, J Mar Biol Assoc UK 62, 667-674.
- 452 Jaishankar, M., Tseten, T., Anbalagan, N., Mathew, B. B., Beeregowda, K.N. 2014. Toxicity,
- 453 mechanism and health effects of some heavy metals. Interdiscip Toxicol. 7(2), 60–72. doi:
  454 10.2478/intox-2014-0009
- 455 Jakimska, A., Konieczka, P., Skóra, K., Namieśnik, J. 2011. Bioaccumulation of Metals in Tissues
- of Marine Animals, Part I: The Role and Impact of Heavy Metals on Organisms. Pol. J. Environ.
  Stud. 20, 5, 1117-1125.
- 458 Jitar O., Teodosiu, C., Oros, A., Plavan, G., Nicoara, M. 2015. Bioaccumulation of heavy metals in
- 459 marine organisms from the Romanian sector of the Black Sea. New Biotechnology. 32, 3, 369-378.
- 460 Kamaruzzaman, B.Y., Rina, John, B.A., Jalal, K.C.A. 2011. Heavy Metal Accumulation in
- 461 Commercially Important Fishes of South West Malaysian Coast. Research Journal of
- 462 Environmental Sciences 5, 6, 595-602.

- 463 Kelepertzis, E. 2013. Heavy Metals baseline concentrations in soft tissues of *Patella* sp. From the
- 464 Stratoni coastal environment, NE Greece. Ecol Chem Eng S. 20, 1, 141-149.
- 465 Mitchelmore, C.L., Verde, E.A., Ringwood, A.H., Weis, V.M. 2003. Differential accumulation of
- heavy metals in the sea anemone Anthopleura elegantissima as a function of symbiotic state.Aquatic Toxicology 64, 317-329.
- Perez-Lopez, M., Alonso, J., Novoa-Valinas, M. C., Melgar, M. J. 2003. Assessment of Heavy
  Metal Contamination of Seawater and Marine Limpet, *Patella vulgata* L., from Northwest Spain.
- 470 Journal of Environmental science and Health, Part A—Toxic/Hazardous Substances &
  471 Environmental Engineering 38, 12, 2845–2856.
- Philips, D. 1977. The use of biological indicator organisms to monitor trace metal pollution in
  marine and estuarine environments a review, Environ Pollut 13, 281-317.
- 474 Ramirez, R. 2013. The gastropod Osilinus atrata as a bioindicator of Cd, Cu, Pb and Zn
- 475 contamination in the coastal waters of the Canary Islands. Chemistry and Ecology. 29, 3, 208-220.
  476 https://doi.org/10.1080/02757540.2012.735659
- 477 Rodríguez-Figueroa, G. M., Shumilin, E., Sánchez-Rodríguez, I. 2008. Heavy metal pollution
- 478 monitoring using the brown seaweed Padina durvillaei in the coastal zone of the Santa Rosalía
  479 mining region, Baja California Peninsula, Mexico. J Appl Phycol. 21:19–26. DOI 10.1007/s10811-
- 480 008-9346-0
- 481 Roesijadi, G., Robinson, W.E. 1994. Metal regulation in aquatic animals: mechanisms of uptake,
- 482 accumulation, and release. In Aquatic Toxicology: Molecular, Biochemical, and Cellular
- 483 Perspectives; Malins, D.C., Ostrander, G.K., Eds.; Lewis Publishers: Boca Raton, 387–420.
- 484 Ryabushko, V. I., Prazukin, A. V., Gureeva, E. V., Bobko, N. I., Kovrigina, N. G., Nekhoroshev,
- 485 М. V. 2017. Морской биологический журнал (Marine Biological Journal). 2, 2, 70–79. doi:
  486 10.21072/mbj.2017.02.2.07
- 487 Salvo, A., Giorgi, A.G., Cicero, N., Bruno, M., Lo Turco, V., Di Bella, G., Dugo, G. 2014.
- 488 Statistical characterization of heavy metal contents in Paracentrotus lividus from Mediterranean 19

- 489 Sea. Natural Product research formely Natural Products letters. 1-9. DOI:490 10.1080/14786419.2013.878937
- 491 Vinogrjidov, A., Efron, J., Setlow, J. 1953. Elementary Composition of Coelenterata. In ODUM V.
- 492 (Ed.), Memoir II: The Elementary Chemical Composition of Marine Organisms (pp. 194-219).
  493 NEW HAVEN: Yale University Press. Retrieved March 25, 2020, from
  494 www.jstor.org/stable/j.ctvbcd0gk.11
- 495 Wenzl, T., Haedrich, J., Schaechtele, A., Robouch, P., Stroka, J. 2016. Guidance Document on the
- Estimation of LOD and LOQ for Measurements in the Field of Contaminants in Feed and Food.
  EUR 28099, Publications Office of the European Union, Luxembourg, ISBN 978-92-79-61768-3;
  doi:10.2787/8931
- 499 , Y.J., Yang, Z., Zhang, S. 2011. Ecological risk assessment of heavy metals in sediment and human
  500 health risk assessment of heavy metals in fishes in the middle and lower reaches of the Yangtze
- 501 River basin. Environmental Pollution, Volume 159, Issue 10, October 2011, Pages 2575-2585
- 502 Yüzereroğlu, T.A., Firidin, G.G., Coğun, H.Y., First, O., Aslanyavrsu, S., Maruldali, O., Kargin, F.
- 503 2009. Heavy metals in Patella caerulea (Mollusca, Gastropoda) in polluted and non-polluted areas
- from the Iskenderun Gulf (Mediterranean Turkey). Environmental Monitoring and Assessment 167,
- 505 1-4, 257-64. DOI: 10.1007/s10661-009-1047-x
- 506
- 507
- 508
- 509
- 510
- 511 512

# **Table 1.** Humidity and ash content of the edible part of gastropods, Paracentrotus, anemonia, and of the seaweed sampled.

		Humidity		Ash			
	g/1	$00g \pm RSD F$	W	$g/100g \pm RSD FW$			
	C. Zafferano	C. Spartivento	C. Frasca	C. Zafferano	C. Spartivento	C. Frasca	
	(CZ)	(CS)	(CF)	(CZ)	(CS)	(CF)	
Patella vulgata	79.46 a(a)	78.91 a(a)	78.01 a(a)	4.61 a(a)	4.29 a(a)	5.04 ab(a)	
Osilinus turbinatus	74.44 a(a)	77.54 a(a)	75.62 a(a)	8.91 b(a)	5.10 a(b)	6.59 a(ab)	
Thais haemastoma	76.53 a(a)	74.90 a(a)	76.20 a(a)	5.18 a(a)	3.75 ab(b)	4.24 b(ab)	
Paracentrotus lividus	84.06 b(a)	79.54 a(a)	81.56 b(a)	2.10 c(a)	2.75 b(a)	3.30 b(a)	
Anemonia sulcata	89.06 b(a)	91.44 b(a)	91.96 c(a)	3.50 c(a)	3.85 ab(a)	3.71 b(a)	
Padina pavonica	81.67 b(a)	81.50 c(a)	84.18 b(a)	15.35 d(a)	13.13 c(a)	6.30 a(b)	
Cystoseira mediterranea	70.21 c(a)	68.91 d(a)	78.44 b(b)	10.85 e(a)	12.86 c(a)	10.48 c(a)	

<sup>a</sup> Values among species (without brackets) or among sites (in brackets) followed by unlike letters differ significantly by Fisher's least significant difference (LSD) procedure,  $P \leq 0.05$ . Letters without brackets relate to comparisons of the influence of species within each site. Letters in brackets relate to comparisons of the influence of the sites among each species for each parameter.

 Table 2. Heavy metals method parameters.

Element $\lambda$		MRL µg/g	LOQ	Linear regression equation	$\mathbb{R}^2$
Al			0.10 y=670.89x-45.		0.9978
AS	188.98		0.10	y=79.38x-10.14	0.9991
В	249.77		0.50	y=5151.33x+789.8	0.9995
Ba	493.40		0.05	y=8035x-289.9	0.9993
Be	313.04		0.005	y=259128x-6848	0.9985
Cd	226.50	0.05	0.005	y=4804x-3.55	0.9995
Со	228.61		0.025	y=6622.7x+58.6	0.9997
Cr	267.71		0.01	y=9908.7x+97.7	0.9994
Cu	324.75		0.01	y=17117.2x-95.93	0.9996
Fe	259.94		0.10	y=3059.9x+80.8	0.9992
Hg	194.16	0.5	0.05	y=392.2x+11	0.9994
Mn	257.61		0.01	y=70780.6x-8029	0.9990
Мо	204.59		0.05	y=604.4x-5.5	0.9988
Ni	216.55		0.05	y=1642.2x-22.3	0.9986
Pb	220.35	0.3	0.05	y=297.3x-19.4	0.9967
Sb	217.58		0.10	y=49.7x-5.8	0.9981
Se	196.02		0.025	y=157.7x-11.5	0.9992
Sn	189.92		0.05	y=74.3x-4.9	0.9986
Sr	407.77		0.10	y=829945.3x+17943	0.9984
Те	214.28		0.10	y=106.6x-10.1	0.9972

Ti	336.12	0.01	y=14275x+521.8	0.9976
v	292.40	0.005	y=13753.2x+231.2	0.9996
Zn	213.85	0.05	y=4785.9x-316	0.9986

Table 3. Heavy	y metals ( $\mu g/g \pm RSD\%$	) in gastropods	from the three	sites investigated.

metal		Patella vulgata		(	Osilinus turbinatus			Thais		
	C. Zafferano	C. Spartivento	C. Frasca	C. Zafferano	C. Spartivento	C. Frasca	C. Zafferano	C. Spartivento	C. Frasca	
	(CZ)	(CS)	(CF)	(CZ)	(CS)	(CF)	(CZ)	(CS)	(CF)	
Al	6.44a(a)	1,41b(b)	1.35b(b)	8,77 a(a)	2,94 b(c)	2.80 b(c)	0,37B a(d)	0,18 b(e)	0.15 b(e)	
AS	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
В	1,63a(a)	1,44a(a)	1.37a(a)	1,54 a(a)	1,35 a(a)	1.40 a(a)	1,14 a(b)	0,90 a(b)	0.92 a(b)	
Ba	1,21a(a)	0,52b(b)	0.48b(b)	1,98 a(c)	1,63 a(c)	1.57 a(c)	10,04 a(d)	7,07 b(e)	6.58 b(e)	
Be	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Cd	0,19a(a)	0,23a(a)	0.20a(a)	0,08 a(b)	0,10 a(b)	0.09 a(b)	0,90 a(c)	1,14 a(c)	1.01 a(c)	
Co	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Cr	0,03a(a)	0,03a(a)	0.03a(a)	0,11 a(a)	0,03 b(a)	0.03 b(a)	0,04 a(a)	0,03 a(a)	0.03 a(a)	
Cu	0,28a(a)	0,26a(a)	0.27a(a)	2,29 a(a)	1,64 b(a)	1.58 b(a)	29,68 a(a)	8,71 b(a)	7.58 b(a)	
Fe	34,28a(a)	7,37b(b)	6.58b(b)	40,22 a(a)	15,36 b(c)	14.91 b(c)	36,89 a(a)	10,73 b(d)	11.51 b(d)	
Hg	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Mn	0,33a(a)	0,07b(b)	0.05b(b)	0,69 a(c)	0,23 b(a)	0.21 b(a)	0,30 a(a)	0,36 a(a)	0.32 a(a)	

Mo	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Ni	< LOQ	0,13a(a)	0.11a(a)	7,93 a(b)	0,26 b(c)	0.25 b(c)	< LOQ	0,08 a(a)	0.05 a(a)
Pb	0,50a(a)	0,07b(b)	0.08b(b)	0,18 a(c)	< LOQ	< LOQ	0,46 a(a)	0,16 b(c)	0.14 b(c)
Sb	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Se	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Sn	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Sr	7,06a(a)	1,57b(b)	1.65b(b)	19,08 a(c)	8,04 b(a)	7.58 b(a)	6,01 a(a)	6,10 a(a)	6.05 a(a)
Te	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Ti	0,13a(a)	0,07b(b)	0.06b(b)	0,18 a(a)	0,09 b(b)	0.10 b(b)	< LOQ	< LOQ	< LOQ
V	0,28a(a)	0,30a(a)	0.31a(a)	0,60 a(b)	0,59 a(b)	0.54 a(b)	0,06 a(c)	0,51 b(b)	0.48 b(b)
Zn	1,69a(a)	1,88a(a)	1.75a(a)	0,62 a(b)	2,16 b(c)	1.98 b(c)	21,91 a(d)	20,31 a(d)	21.03 a(d)

<sup>a</sup> Values among species (without brackets) or among sites (in brackets) followed by unlike letters differ significantly by Fisher's least significant difference (LSD) procedure,  $P \leq 0.05$ . Letters without brackets relate to comparisons of the influence of species within each site. Letters in brackets relate to comparisons of the influence of the sites among each metal.

# Table 4. Heavy metals seaweeds.

Element	1	Padina pavonica		Cystoseira mediterranea			
	C. Zafferano	C. Spartivento	C. Frasca	C. Zafferano	C. Spartivento	C. Frasca	
	(CZ)	(CS)	(CF)	(CZ)	(CS)	(CF)	
Al	104,40a(a)	9,13b(b)	7,30b(b)	41,85a(c)	3,85b(d)	15,09c(e)	
AS	0,73a(a)	0,25b(b)	0,26b(b)	1,06a(c)	2,56a(c)	1,05a(c)	
В	< LOQ	0,74a(a)	< LOQ	12,60a(b)	5,10b(c)	< LOQ	
Ba	9,96a(a)	3,52b(b)	0,68c(c)	32,08a(d)	17,77b(e)	4,46c(b)	
Be	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Cd	0,07a(a)	0,01b(b)	0,02b(b)	0,08a(a)	0,02b(b)	0,08a(a)	
Co	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Cr	0,18a(a)	0,02b(b)	0,06b(b)	0,15a(a)	0,03b(b)	0,24c(ac)	
Cu	0,42a(a)	0,10b(b)	0,15b(b)	0,79a(c)	0,18b(b)	0,15b(b)	
Fe	225,46a(a)	11,41b(b)	15,79b(b)	184,03a(a)	3,94b(c)	3,56b(c)	

Hg	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Mn	2,88a(a)	0,84b(b)	< LOQ	6,58a(c)	1,40b(a)	0,73c(b)
Mo	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Ni	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Pb	0,71a(a)	0,11b(b)	0,25c(c)	0,58a(a)	0,13b(b)	0,92c(d)
Sb	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Se	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Sn	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Sr	77,46a(a)	5,39b(a)	5.65b (b)	55,46a(a)	1,53b(c)	1,15b(c)
Te	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ
Ti	2,60a(a)	0,20b(b)	0,92c(c)	1,18a(c)	0,03b(d)	4,41c(e)
V	0.72a(a)	0,21b(b)	0,25b(b)	1.10a(a)	0,52b(c)	1,50c(d)
Zn	104,40a(a)	0,23b(b)	11,42c(c)	41,85a(d)	1,50b(e)	2,97c(f)

<sup>a</sup> Values among species (without brackets) or among sites (in brackets) followed by unlike letters differ significantly by Fisher's least significant difference (LSD) procedure,  $P \le 0.05$ . Letters without brackets relate to comparisons of the influence of species within each site. Letters in brackets relate to comparisons of the influence of the sites among each metal.

 Table 5. Heavy metals in echinoderms, and coelenterate.

Element	Paracentrotus lividus			Anemonia sulcata			
	C. Zafferano	C. Spartivento	C. Frasca	C. Zafferano	C. Spartivento	C. Frasca	
	(CZ)	(CS)	(CF)	(CZ)	(CS)	(CF)	
Al	5.64a(a)	0,27a(a)	1,74a(a)	2,21a(a)	1.95a(a)	2.05a(a)	
AS	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
В	1.25a(a)	0,53a(a)	1,27a(a)	0,63a(a)	0.59a(a)	0.60a(a)	
Ba	0.98a(a)	0.25a(a)	0,40a(a)	0,13a(a)	0.12a(a)	0.10a(a)	
Be	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Cd	0.06a(a)	0,04a(a)	0,03a(a)	< LOQ	< LOQ	< LOQ	
Co	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Cr	0.02a(a)	0,04a(a)	0,02a(a)	< LOQ	< LOQ	< LOQ	
Cu	0.27a(a)	0,14a(a)	0,17a(a)	0,20a(a)	0.12a(a)	0.16a(a)	
Fe	4.25a(a)	1,85a(a)	4,38a(a)	4,07a(a)	3.81a(a)	2.95a(a)	
Hg	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Mn	< LOQ	< LOQ	< LOQ	0,22a(a)	0.35a(a)	0.27a(a)	
Mo	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Ni	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Pb	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Sb	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Se	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Sn	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Sr	5.64a(a)	0,79a(a)	2,89a(a)	0,64a(a)	0.29a(a)	0.31a(a)	
Te	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	< LOQ	
Ti	0.09a(a)	0,02a(a)	0,08a(a)	< LOQ	< LOQ	< LOQ	
V	0.12a(a)	0,08a(a)	0,10a(a)	0,02a(a)	0.03a(a)	0.02a(a)	
Zn	9.24a(a)	4,87a(a)	2,47a(a)	0,62a(a)	0.22a(a)	0.13a(a)	

670
671
<sup>a</sup> Values among species (without brackets) or among sites (in brackets) followed by unlike letters differ significantly by Fisher's least significant difference (LSD) procedure, P ≤ 0.05. Letters without brackets relate to comparisons of the influence of species within each site. Letters in brackets relate to comparisons of the influence of the sites among each metal.



**Figure 1**. Localization of the three sites in Sardinia, main current **NW**, and **•** industrial area.

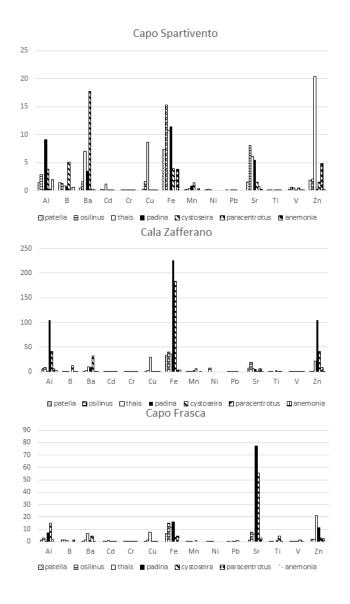


Figure 2. Histograms of the distribution of the metals among the species in the three locations.

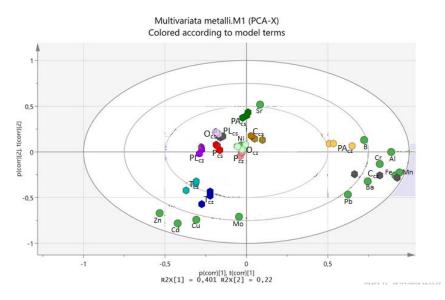




Figure 3. PCA biplot of the score and loading in CZ and CS sites.