



Review

The dairy biorefinery: Integrating treatment processes for cheese whey valorisation

Fabiano Asunis^{a,b}, Giorgia De Gioannis^{a,c}, Paolo Dessì^{b,*}, Marco Isipato^{a,b}, Piet N.L. Lens^b, Aldo Muntoni^{a,c}, Alessandra Poletti^d, Raffaella Pomi^d, Andreina Rossi^d, Daniela Spiga^a

^a DICAAR - Department of Civil and Environmental Engineering and Architecture, University of Cagliari, Piazza D'Armi 1, 09123, Cagliari, Italy

^b Microbiology, School of Natural Sciences and Ryan Institute, National University of Ireland Galway, University Road, Galway, H91 TK33, Ireland

^c IGAG-CNR, Environmental Geology and Geoengineering Institute of the National Research Council - Piazza D'Armi 1, 09123, Cagliari, Italy

^d Department of Civil and Environmental Engineering, University of Rome "La Sapienza", Via Eudossiana 18, 00184, Rome, Italy



ARTICLE INFO

Keywords:

Anaerobic digestion
Bioplastic
Circular economy
Dairy industry
Fermentation
Microbial electrochemical technology

ABSTRACT

With an estimated worldwide production of 190 billion kg per year, and due to its high organic load, cheese whey represents a huge opportunity for bioenergy and biochemicals production. Several physical, chemical and biological processes have been proposed to valorise cheese whey by producing biofuels (methane, hydrogen, and ethanol), electric energy, and/or chemical commodities (carboxylic acids, proteins, and biopolymers). A biorefinery concept, in which several value-added products are obtained from cheese whey through a cascade of biotechnological processes, is an opportunity for increasing the product spectrum of dairy industries while allowing for sustainable management of the residual streams and reducing disposal costs for the final residues. This review critically analyses the different treatment options available for energy and materials recovery from cheese whey, their combinations and perspectives for implementation. Thus, instead of focusing on a specific valorisation platform, in the present review the most relevant aspects of each strategy are analysed to support the integration of different routes, in order to identify the most appropriate treatment train.

1. Introduction

Fossil fuels, which include coal, oil and natural gas, supply about 80% of the world total energy (International Energy Agency, 2017). These non-renewable sources provide electricity, heat, and transportation fuels, as well as supply raw materials and platform chemicals for the manufacturing of a wide range of products. Fossil fuels and industrial processes, on the other hand, account for 65% of the global greenhouse emissions to the atmosphere (IPCC, 2014). In 2015, the increased awareness of climate change issues led to the Paris agreement, in which 195 countries committed themselves to reducing their greenhouse gases emissions by 40% by 2030. Achieving such an ambitious target requires a shift from fossil fuels to renewable sources for energy and chemicals production. Among renewable resources, biodegradable waste streams are a promising source of green energy and (bio)-chemicals.

Recently, the awareness of the unexploited potential of waste has increasingly driven the industrial sector to implement integrated systems, the so-called biorefineries, to produce not only biofuels, but also a

wide spectrum of bio-based chemicals from organic by-products and waste streams (Cherubini, 2010; Mohan et al., 2016; Moscoviz et al., 2018). Such a transition is fully in line with the efforts the EU is making towards a circular bioeconomy (European Commission, 2018) as well as its commitment to becoming the first climate-neutral area in the world by 2050 (European Commission, 2020).

Among the business areas producing waste and wastewater potentially suitable for biorefineries, the dairy sector plays a significant role in the EU economy and many dairy companies are making tremendous efforts to meet the European environmental protection measures and targets (European Dairy Association, 2019). In the EU, a total of 170 billion kg of milk was produced in 2017, 93% of which was converted into dairy products including cheese (37%), butter (30%), cream (13%), fresh milk (11%), acidified milk (4%), milk powder (2%), and other minor products (Eurostat, 2018). Dairy industries produce an average of 2.5 L of wastewater per L of processed milk, as well as about 9–10 L of cheese whey (CW) per kg of cheese produced, resulting in approximately 400 billion L of wastewater per year (Eurostat, 2018). Dairy effluents, and CW in particular, are characterised by a high organic load representing, at the same time, a severe hazard for the environment and a

* Corresponding author.

E-mail addresses: paolo.dessi@nuigalway.ie, paolo.dessi89@yahoo.it (P. Dessì).

<https://doi.org/10.1016/j.jenvman.2020.111240>

Received 10 June 2020; Received in revised form 13 August 2020; Accepted 15 August 2020

Available online 28 August 2020

0301-4797/© 2020 The Authors. Published by Elsevier Ltd. This is an open access article under the CC BY license (<http://creativecommons.org/licenses/by/4.0/>).

Abbreviations			
AD	Anaerobic digestion	MEC	Microbial electrolysis cell
AFBR	Anaerobic fluidized bed reactors	MET	Microbial electrochemical technology
AS	Activated sludge	MFC	Microbial fuel cell
ASTBR	Anaerobic structured-bed reactor	OLR	Organic loading rate
BMP	Biomethane potential	PABR	Periodic anaerobic baffled reactor
BOD	Biological oxygen demand	PBR	Packed bed reactor
CE	Coulombic efficiency	PEM	Proton exchange membrane
COD	Chemical oxygen demand	PHA	Polyhydroxyalkanoates
CSTR	Continuous stirred tank reactor	PHB	Polyhydroxybutyrate
CW	Cheese whey	PHV	Polyhydroxyvalerate
DF	Dark fermentation	PLA	Poly(lactic acid)
FBR	Fluidized bed reactor	PTFE	Polytetrafluoroethylene
GDL	Gas diffusion layer	SBR	Sequence batch reactor
HAc	Acetic acid	TOC	Total organic carbon
HBu	Butyric acid	TRL	Technology readiness level
HCa	Caproic acid	UASB	Upflow anaerobic sludge blanket
HPr	Propionic acid	UFAF	Up-flow anaerobic filter
HRT	Hydraulic retention time	VFAs	Volatile fatty acids
		VS	Volatile solids
		VSS	Volatile suspended solids

huge opportunity for bioenergy and biochemicals production (Ahmad et al., 2019).

Currently, a large share of dairy effluents, including about 50% of the CW produced worldwide, is discharged into the environment without any treatment (Bosco et al., 2018; Slavov, 2017). Among the available treatment options, traditional activated sludge processing is economically not sustainable due to the high organic load of dairy effluents, and the consequent huge quantities of both oxygen required for aeration and excess sludge produced. Activated sludge treatment consumes an average of 900 kWh_(el) d⁻¹, including 100 kWh_(el) d⁻¹ for dewatering (using a filter press) and 800 kWh_(el) d⁻¹ for aerobic stabilization, accounting for 30% of the total energy required for aerobic treatment of dairy effluents (Dąbrowski et al., 2017). Thermo-catalytic treatment has also been proposed for CW valorisation (Remón et al., 2016), but the high temperature required (450–600 °C) and the production of solids make such a process expensive.

Bioprocesses such as anaerobic digestion or fermentation, as well as biological production of polymers and bioelectrochemical systems, have the advantage of coupling the treatment of dairy effluents with the production of bioenergy and/or biochemical commodities at mild temperature conditions, typically within the range 20–55 °C. Though promising, none of the mentioned options alone represents the ultimate solution for CW treatment, since the energy/chemicals production rates are too small for an economically sustainable scale-up. The implementation of an integrated process, including a combination of physical, chemical and biological processes, is therefore the key for a cost-effective and efficient valorisation of dairy effluents. The aim of this review is to summarize and critically discuss the progress made towards the implementation of biorefineries for energy and chemicals recovery in dairy industries, with a specific focus on CW. This review provides an insight into the most promising biorefinery models for resource recovery from CW, based on critical considerations on the potentials, prospects and limitations of the available options, to support the creation of an innovative and scalable industrial chain.

2. Cheese whey characterisation

Cheese production usually generates three different waste streams, including CW and secondary CW (from cheese and ricotta/cottage cheese production, respectively), and dairy wastewater (from washing of tanks and equipment) (Fig. 1).

CW is a green-yellow by-product of cheese and casein powder

production, with an estimated worldwide production of about 190 billion kg year⁻¹. Due to its high organic and volumetric load, CW is considered the main polluting waste stream in dairy industries (Ryan and Walsh, 2016; Slavov, 2017). The CW composition depends on the cheese production process, on the milk source (sheep, goat, cow or buffalo), as well as on the quantity of water, detergents and sanitizing agents used (Demirel et al., 2005; Shete and Shinkar, 2013). In general, CW accounts for 85–95% of the milk volume, retains 55% of milk nutrients (vitamins and minerals) and 20% of milk proteins, and is characterised by COD and BOD concentrations of 50–102 and 27–60 g L⁻¹, respectively, more than 90% of which is made up of lactose (Carvalho et al., 2013; Ryan and Walsh, 2016). CW also contains sodium, potassium and calcium salts (0.46–10%), and has a pH of 3.8–6.5 depending on the whey type (acidic or sweet), and a low alkalinity (Prazeres et al., 2012). More details can be found in the comprehensive review by Carvalho et al. (2013).

CW can be processed to obtain cottage, curd, or ricotta cheese, generating secondary CW as a by-product. Secondary CW retains about 60% of the dry matter contained in CW, and is characterised by a lower protein concentration and a higher salinity because of the second flocculation step and addition of salts in the manufacturing process (Carvalho et al., 2013). Dairy wastewater contains similar compounds as CW, but at lower concentrations. Another waste stream, whey permeate,

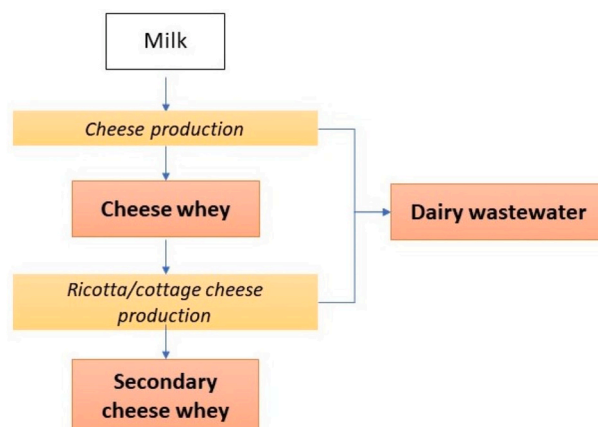


Fig. 1. Cheese production process and effluents generated.

can be obtained as a by-product of protein recovery from CW by ultrafiltration. CW permeate retains about 80% of the lactose contained in the original CW, has a high salinity and low concentration of proteins and fats, depending on the efficiency of the ultrafiltration process (Bosco et al., 2018).

3. Biotechnologies for bioenergy and biochemicals production from cheese whey

3.1. Anaerobic digestion

Anaerobic digestion (AD) is a well-established process to exploit the energy content of CW (De Gioannis et al., 2017; Traversi et al., 2013). However, due to the high organic load and low alkalinity of CW, AD may result in the accumulation of volatile fatty acids (VFAs). This leads to acidification and inhibition of the methanogenic activity, adversely affecting the CH₄ yield and process stability (De Gioannis et al., 2014; Escalante-Hernández et al., 2017; Hagen et al., 2014; Prazeres et al., 2012; Traversi et al., 2013). As a consequence, relatively low biomethane potentials (BMPs) ranging from 270 to 600 L CH₄ kg⁻¹ VS have been reported for AD of CW under mesophilic conditions (35–37 °C) (Escalante et al., 2018; Labatut et al., 2011; Vivekanand et al., 2018), implying that HRT values >5 days are required in continuously operated AD systems to prevent process instability (Table 1).

In AD, alkali addition or dilution is generally required to mitigate acidification, but both strategies increase the operational costs, and/or the volumes to be treated. A more economic option is co-digesting CW with substrates characterised by a high buffering capacity, such as sewage sludge (Carrieri et al., 1993), dairy manure (Kavacik and Topaloglu, 2010; Rico et al., 2015; Vivekanand et al., 2018), poultry manure (Gelegenis et al., 2007), cattle slurry (Comino et al., 2012), or fish ensilage (Vivekanand et al., 2018), although literature results are controversial. Furthermore, when digesting CW in combination with pathogenic waste streams, health and safety issues may hamper the use of the digestate as a fertilizer. Labatut et al. (2011) reported that co-digestion of CW with dairy manure, at 10:90 or 25:75 ratios, resulted in a lower CH₄ yield (238–252 L kg⁻¹ VS) than with raw CW (424 L kg⁻¹ VS). Vivekanand et al. (2018) also reported a decreased CH₄ yield when blending CW with cattle manure, fish ensilage, or both. On the other

hand, Comino et al. (2012) obtained the highest CH₄ yield of 343 L CH₄ kg⁻¹ VS co-digesting 50% CW and 50% cattle slurry at 35 °C and 42 d HRT. Hublin and Zelić (2013) reported a maximum CH₄ yield of 288 L kg⁻¹ VS by co-digestion of CW and cow manure at 55 °C, with an optimum mixing ratio of 10:90, and addition of 5 g NaHCO₃ L⁻¹ for alkalinity control.

In co-digestion, not only the maximum CH₄ yield, but also the process stability, may be affected by the mixing ratio of substrates. When co-digesting CW and diluted poultry manure in a continuous-flow stirred tank reactor (CSTR), Gelegenis et al. (2007) reported an increasing CH₄ yield for mixtures with CW concentrations up to 35%, but the process became unstable when the CW fraction exceeded 50% (based on VS). In contrast, when co-digesting CW and the screened liquid fraction of dairy manure, Rico et al. (2015) reported an increase in the CH₄ yield from 161 to 187 L CH₄ kg⁻¹ COD when increasing the CW proportion from 15 to 85% at 35 °C and 15.6 d HRT, with no instability concerns.

A two-stage process, where hydrolysis-acetogenesis and methanogenesis are carried out in two separate reactors, is another strategy to avoid process instability (Fernández et al., 2015) and enhance COD removal and CH₄ production (Bertin et al., 2013; Yazar et al., 2016), although with increased capital and operational costs. Another advantage of two-stage AD is the possibility of operating the methanogenic reactor at lower HRTs (<5 d) compared to the single-stage process. More innovative systems involve a two-stage process in which H₂ is recovered in the acidogenic reactor (see Section 3.2), which can be used as a fuel, either alone or in combination with CH₄ (hythane). Yilmazer and Yenigün (1999) and Saddoud et al. (2007) reported a biogas yield of 550 and 300 L kg⁻¹ COD_{removed}, respectively, with COD removal efficiencies above 90%, in a two-stage AD process with 4 d HRT in the methanogenic reactor. With a HRT of 4.4 d, Antonopoulou et al. (2008) obtained a CH₄ yield of 75.6 L CH₄ d⁻¹ (or 383 L CH₄ kg⁻¹ COD_{feed}), notably higher than that obtained by Venetsaneas et al. (2009) with a CSTR at 20 d HRT (1 L CH₄ d⁻¹ or 134 L CH₄ kg⁻¹ COD_{feed}). Fernández et al. (2015) compared single- and two-stage AD of CW under thermophilic conditions (55 °C), reporting for the former a maximum yield of 349 L CH₄ kg⁻¹ COD_{feed} at 8.3 d HRT, whereas for the latter an inhibition effect at a HRT < 12.5 d. This suggests that two-stage processes may not be optimal for thermophilic AD.

Table 1

Continuous methane production from CW, either as the sole substrate or in co-digestion, in single-stage or two-stage (acidogenesis and methanogenesis in separate reactors) AD processes.

Process	Substrate	Inoculum	Reactor ^a	T (°C)	pH	HRT (d)	Methane production	COD removal (%)	Reference
One-stage	50% CW	None	CSTR	35	6.9–8.7	42	187 L CH ₄ kg ⁻¹ COD _{feed}	82	Comino et al. (2012)
	50% cattle slurry (v/v)	Granular anaerobic cultures	UASB	35	n.a. ^c	2–4.95	424 L CH ₄ kg ⁻¹ COD _{feed}	95–97	Erguder et al. (2001)
	2 L CW + 1 kg Dairy manure + 1 L water	None	CSTR	34	6.5–7.5	5	0.9 L CH ₄ L ⁻¹ d ⁻¹	n.a. ^c	Kavacik and Topaloglu (2010)
	85% CW	None	CSTR	35	6.4–7.1	15.6	392 L CH ₄ kg ⁻¹ VS _{feed}	n.a. ^c	Rico et al. (2015)
Two-stage	15% liquid fraction of dairy manure (v/v)	None	CSTR	35	6.4–7.1	15.6	392 L CH ₄ kg ⁻¹ VS _{feed}	n.a. ^c	Rico et al. (2015)
	CW	None	PABR	35	8.0	4.4	383 L CH ₄ kg ⁻¹ COD _{removed} ^b	94	Antonopoulou et al. (2008)
	CW	None	SBR	55	n.a. ^c	25	349 L CH ₄ kg ⁻¹ COD _{feed} ^b	n.a. ^c	Fernández et al. (2015)
	Diluted CW	Anaerobic sludge	CSTR	37	7.3–8.5	4	300 L biogas kg ⁻¹ COD _{removed} ^b	99	Saddoud et al. (2007)
	CW	Anaerobic sludge	CSTR	35	7.7	20	134 L CH ₄ kg ⁻¹ COD _{feed} ^b	95	Venetsaneas et al. (2009)
CW	Anaerobic sludge	UFAF	n.a. ^c	n.a. ^c	4	550 L biogas kg ⁻¹ COD _{removed} ^b	90	Yilmazer and Yenigün (1999)	

^a CSTR, continuously stirred tank reactor; PABR, periodic anaerobic baffled reactor; SBR, sequencing batch reactor; UASB, upflow anaerobic sludge blanket; UFAF, up-flow anaerobic filter.

^b For two-stage processes, it refers to the COD of the acidogenic effluent rather than the initial substrate.

^c Not available.

3.2. Fermentative processes

3.2.1. Dark fermentation

Dark fermentation (DF) is a promising option for CW valorisation due to its high carbohydrate content, which can be converted to biohydrogen and VFAs (Akhlaghi et al., 2019; Asunis et al., 2019; De Gioannis et al., 2014). In the absence of CW pre-treatments and external inoculum, DF of CW mainly involves three steps, including (i) lactose hydrolysis into glucose and galactose, (ii) conversion of monomeric sugars into lactate by homolactic microorganisms, such as *Lactobacillus*, and (iii) conversion of lactate into H_2 and VFAs by fermentative microorganisms, such as *Clostridium* (Fig. 2).

A theoretical maximum yield of $8 \text{ mol } H_2 \text{ mol}^{-1}$ lactose can be obtained by DF, if acetate is the only soluble reaction product. However, DF is sensitive to substrate composition, organic loading rate, inoculum type and pre-treatment, reactor type and operation regime, temperature, pH, hydraulic and cell residence time (Akhlaghi et al., 2017). This results in actual H_2 yields between 1 and $4 \text{ mol } H_2 \text{ mol}^{-1}$ lactose, accompanied by the production of a mixture of VFAs, mainly acetic, propionic, and butyric acid (Table 2).

Inocula of different origin, including pure cultures, anaerobic sludge, activated sludge, and compost, with or without pre-treatment, have been used in DF of CW (Table 2). However, some studies relied exclusively on the indigenous biomass of CW (Akhlaghi et al., 2017; Antonopoulou et al., 2008; De Gioannis et al., 2014; Montecchio et al., 2018; Venetsaneas et al., 2009), reporting as high H_2 yields as those obtained using external inocula. De Gioannis et al. (2014) compared batch DF of CW with pre-treated activated sludge and without an external inoculum, obtaining similar yields of $160\text{--}170 \text{ L } H_2 \text{ kg}^{-1} \text{ TOC}$ at pH 6–6.5. To achieve faster start-up, pre-fermented CW can be used as the inoculum in large-scale plants in place of methanogenic inocula that require chemical or thermal pre-treatment to inhibit methanogenesis. However, Perna et al. (2013) obtained a yield of only $0.7 \text{ mol } H_2 \text{ mol}^{-1}$ lactose ($40 \text{ L } H_2 \text{ kg}^{-1} \text{ COD}_{\text{lactose}}$) when using fermented CW as the inoculum in a packed bed reactor (PBR), with a relatively high production of acetate (10 g L^{-1}), which suggests the onset of H_2 -consuming homoacetogenic pathways. Among pure cultures, both *Clostridium saccharoperbutyacetonicum* (Ferchichi et al., 2005) and *Escherichia coli* (Rosales-Colunga et al., 2010) yielded $2.7 \text{ mol } H_2 \text{ mol}^{-1}$ lactose ($158 \text{ L } H_2 \text{ kg}^{-1} \text{ COD}_{\text{lactose}}$) from diluted CW and CW powder, respectively, therefore of the same order of magnitude as mixed cultures. The use of pure

cultures, which increases operational costs, does not seem a cost-effective approach for CW fermentation when H_2 is the desired end product.

Various CW-based substrates have been used for DF. Raw CW is easily degraded by indigenous bacteria, even at 4°C , making storage difficult (Tribst et al., 2019). Thus, many studies used re-hydrated CW powder (Table 2), adjusting the water content to restore the original content of raw CW. Addition of bicarbonate was proposed to prevent acidification (Perna et al., 2013), although co-digestion with an alkaline substrate such as manure can also be done (Ghimire et al., 2017). Dilution of CW can prevent acidification of the fermentation broth, thus increasing the H_2 yields, but also drastically increasing the already huge amount of wastewater to be treated. Furthermore, dilution of CW would reduce the concentration of micro and macro nutrients available to the microorganisms. Yields above $3 \text{ mol } H_2 \text{ mol}^{-1}$ lactose, and acetate and isobutyrate concentrations above 5 g L^{-1} were obtained by supplementing CW with micronutrients such as calcium (Azbar et al., 2009a), whereas yields below $2 \text{ mol } H_2 \text{ mol}^{-1}$ lactose ($117 \text{ L } H_2 \text{ kg}^{-1} \text{ COD}_{\text{lactose}}$), as well as low VFA concentrations were obtained from deproteinized or ultrafiltered CW (Fernández et al., 2015; Montecchio et al., 2018). Since this was likely due to the lack of nitrogen to support microbial growth, excessive dilution or inclusion of a protein recovery step before DF of CW are not recommended.

Bioreactors with high biomass retention, such as fluidized bed reactors (FBR) (Ferreira Rosa et al., 2014a, 2014b; Ottaviano et al., 2017), or sequencing batch reactors (SBR) (Fernández et al., 2015) can be advantageous for DF of CW, compared to CSTRs, as much lower HRTs can be applied (Table 2). However, too short HRTs, below 4 h, may decrease the H_2 yield (Ferreira Rosa et al., 2014a). Among the operating parameters, pH has the strongest impact on both H_2 yield and VFA production. An optimum pH between 5.5 and 6.5 for H_2 production from CW under mesophilic conditions was identified in several studies (Asunis et al., 2019; Azbar et al., 2009b; Davila-Vazquez et al., 2008; De Gioannis et al., 2014; Ferchichi et al., 2005). However, an optimum pH of 4.5 was reported under thermophilic conditions by Azbar et al. (2009b). Ottaviano et al. (2017) obtained a remarkable yield of $3.67 \text{ mol } H_2 \text{ mol}^{-1}$ lactose ($214 \text{ L } H_2 \text{ kg}^{-1} \text{ COD}_{\text{lactose}}$) from diluted CW in a thermophilic (55°C) FBR operated at pH 4–4.5 and HRT of 4 h.

Besides H_2 production, the pH affects the yield and spectrum of VFAs, so that the operating conditions of DF reactors can be adjusted to target specific VFAs (or a mixture of them), in combination with or as an

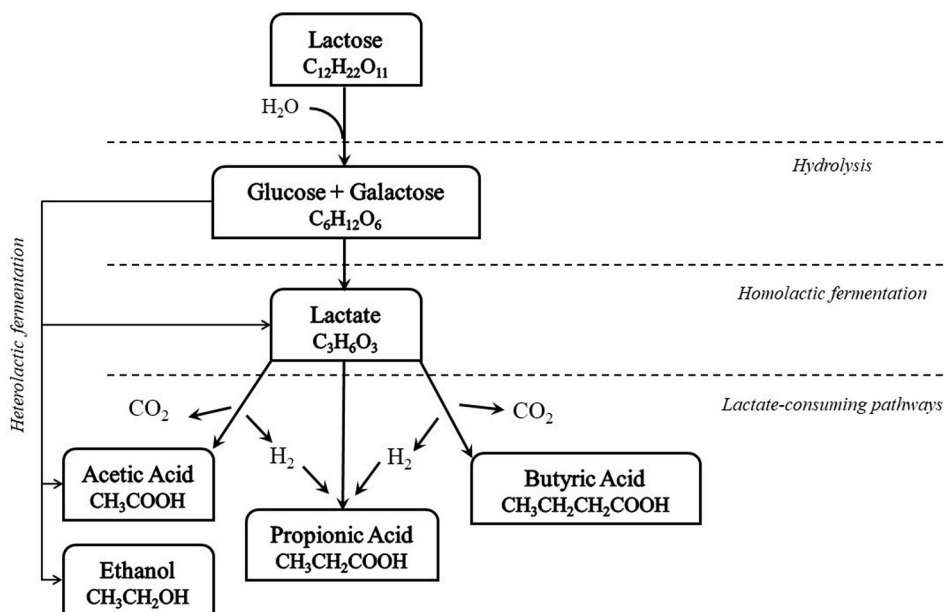


Fig. 2. Most common lactose fermentation pathways from CW indigenous microorganisms.

Table 2
Continuous hydrogen and VFA production from CW-based substrates.

Substrate	Inoculum	Reactor ^a	T (°C)	pH	OLR	HRT (h)	H ₂ yield	VFA ^b production	Reference
Cheese whey	None	CSTR	35	5.2 (at steady state)	n.a. ^c	24	41 L H ₂ kg ⁻¹ COD	HAc: 9.4 g L ⁻¹ HBu: 7.2 g L ⁻¹	Antonopoulou et al. (2008)
Cheese whey powder	Pre-fermented cheese whey	ASTBR	25	5.1 (average)	24 g _{COD} L ⁻¹ d ⁻¹	24	82 L H ₂ kg ⁻¹ COD _{lactose}	HAc: 5.0 g L ⁻¹ HBu: 3.0 g L ⁻¹	Blanco et al. (2019)
Cheese whey	Acidogenic sludge	UASB	30	5.0 (average)	10–20 g _{COD} L ⁻¹ d ⁻¹	24–12	122 mL H ₂ L ⁻¹ d ⁻¹	n.a. ^c	Castelló et al. (2009)
Cheese whey	Kitchen waste compost	CSTR	30	5.5 (controlled)	29 g _{COD} L ⁻¹ d ⁻¹	24	52 L H ₂ kg ⁻¹ COD _{lactose}	HAc: 3.0 g L ⁻¹ HBu: 1.6 g L ⁻¹	Castelló et al. (2018)
Cheese whey powder	Pretreated anaerobic granular sludge	CSTR	37	5.9 (controlled)	92.4–184.4 g _{lactose} L ⁻¹ d ⁻¹	4–10	163 L H ₂ kg ⁻¹ COD _{lactose}	HAc: 4.5 g L ⁻¹ HPr: 6.2 g L ⁻¹ HBu: 10.6 g L ⁻¹	Davila-Vazquez et al. (2009)
Cheese whey	Pretreated digested sludge	UASB	35	5.0 (controlled)	20–80 g _{COD} L ⁻¹ d ⁻¹	24	40 L H ₂ kg ⁻¹ COD	HAc: 9.1 g L ⁻¹ HBu: 13.5 g L ⁻¹	Dessi et al. (2020)
Deproteinized cheese whey	None	SBR	35	4.5–5.5 (pulse controlled) ^d	12.7–25.3 g _{COD} L ⁻¹ d ⁻¹	1.5–3.0	12 L H ₂ kg ⁻¹ COD	HAc: 2.3–3.4 g L ⁻¹ HPr: 1.0 g L ⁻¹ HBu: 0.5 g L ⁻¹	Fernández et al. (2015)
Cheese whey powder supplemented with medium	Pretreated anaerobic granular sludge	AFBR	30	4.0–4.5 (controlled)	30–120 g _{COD} L ⁻¹ d ⁻¹	1–4	77 L H ₂ kg ⁻¹ COD _{lactose}	HAc: 0.2 mol mol ⁻¹ lactose HBu: 0.4 mol mol ⁻¹ lactose HPr: 0.4 mol mol ⁻¹ lactose	Ferreira Rosa et al. (2014a)
Cheese whey powder supplemented with medium	Pretreated anaerobic granular sludge	AFBR	30	4.0–4.5 (controlled)	n.a. ^c	6	74 L H ₂ kg ⁻¹ COD _{lactose}	n.a. ^c	Ferreira Rosa et al. (2014b)
Cheese whey + buffalo manure	Pretreated anaerobic sludge	CSTR	55	4.8–5.0 (at steady state)	0.7–2.6 g _{VS} L ⁻¹ d ⁻¹	192–288	131.8 L H ₂ kg ⁻¹ VS	HAc: 4.2 mmol g ⁻¹ VS HBu: 14.1 mmol g ⁻¹ VS HPr: 0.5 mmol g ⁻¹ VS	Ghimire et al. (2017)
CW powder	Acclimated anaerobic sludge	CSTR	30	4.5–7.0 (controlled)	15 g _{COD} L ⁻¹ d ⁻¹	1	n.a. ^c	HAc: 3.5–12 g L ⁻¹ HBu: 2–3 g L ⁻¹ HPr: 2–3 g L ⁻¹	Gouveia et al. (2017)
Ultrafiltered cheese whey	None	CSTR	36	5.5 (controlled)	n.a. ^c	6–12	78–107 L H ₂ kg ⁻¹ COD _{lactose}	n.a. ^c	Montecchio et al. (2018)
CW powder solution	Pretreated anaerobic granular sludge	AFBR	55	4.0–4.5 (controlled)	235.2 g _{lactose} L ⁻¹ d ⁻¹	4	214 L H ₂ kg ⁻¹ COD _{lactose}	HAc: 0.5 g L ⁻¹ HBu: 0.7 g L ⁻¹	Ottaviano et al. (2017)
Cheese whey powder supplemented with sodium bicarbonate	Pre-fermented cheese whey	PBR	30	5.6 (controlled)	22–37 g _{COD} L ⁻¹ d ⁻¹	24	39 L H ₂ kg ⁻¹ COD _{lactose}	HAc: 10 g L ⁻¹ HBu: 2 g L ⁻¹	Perna et al. (2013)
Cheese whey	None	CSTR	35	5.0–6.0 (controlled)	60 g _{COD} L ⁻¹ d ⁻¹	24	48 L H ₂ kg ⁻¹ COD _{consumed}	HAc: 9.2 g L ⁻¹ HBu: 14.5 g L ⁻¹	Venetsaneas et al. (2009)
Dry whey permeate powder	Anaerobic sludge	CSTR	35	Uncontrolled condition	14 g _{COD} L ⁻¹ d ⁻¹	24	52 L H ₂ kg ⁻¹ COD	HAc: 2.1 g L ⁻¹ HPr: 0.1 g L ⁻¹ HBu: 0.8 g L ⁻¹ HCa: 1.2 g L ⁻¹	Yang et al. (2007)

^a AFBR, anaerobic fluidized bed reactor; ASTBR, anaerobic structured-bed reactor; CSTR, continuously stirred tank reactor; PBR, packed bed reactor; SBR, sequence batch reactor; UASB, upflow anaerobic sludge blanket reactor.

^b HAc, acetic acid; HBu, butyric acid; HCa, caproic acid; HPr, propionic acid.

^c Not available.

^d pH adjusted to 5.5 whenever it dropped to 4.5.

alternative to H₂. The production of butyrate and acetate from CW occurs at pH 5–6 (Table 2), whereas propionate production prevails over the butyrate pathway at pH 7–7.5 (Asunis et al., 2019). Generally, a total of 0.5–0.6 g VFA g⁻¹ CW is obtained in the pH range 5–7 (Asunis et al., 2019; Colombo et al., 2016; Duque et al., 2014; Gouveia et al., 2017). CW fermentation at pH < 5 can promote the accumulation of lactic acid (Asunis et al., 2019; Gouveia et al., 2017). Interestingly, in continuous reactors, the type and concentration of VFAs produced at a given pH appear to be the same irrespective of the starting conditions (Gouveia et al., 2017). This suggests that the fermentation pH may be adjusted during continuous operation to target specific metabolic products.

In-line VFA extraction can be implemented during DF of CW to improve process stability and allow a continuous recovery of VFAs (Dessi et al., 2020).

3.2.2. Lactate fermentation

Lactic acid is characterised by an increasing global demand from 0.7 Mt in 2013 to 1.9 Mt in 2020 (grandviewresearch.com), mainly as the building block for polylactic acid (PLA) production, but also for application in food, pharmaceutical and chemical industries. Most commercial lactic acid is currently produced by bacterial fermentation of corn, sugarcane, molasses and other crops. CW has been proposed as an

alternative feedstock to avoid competition with food production.

Lactic acid is mainly produced by bacteria belonging to the genera *Lactobacillus*, *Lactococcus*, *Streptococcus*, *Bacillus*, and *Enterococcus* (Miller et al., 2011; Ryan and Walsh, 2016). Lactose is hydrolysed by enzymes such as β -galactosidase, produced by lactic acid bacteria, into glucose and galactose, and then converted into lactic acid via homolactic fermentation (Fig. 2), resulting in a yield of 4 mol lactate mol⁻¹ lactose. However, ethanol or acetate can be produced along with lactate via heterolactic fermentation (Castillo Martinez et al., 2013; Sikora et al., 2013), halving the lactate yield. The type of fermentation pathway and the specific lactate isomer (L- or D-) produced depend on the lactic acid bacteria involved and the operating conditions, in particular pH (Mazzoli et al., 2014; Miller et al., 2011; Xu et al., 2018).

Since lactic acid bacteria have limited potential to biosynthesize amino-acids, the presence of a nitrogen source is crucial for their growth (Mazzoli et al., 2014; Prazeres et al., 2012). Due to its high protein content, raw CW can be used for lactate production, although enzymatic hydrolysis of lactose might be necessary. Xu et al. (2018) reported a D-lactic acid productivity of 2.4 g L⁻¹ d⁻¹ from hydrolysed CW powder by *Lactobacillus bulgaricus* in non-sterile conditions and without the addition of extra nutrients, which was further enhanced by the addition of 9 g L⁻¹ yeast extract. The yield can be further improved by continuous extraction of the lactic acid produced, since its accumulation inhibits the biomass activity. Taleghani et al. (2018) reported a lactic acid production rate of 6.1 g L⁻¹ h⁻¹ in a fermentative reactor with an integrated membrane extraction system, as opposed to 3.4 g L⁻¹ h⁻¹ obtained in the control reactor without membrane extraction.

3.2.3. Ethanol fermentation

Bioethanol is considered one of the most promising candidates for replacing fossil fuels, and thus its global demand is constantly increasing (marketsandmarkets.com). CW fermentation into ethanol is currently hardly competitive with the established processes that use sugarcane, corn starch or lignocellulosic biomass as raw materials (Guimarães et al., 2010). Solventogenesis from CW has been attempted with yeasts such as *Saccharomyces cerevisiae* (Staniszewski et al., 2007), but low ethanol yields were obtained due to the low lactose conversion and product inhibition. Conversely, the *Kluyveromyces marxianus* yeast was shown to hydrolyse lactose, form a biofilm and tolerate ethanol, and is thus a potential candidate for CW conversion into bioethanol (Joshi et al., 2011; Lane and Morrissey, 2010). Continuous fermentation is potentially superior to batch processes, as it results in higher ethanol production while reducing the fermentation time (Gabardo et al., 2014). Several strategies have been proposed to retain the microorganisms into the bioreactor, including cell immobilization (Dahiya and Vij, 2012), cell recycle (Santos et al., 2016) and membrane retention (Wei et al., 2014). Christensen et al. (2011) obtained continuous ethanol production from CW, with a rate of 2.5–4.5 g L⁻¹ h⁻¹, using a pure culture of *K. marxianus* immobilized in Ca-alginate.

The ethanol yield strictly depends on the operating parameters such as substrate concentration, pH and temperature (Table 3). Using a

continuous FBR with Ca-alginate immobilized-cells, Gabardo et al. (2014) obtained the highest ethanol productivity of 6.0 g L⁻¹ h⁻¹ from CW permeate at a concentration of 150 g L⁻¹ although the highest ethanol yield was obtained at 90 g L⁻¹. Dragone et al. (2011) reported that a lactose concentration of 200 g L⁻¹ and a temperature of 35 °C were optimal for ethanol production (81 g L⁻¹ in 44 h) from CW powder by *K. fragilis*. Using the response surface methodology, Diniz et al. (2014) reported that temperatures between 33.3 and 38.5 °C, pH between 4.7 and 5.7, lactose concentrations between 50 and 108 g L⁻¹ and biomass concentrations between 2.4 and 3.3 (optical density at 600 nm) are optimal for ethanol production from CW by *K. marxianus*, with yields above 90% of the theoretical value.

3.3. Biopolymers

CW fermentation products, mainly VFAs, can be used as building blocks for biopolymer production (Colombo et al., 2016; Duque et al., 2014; Gouveia et al., 2017; Ryan and Walsh, 2016). Biopolymers such as polylactic acid (PLA) and polyhydroxyalkanoates (PHA) are a bio-based, biodegradable alternative to petroleum-based plastics, and their market size is expected to increase from 2.11 Mt in 2018 to 2.63 Mt in 2023 (European Bioplastics, 2018).

3.3.1. PLA

PLA is a versatile biopolymer used in a wide range of industrial sectors, such as food packaging, textile, agriculture, electronics, transportation as well as in the biomedical field. PLA is currently the largest compostable synthetic plastic produced worldwide and its production is projected to increase up to 0.6 Mtons year⁻¹ in 2025 (IEA Bioenergy Task42, 2012). However, the high costs of the building block used for PLA production, mostly lactic acid from microbial corn starch fermentation, hinders full exploitation of its potential. This may be fostered by optimized lactate production from residual organic materials (including CW), as outlined in section 3.2.2.

3.3.2. PHA

PHAs are polyesters produced from organic substrates by various microorganisms, which accumulate them inside the cell for energy storage purposes. PHA production from CW has been reported from microorganisms able to synthesize polymers from lactose, such as *Thermus thermophilus* (Pantazaki et al., 2009), *Pseudomonas hydrogenovora* (Koller et al., 2008), and *Bacillus megaterium* (Das et al., 2018) or engineered *Cupriavidus necator* (Povolo et al., 2010). Although higher PHA accumulation can be attained with pure cultures, mixed microbial communities can produce PHAs from complex and cheaper substrates, such as dairy biowaste, and do not require sterilisation. PHA-producing microorganisms are selected and enriched by alternating short feast (presence of carbon) and long famine (absence of carbon) regimes (Reis et al., 2003). Despite nutrient addition being commonly reported in the literature for selecting high-capacity PHA-storing microbial communities (Oliveira et al., 2018), the high N and P contained in CW might

Table 3
Batch and continuous bioethanol fermentation from CW-based substrates using *Kluyveromyces marxianus*.

Substrate	Substrate concentration (g lactose L ⁻¹)	Reactor	T (°C)	Operating conditions	Ethanol production	Reference
Cheese whey	46.8	Continuous fluidized-bed bioreactor (alginate-immobilized cells)	32	Dilution rate: 0.2 h ⁻¹	4.5 g L ⁻¹ h ⁻¹	Christensen et al. (2011)
Cheese whey powder	150	Batch reactor	35	pH: 4.5	43.7 g L ⁻¹	Das et al. (2016)
Cheese whey permeate	150	Continuous fluidized-bed bioreactor (alginate-immobilized cells)	30	Dilution rate: 0.3 h ⁻¹	6.0 g L ⁻¹ h ⁻¹	Gabardo et al. (2014)
Cheese whey	48	Fed-batch reactor	30	Uncontrolled condition	8.0 g L ⁻¹	Hadiyanto et al. (2014)
Cheese whey	43.6	Batch reactor	28	Uncontrolled condition	17 g L ⁻¹	Zoppellari and Bardi (2013)

reduce, or even eliminate, the need for an external nutrient supply (Colombo et al., 2016).

PHA production from fermented CW by mixed cultures resulted in storage yields of 0.7–0.8 mol PHA mol⁻¹ VFA, with a PHA content of 65–75% (Table 4). The PHA composition (polyhydroxybutyrate, PHB, or polyhydroxyvalerate, PHV) depends on the carboxylic acid present in the CW fermentate: the higher the concentration of acetate and butyrate, the higher the PHB fraction, whereas high concentrations of propionate result in PHV accumulation. PHV fractions up to 40% have been reported from fermented CW (Table 4). Recently, fermented CW has also been used as the substrate for PHA production by phototrophic mixed cultures, using light intensities comparable with those typical of sunny regions, yielding 0.6 g COD_{PHA} g⁻¹ COD_{VFA} and PHA contents of 20–25% (Fradinho et al., 2019).

3.4. Bioelectrochemical systems

Microbial electrochemical technologies (METs) can be implemented to recover the energy contained in CW as electricity in microbial fuel cells (MFCs) (Table 5) or as H₂ in microbial electrolysis cells (MECs) (Table 6). In MFCs, specific microorganisms, namely exoelectrogens, oxidise the organic substrate and transfer the electrons to an anode electrode. Electrons then flow to a cathode electrode through an external circuit, producing electric power, and combine with an electron acceptor, such as oxygen, closing the circuit (Logan et al., 2006). In MECs, the protons resulting from substrate oxidation are the final electron acceptors, producing H₂, if enough energy is provided as input current to drive the reaction (Rago et al., 2016).

Antonopoulou et al. (2010) were the first to test CW, diluted to 0.73 g COD L⁻¹ and amended with nutrients, as the substrate for MFC, yielding a maximum power density of 18.4 mW m⁻² and a coulombic efficiency (CE) of only 1.9%, due to the presence of undesired microorganisms in CW. To address this issue, Stamatelatou et al. (2011) filter-sterilised CW prior to dilution, obtaining power densities up to 40 mW m⁻². The effect of COD concentration (0.35–6.7 g L⁻¹) was investigated by Tremouli et al. (2013), who reported the highest power production (46 mW m⁻²) and CE (11.3%) from diluted CW at 6.7 g COD L⁻¹, with a 95% COD removal efficiency. Ghasemi et al. (2017) compared CW (50 g lactose L⁻¹) and concentrated CW (100 g lactose L⁻¹) as the substrate in a two-chamber MFC, reporting a higher power density from raw CW (288 mW m⁻²) than from concentrated CW (188 mW m⁻²).

Since carboxylic acids are favourable substrates for exoelectrogenic microorganisms, Wenzel et al. (2017) proposed fermented CW as the substrate for a single-chamber MFC, obtaining a dramatically higher

power production (439 mW m⁻²) than a control reactor fed with raw CW (0.34 mW m⁻²). Indeed, exoelectrogenic microorganisms were enriched in the MFC fed with fermented CW, due to the high concentration of VFAs, whereas the high lactose and lactate concentrations of the raw CW resulted in a prevalence of fermentative microorganisms.

Both CW and fermented CW, as well as digestate from AD of CW, have been used as the substrates for H₂ production in MEC (Table 6). Diluted CW (2 g COD L⁻¹), amended with a phosphate buffer solution, was resulted in a production of 0.8 L H₂ L⁻¹ d⁻¹, with energy recoveries up to 71% (Rago et al., 2016). Moreno et al. (2015) combined DF and MEC for two-stage H₂ production from CW, obtaining 0.5 L H₂ L⁻¹ d⁻¹ from filtered, eight-times diluted CW fermentate, supplemented with acetate, in a MEC. However, a rapid decrease in the MEC performance occurred, probably due to the lack of nutrients in the diluted substrate. Rivera et al. (2017) compared raw, fermented and digested CW for H₂ production in a single-chamber MEC. H₂ production yields of 61 and 48 mL H₂ g⁻¹ COD_{removed} were obtained from digested and fermented CW, with a CE of 93 and 32%, respectively, whereas a negligible H₂ production (CE 1%) was obtained from raw CW. However, besides their composition, the different initial organic load (19.9, 1.6 and 4.0 g L⁻¹ COD for raw, fermented and digested CW, respectively) may have affected the results. Fermented CW, rather than raw CW, should thus be used as the substrate for energy recovery in METs. METs can also be seen as a final polishing stage after the AD process. Filtration and dilution should be avoided, since they may result in a lack of nutrients which can hinder the electrogenic activity.

3.5. Integrated processes

A combination of treatment processes to produce an array of valuable products is required for the implementation of a zero-waste-approaching dairy biorefinery (Morais and Bogel-Lukasik, 2013). Combinations of physical, chemical and biological processes (Table 7; Fig. 3) can be implemented. Protein recovery, e.g. by isoelectric or thermocalcic precipitation or nano- or ultrafiltration, may be applied prior to the biological treatment (Bosco et al., 2018; Chen et al., 2016), although negatively affecting the availability of nutrients for the subsequent biological processes.

DF was applied as the first step in most of the studies combining biological treatments for CW valorisation (Table 7), standing as the core of the biorefinery. Several biological downstream processes can be then applied for further valorisation of the DF effluent. Among these, AD is the most applied process on fermented dairy effluent (Table 7). Combination of DF and AD typically leads to high COD removal efficiencies

Table 4
PHA production from CW derived fermentates using mixed microbial communities.

Substrate	Fermentation yield (g COD g ⁻¹ COD)	Fermentation products (PHA precursors) HLa/HAc/HBu/HPr/HVa/HCa/EtOH ^a (% Organic Acid as COD)	Max PHA content (g PHA kg ⁻¹ VSS)	PHA storage yield (g COD _{PHA} g ⁻¹ COD)	Productivity (g PHA L ⁻¹ d ⁻¹)	Polymer composition (% HB:%HV) ^b	Reference
Cheese whey	0.4	58/16/26/0/0/0/0	659 ± 46	0.6 ± 0.0	10.9 ± 0.8	100:0	Colombo et al. (2016)
Sterilised cheese whey	0.6 ± 0.1	6/58/13/19/4/0/0	814 ± 57	0.7 ± 0.1	28.2 ± 2.0	60:40	Colombo et al. (2016)
Cheese whey	0.7 ± 0.2	1/58/22/6/4/0/9	650	0.7 ± 0.1	13.4	81:19	Duque et al. (2014)
Cheese whey	0.6–0.7	16/45/23/14/6/0/5	300	0.6	n.a. ^c	88:12	Fradinho et al. (2019)
Sweet cheese whey powder	0.64 ± 0.05	0/46/44/4/5/0/0	430	0.85 ± 0.12	0.20	89:11	Oliveira et al. (2018)
Filtered whey permeate	0.5	0/44/50/2/1/3/0	530–630	0.41–0.63	n.a. ^c	85:15	Valentino et al. (2015)

^a HLa, Lactic acid; HAc, Acetic acid; HBu, Butyric acid; HPr, Propionic acid; HVa, Valeric acid; HCa, Caproic acid; EtOH, Ethanol.

^b HB, hydroxybutyrate; HV, hydroxyvalerate.

^c Not available.

Table 5
Electricity production from dairy wastewater or CW-based substrates in MFCs.

Substrate	Inoculum	Reactor characteristics ^a	T (°C)	HRT (h)	Maximum power production	CE (%)	Reference
Cheese whey	Anaerobic sludge	H-type (310 mL) Anode: Teflon treated carbon filter paper Cathode: Carbon cloth with Pt Membrane: PEM	35.5	Batch	18.4 mW m ⁻²	1.9	Antonopoulou et al. (2010)
Cheese whey powder	Anaerobic sludge	Single chamber (125 mL) Anode: Carbon cloth Cathode: Carbon cloth with GDL Membrane: None	35	Batch	n.a. ^b	0.8–2.0	Colombo et al. (2017)
Synthetic dairy wastewater	Municipal wastewater	Dual chamber (480 mL) Anode: Untreated carbon paper Cathode: Untreated carbon paper Membrane: PEM	22	8.4	90 mW m ⁻²	10.5 ± 10	Faria et al. (2017)
Whey	Anaerobic sludge	Cube-shaped dual chamber (420 mL) Anode: Carbon paper Cathode: Carbon paper with Pt Membrane: PEM	30	Batch	188.8 mW m ⁻²	26	Ghasemi et al. (2017)
Concentrated whey	Anaerobic sludge	Cube-shaped dual chamber (420 mL) Anode: Carbon paper Cathode: Carbon paper with Pt Membrane: PEM	30	Batch	288.1 mW m ⁻²	15	Ghasemi et al. (2017)
Cheese whey (diluted 10 times)	Digested sludge	Tubular dual chamber (500 mL) Anode: Carbon fibre brush Cathode: Carbon cloth and activated carbon powder Membrane: PEM	21	Batch	0.4 W m ⁻³	3.9 ± 1.7 (based on total COD)	Kelly and He (2014)
Cheese whey	MFC enriched community	Single chamber (28 mL) Anode: Graphite fibre brush Cathode: Graphite fibre cloth with PTFE and Pt Membrane: None	n.a. ^b	Batch	3.46 mW m ⁻²	49 ± 8	Rago et al. (2016)
Cheese whey	Anaerobic sludge	Dual chamber (310 mL) Anode: Carbon paper Cathode: Carbon cloth Membrane: PEM	30	Batch	46 mW m ⁻²	5.5–11.3	Tremouli et al. (2013)
Dairy wastewater	Anaerobic mixed consortia	Single chamber (550 mL) Anode: Graphite plate Cathode: Graphite plate Membrane: PEM	29	Batch	6.71 mW m ⁻²	4.3–14.2	Venkata Mohan et al. (2010)
Cheese whey	Planktonic MFC community	Single chamber air cathode (25 mL) Anode: Graphite felt Cathode: Carbon cloth with Nafion and Pt Membrane: PEM	30	Batch	0.34 mW m ⁻²	14	Wenzel et al. (2017)
Fermented cheese whey	Planktonic MFC community	Single chamber air cathode (25 mL) Anode: Graphite felt Cathode: Carbon cloth with Nafion and Pt Membrane: PEM	30	Batch	439 mW m ⁻²	24	Wenzel et al. (2017)

^a GDL, gas diffusion layer; PEM, proton exchange membrane; PTFE, polytetrafluoroethylene.

^b Not available.

(>80%), due to the final conversion of VFAs to methane. However, considering the higher pH and HRT required for AD than for DF, pH buffering and high reactor volumes are commonly required for the AD step. Furthermore, the high concentrations (up to 20–30 g L⁻¹) of VFAs produced in DF, as well as the low buffering capacity of CW, may inhibit the AD process (Bertin et al., 2013).

The DF effluent can also be used for further H₂ production, although an external source of energy e.g. in the form of light (photofermentation) or electricity (MEC) is required to overcome the thermodynamic constraints. Rai et al. (2012) combined dark and photofermentation of diluted CW (10 g L⁻¹ lactose) using immobilized pure cultures, obtaining a yield of 199 L H₂ kg⁻¹ COD, although the COD removal efficiency was low (36%). The application of METs to fermented CW could be favoured by the low ohmic resistance associated to the typically high

salinity of CW. A remarkably high yield of over 800 L H₂ g⁻¹ COD was obtained from deproteinized ricotta CW, diluted to 3 g COD L⁻¹, by combining DF and MEC, with 63% COD removal efficiency (Marone et al., 2017).

Since DF effluents are rich in VFAs, DF can also be coupled to PHA production. Colombo et al. (2019) proposed an integrated two-stage bioprocess aimed at simultaneously recovery of H₂ (2.4–5.1 L H₂ L⁻¹ d⁻¹) and PHB (274–268 g kg⁻¹ COD_{feed}) from deproteinized CW. In order to produce PHAs at high concentrations, a VFA extraction and concentration step, e.g. via electro dialysis, can be included in the integrated process. Domingos et al. (2018) applied electro dialysis to fermented CW obtaining a concentrated VFA stream (up to 63 g L⁻¹ from the original concentration of 13 g L⁻¹), from which a PHA yield of 0.60 g PHA g⁻¹ VFA was obtained, comparable to the yields reported from

Table 6
Hydrogen production from dairy wastewater or CW-based substrates in MECs.

Substrate	Inoculum	Reactor characteristics ^a	T (°C)	HRT (h)	Maximum hydrogen production	CE (%)	Reference
Ricotta cheese production wastewater (scotta)	Anaerobic sediments	Cylindrical two-chamber (400 mL) Anode: Carbon felt Cathode: Pt-Ir (90%;10%) mesh Membrane: AEM	37	Batch	0.023 L H ₂ L ⁻¹ d ⁻¹	75 (estimated)	Marone et al. (2017)
Fermented cheese whey (diluted)	Planktonic MEC community	Continuous flow membrane-less (50 mL) Anode: Carbon felt Cathode: GDL with Ni nanoparticles Membrane: None	25	10	MEC failure	n.a. ^b	Moreno et al. (2015)
Fermented cheese whey and acetate	Planktonic MEC community	Continuous flow membrane-less (50 mL) Anode: Carbon felt Cathode: GDL with Ni nanoparticles Membrane: None	25	10	MEC failure	n.a. ^b	Moreno et al. (2015)
Fermented cheese whey and salts	Planktonic MEC community	Continuous flow membrane-less (50 mL) Anode: Carbon felt Cathode: GDL with Ni nanoparticles Membrane: None	25	10	0.5 L H ₂ L ⁻¹ d ⁻¹	n.a. ^b	Moreno et al. (2015)
Fermented cheese whey, salts and acetate	Planktonic MEC community	Continuous flow membrane-less (50 mL) Anode: Carbon felt Cathode: GDL with Ni nanoparticles Membrane: None	25	10	0.5 L H ₂ L ⁻¹ d ⁻¹	n.a. ^b	Moreno et al. (2015)
Cheese whey	MFC enriched community	Single chamber (28 mL) Anode: Graphite fibre brush Cathode: Graphite fibre cloth with PTFE and Pt Membrane: None	n.a. ^b	Batch	0.8 L H ₂ L ⁻¹ d ⁻¹	120 ^c	Rago et al. (2016)
Cheese whey	Anaerobic sludge	Single chamber (300 mL) Anode: Graphite felt Cathode: Stainless steel mesh Membrane: None	32	Batch	MEC failure	1	Rivera et al. (2017)
Cheese whey digestate	Anaerobic sludge	Single chamber (300 mL) Anode: Graphite felt Cathode: Stainless steel mesh Membrane: None	32	Batch	0.16 L H ₂ L ⁻¹ d ⁻¹	31.8	Rivera et al. (2017)
Fermented cheese whey	Anaerobic sludge	Single chamber (300 mL) Anode: Graphite felt Cathode: Stainless steel mesh Membrane: None	32	Batch	0.06 L H ₂ L ⁻¹ d ⁻¹	92.7	Rivera et al. (2017)

^a AEM, anion exchange membrane; GDL, gas diffusion layer; PEM, proton exchange membrane; PTFE, polytetrafluoroethylene.

^b Not available.

^c Due to H₂ recycling by homoacetogenic bacteria.

VFA-containing synthetic solutions.

4. Full-scale applications

Despite the potential for CW valorisation, the implementation of integrated, multiple treatment schemes is still limited. Existing full-scale plants are mostly AD plants producing biogas to cover part of the dairy industry energy needs. To be economically viable, more complex processing sequences would require a plant size that often exceeds the potential of small-to medium-size dairy industries.

Among the full-scale applications, the company Valbio provides AD systems to dairy industries through its patented technology Valbio Methcore®, based on UASB technology. Valbio has commissioned more than 10 full-scale plants for dairy companies mostly located in France, Canada and Bulgaria, treating 0.3–10.5 million L of CW year⁻¹ with an energy production of 0.3–3.5 MWh year⁻¹ and COD removal efficiency higher than 90% (Valbio.com). Dairygold Co-Operative Society Limited recently installed the world's largest above ground anaerobic digester (ADI/BVF®, Evoqua) in Ireland. The Dairygold low-rate anaerobic

digester was designed to treat 5500 m³ d⁻¹ wastewater containing powdered milk, cheese waste and CW (Evoqua.com), meeting the discharge limits and contributing to satisfy the dairy industry energy needs. First Milk's Lake District creamery (Cumbria, UK) was the first dairy industry to feed upgraded biomethane generated from cheese process residues to the national gas grid in 2016 (Clearfleau.com). The CSTR was designed to treat 1650 m³ d⁻¹ of dairy wastewater and whey producing 5.4 MWh of bioenergy (Clearfleau.com).

Industrial-scale bioethanol plants are in operation in Ireland, New Zealand, USA, Denmark and Germany. The Carbery Group factory, the largest single cheese-producing facility in Ireland, started the operation of an industrial-scale whey-to-ethanol plant in 1978 (Carbery, 2018). In addition to cheese, the company produces high-quality ethanol, accounting for 50% of Ireland's industrial ethanol needs for beverage, pharmaceutical and food industries (Carbery, 2018). Since 2005, the company has also been supplying ethanol to petrol companies in Ireland. Anchor Ethanol operates three whey-to-ethanol plants in New Zealand, using deproteinated whey, concentrated from 4 to 8% lactose and fermented for 24 h by *Kluyveromyces* sp., as feedstock attaining an ethanol

Table 7

Combination of at least two chemical or biological processes for energy or resource recovery from CW-based substrates.

Process ^a	Substrate	Inoculum	Temperature (°C)	HRT	Output	COD removal (%)	Reference
Dark fermentation (CSTR, 3 L) + methanogenesis (PABR, 15 L)	Cheese whey (61 g COD L ⁻¹)	Indigenous microflora	35 (both processes)	1 day (dark fermentation); 4.4 days (methanogenesis)	Hydrogen: 41 L kg ⁻¹ COD; Methane: 383 L kg ⁻¹ COD	94	Antonopoulou et al. (2008)
Acidogenesis + methanogenesis (two-stage concentric reactor, 190 mL for acidogenic reactor, 790 mL for methanogenic reactor)	Cheese whey + cattle manure (1:1; 35.2 g COD L ⁻¹)	Methanogenic sludge (both processes)	35 (both processes)	5 days (acidogenesis); 20 days (methanogenesis)	Methane: 258 kg ⁻¹ VS	83	Bertin et al. (2013)
Thermocalcic precipitation, ultrafiltration + PHA production	Cheese whey (50 g COD L ⁻¹)	Dairy plant activated sludge enriched on acetate (PHA production)	45–55 (thermocalcic precipitation) Not reported for PHA production	24–48 h (PHA production)	Proteins: 80 g L ⁻¹ , PHA: 0.75–0.90 g L ⁻¹	n.a. ^b	Bosco et al. (2018)
Isoelectric precipitation, nanofiltration + dark fermentation (UASB, 7.4 L)	Milk powder (3.0 g COD L ⁻¹)	Sewage sludge	25 (precipitation and nanofiltration) 37 (dark fermentation)	12 h (dark fermentation)	Proteins: 192 g kg ⁻¹ COD; Reusable water; Hydrogen (not quantified); VFAs: 2.2 g L ⁻¹	n.a. ^b	Chen et al. (2016)
Enzymatic hydrolysis step with β-galactosidase + Dark fermentation (CSTR, 4 L) + PHA production (SBR, 1 L; Fed-batch assay, 0.5 L)	Secondary cheese whey and concentrated cheese whey permeate (OLR: 8, 11, 15 g sugars L ⁻¹ d ⁻¹ for dark fermentation; 1.5 mg _{COD} L ⁻¹ d ⁻¹ for PHA production)	Thermally pretreated anaerobic digested sludge (dark fermentation); Activated sludge (PHA production)	55 (dark fermentation) 25 (PHA production)	2 days (dark fermentation) 1 day (PHA production)	Hydrogen: 163–233 L kg ⁻¹ COD; PHA: 268–274 g kg ⁻¹ COD	n.a. ^b	Colombo et al. (2019)
Dark fermentation (CSTR, 3 L) + methanogenesis (UASB, 1 L)	Cheese whey powder (45.5 g COD L ⁻¹)	Anaerobic granular sludge	37 (dark fermentation) 25–30 (methanogenesis)	6 h (both processes)	Hydrogen: 137 L kg ⁻¹ COD; Methane: 250 L kg ⁻¹ COD	92	Cota-Navarro et al. (2011)
Dark fermentation (anaerobic column biofilm packed reactor, 1 L) + electro dialysis + PHA production (3 L)	Cheese whey powder (28 g COD L ⁻¹)	Acidogenic sludge (dark fermentation); <i>Cupravidus necator</i> (PHA production)	37 (dark fermentation) 30 (PHA production)	6 h (dark fermentation); 52 h (PHA production, batch mode)	VFAs: 13 g L ⁻¹ (60 g L ⁻¹ after electro dialysis); PHA: 500 g kg ⁻¹ COD	n.a. ^b	Domingos et al. (2018)
Dark fermentation (batch, 2 L) + methanogenesis (batch, 2 L)	Dairy wastewater (11.2 g COD L ⁻¹)	<i>Enterobacter aerogens</i> (dark fermentation); Digested cow dung slurry (methanogenesis)	30 (dark fermentation) 35 (methanogenesis)	13 h (dark fermentation); 7 days (methanogenesis)	Hydrogen: 105 L kg ⁻¹ COD; Methane: 190 L kg ⁻¹ COD	64	Kothari et al. (2017)
Dark fermentation (batch, 500 mL) + biocatalyzed electrolysis (MEC, 400 mL)	Deproteinized ricotta cheese whey (57.8 g COD L ⁻¹) diluted to 3 g _{COD} L ⁻¹	Anaerobic digested sludge (dark fermentation)	37 (both processes)	48 h (dark fermentation); 14 days (MEC)	Hydrogen: 95.1 + 714.7 L kg ⁻¹ COD; Electric current: 7.46 A m ⁻²	63	Marone et al. (2017)
Dark fermentation (batch, 250 mL) + biocatalyzed electrolysis (MEC, 50 mL)	Cheese whey (fermentation; 122 g COD L ⁻¹); Fermented cheese whey (MEC; diluted 8 times and amended with acetate and nutrients)	Digested sludge (dark fermentation); Domestic wastewater-fed MEC effluent (MEC)	35 (dark fermentation) 25 (MEC)	Not reported for dark fermentation; 10 h (MEC)	Hydrogen: 94.2 L kg ⁻¹ VS; Electric current: 10 mA	82	Moreno et al. (2015)
Dark fermentation (batch, 100 mL) + photofermentation (batch, 100 mL)	Diluted cheese whey (10 g lactose L ⁻¹)	<i>Enterobacter aerogens</i> (dark fermentation); <i>Rhodospseudomonas</i> (photofermentation)	30 (dark fermentation) 34 (photofermentation)	84 h (both processes)	Hydrogen: 199 L kg ⁻¹ COD	36	Rai et al. (2012)
Dark fermentation (CSTR, 0.5 L) + methanogenesis (CSTR, 3 L)	Cheese whey (60.5 g COD L ⁻¹)	Indigenous microflora	35 (both processes)	1 day (dark fermentation); 20 days (methanogenesis)	Hydrogen: 48 L kg ⁻¹ COD; Methane: 31 L kg ⁻¹ COD	95	Venetsaneas et al. (2009)

^a CSTR, continuously stirred tank reactor; MEC, microbial electrolysis cell; PABR, periodic anaerobic baffled reactor; UASB, upflow anaerobic sludge blanket.^b Not available.

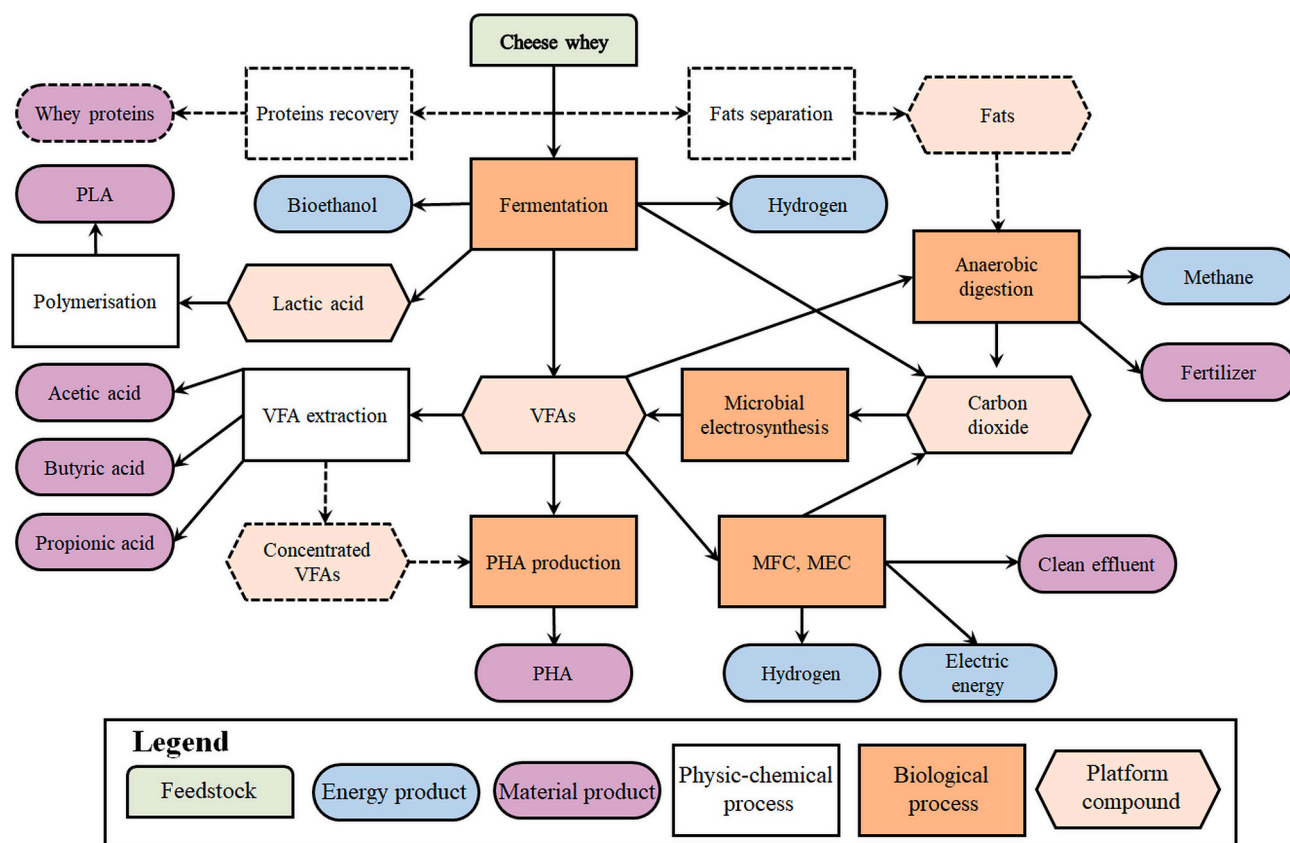


Fig. 3. Integrated treatment processes for cheese whey valorisation according to the circular economy principle. Symbols and colours are represented according to Cherubini et al. (2009). The dashed parts represent optional processes, not essential for the following treatment steps. (For interpretation of the references to colour in this figure legend, the reader is referred to the Web version of this article.)

titre of 4%, successively concentrated to various ethanol grades by distillation (Guimarães et al., 2010).

5. Future perspectives

CW is an abundant substrate, easily available and at low cost, but at the same time needs proper management. Many of the processes applicable for biotechnological valorisation of CW are currently at a medium/high TRL and some integration schemes between these processes are promising. However, some critical aspects need further investigation in order to make the application of the biorefinery concept to the dairy supply chain fully feasible.

CW displays very variable characteristics depending on both the livestock originating the milk, and the geographical context. The milk and the resulting CW production is characterised by a strong seasonal variability in terms of quantity and composition, that follows the lactation period. The seasonal variation could be managed by freezing CW during peak production, and subsequently thawing on demand. A better solution could be based on assessing the availability of CW in the area under concern, and promoting consortia to ensure an even CW supply throughout the year (Ubando et al., 2020).

The optimal combination of the biotechnological processes strictly depends on CW availability and characteristics, as well as legislation and market demand. Pre-treatment of CW might simplify downstream valorisation. For example, a protein extraction stage, already well developed at the industrial scale (TRL 9), could be integrated into the process chain, fostered by the relatively high value of whey proteins (6–22 € kg⁻¹) (Table 8), but addition of nutrients may be required for the subsequent biological treatment stages. Similarly, the need for post-treatments aimed at removing undesired impurities or extracting the compounds of interest must be carefully evaluated, being an important

cost item in the entire process scheme.

Among the soluble products of CW fermentation, acetic acid and ethanol are currently characterised by low economic values, but relatively big market sizes, whereas butyric and lactic acid have smaller, but rapidly growing (15–19% compound annual growth rate, CAGR) markets (Table 8). It should be noted, however, that obtaining individual marketable products from the mixture of carboxylic acids typically obtained from CW fermentate would require highly selective and efficient extraction systems, currently available at TRL 2–3.

As an alternative, the carboxylic acids mixture can be used as a feedstock for biopolymer production, and in particular for PHA production. The technology for PHA production from biowaste is still in the development stage (TRL 3–5). However, the high value of PHAs (2.8–3.2 € kg⁻¹), and the rapidly increasing bioplastics market (16.5% CAGR) could make biological PHA production profitable in the near future. Specific tailored solutions can be investigated within the same dairy industry supply chain, e.g. using the PHA produced from CW as a sustainable packaging for dairy products.

METs are still under development (TRL 3–4). Due to the high cost, and the typically low power density and H₂ yield achievable through MFC and MEC, respectively, their use for treatment of undiluted CW fermentate, characterised by high carboxylic acid concentration, does not appear profitable. In particular, MFCs can hardly compete with technologies such as solar energy and wind power for electricity production at a large scale, unless many cells are stacked together (Gajda et al., 2018). However, due to the high COD removal efficiencies, both MFCs and MECs can be seen as a polishing stage prior to effluent disposal. Among the bioelectrochemical systems, microbial electrosynthesis (MES) can be a key player in limiting the carbon emissions by recycling the CO₂ produced by other bioprocesses, and from the dairy industry itself, and converting it to carboxylic acids for downstream

Table 8

Treatment processes applicable in a biorefinery concept (Fig. 3), including their technological readiness level (TRL), and market value and market size (converted into € from original data in USD) of the main products obtained.

Process	TRL ^a	Products	Indicative price (€ kg ⁻¹)	Global market size (€)	Global market forecast (€)	CAGR ^b (%)	References
Protein recovery	9	Functional proteins	6.4–22.0	3.9 × 10 ⁹ (2017)	5.3 × 10 ⁹ (2022)	6.3	Marketsandmarkets.com
Dark fermentation	4–5	Hydrogen	0.9–7.3	124.4 × 10 ⁹ (2018)	183.4 × 10 ⁹ (2025)	8.0	Marketsandmarkets.com
VFA extraction from fermentation broth	2–3	Acetic acid	0.4–0.7	8.8 × 10 ⁹ (2015)	12.2 × 10 ⁹ (2022)	4.8	Grandviewresearch.com
		Butyric acid	1.4–1.6	114.8 × 10 ⁶ (2014)	218.1 × 10 ⁶ (2020)	15.1	Marketsandmarkets.com
		Propionic acid	1.8–2.3	1.2 × 10 ⁹ (2014)	1.4 × 10 ⁹ (2020)	2.8	Grandviewresearch.com
Lactic acid fermentation	8	Lactic acid	0.9	1.9 × 10 ⁹ (2015)	3.5 × 10 ⁹ (2020)	18.6	Marketsandmarkets.com
Alcohol fermentation	9	Bioethanol	0.6–1.4	48.5 × 10 ⁹ (2016)	63.5 × 10 ⁹ (2022)	5.3	Marketsandmarkets.com
Polymerisation	9	PLA	2.0	5.5 × 10 ⁹ (2017)	13.8 × 10 ⁹ (2023)	16.5	Marketsandmarkets.com
Biological biopolymer production	3–5	PHA	2.8–3.2				
Microbial fuel cell	3–4	Renewable electric power	48.9 ^c	5 × 10 ³ TWh (2018)	7 × 10 ³ TWh (2023)	13.1	International Energy Agency, IEA
Microbial electrolysis cell	3–4	Hydrogen	0.9–7.3	124.7 × 10 ⁹ (2018)	183.3 × 10 ⁹ (2025)	5.7	Marketsandmarkets.com
Anaerobic digestion	9	Biomethane	0.4–0.7	1.4 × 10 ⁹ (2017)	2.4 × 10 ⁹ (2025)	7.1	Transparencymarketresearch.com
		Fertilizer	0–6 ^d	5.6 × 10 ⁹ (2019) ^e	8.5 × 10 ⁹ (2024)	6.8	Globenewswire.com
Microbial electrosynthesis	3–4	Acetic acid	0.4–0.7	8.8 × 10 ⁹ (2015)	12.2 × 10 ⁹ (2022)	4.8	Grandviewresearch.com
		Butyric acid	1.4–1.6	114.8 × 10 ⁶ (2014)	218.1 × 10 ⁶ (2020)	15.1	Marketsandmarkets.com
		Caproic acid	1.5	9.2 × 10 ⁶ (2018)	11.5 × 10 ⁶ (2024)	3.2	Marketwatch.com

^a Technology readiness level.

^b Compound annual growth rate.

^c €/MWh, average price in EU.

^d €/m³; data from compost global market.

applications (Batlle-Vilanova et al., 2016; Vassilev et al., 2018). This would close the loop in the carbon recovery chain towards a zero-waste-approaching biorefinery.

6. Conclusions

Cheese whey is an outstanding resource for production of green energy and platform chemical compounds, but currently its potential is not fully exploited. In this review, the most promising biotechnologies for cheese whey valorisation were compared, and the strong and weak points of each one were critically analysed. Due to its simple and efficient application on raw CW, the current and potential huge market size of its products (H₂ and VFAs), and the more and more stringent regulations on carbon emissions, fermentation is likely to gradually replace anaerobic digestion as the core of dairy biorefinery. H₂ is indeed a key player towards the decarbonisation of the energy production system, whereas VFAs have several industrial applications, and may also be regarded as precursors for bioplastic production, the market size of which is expected to increase in response to the policies to reduce the use of traditional plastics. Due to the high organic load of CW, inhibitory for electrogenic microorganisms, MFC and MEC can only find application as the final polishing stage of the dairy biorefinery. MES is a promising technology to recycle the CO₂ generated in the other biological treatment processes and in the energy production plants, providing heat and electricity to the dairy industry, such as boilers and co-generation heat and power (CHP) plants, closing the carbon loop.

Web references

<http://www.valbio.com/en/VALBIO-Projects/VALBIO-projects/>

id_1024

<https://clearfleau.com/portfolio/lake-district-biogas-green-gas-from-cheese-residues/>

<https://www.evoqua.com/en/brands/adi-systems/Pages/bvf-worlds-largest-above-ground-digester-installed-in-ireland.aspx>

<https://www.globenewswire.com/news-release/2019/10/08/1926776/0/en/The-global-compost-market-is-expected-to-reach-an-estimated-9-2-billion-by-2024-with-a-CAGR-of-6-8-from-2019-to-2024.html>

<https://www.grandviewresearch.com/industry-analysis/acetic-acid-market>

<https://www.grandviewresearch.com/industry-analysis/lactic-acid-and-poly-lactic-acid-market>

<https://www.grandviewresearch.com/industry-analysis/propionic-acid-market>

<https://www.iea.org/reports/world-energy-outlook-2019/electricity#abstract>

<https://www.marketsandmarkets.com/Market-Reports/bioethanol-market-131222570.html>

<https://www.marketsandmarkets.com/Market-Reports/biopolymers-bioplastics-market-88795240.html>

<https://www.marketsandmarkets.com/Market-Reports/butyric-acid-market-76962011.html>

<https://www.marketsandmarkets.com/Market-Reports/dairy-ingredients-market-974.html>

<https://www.marketsandmarkets.com/Market-Reports/hydrogen-generation-market-494.html>

<https://www.marketsandmarkets.com/Market-Reports/poly-lactic-acid-387.html>

<https://www.marketwatch.com/press-release/at-29-cagr-hexano>

ic-acid-market-size-poised-to-touch-usd-45-million-by-2024-2020-01-14

<https://www.transparencymarketresearch.com/bio-methane-market-2018-2026.html>

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgements

This work was supported by the Science Foundation of Ireland (SFI) Research Professorship Programme on *Innovative Energy Technologies for Bioenergy, Biofuels and a Sustainable Irish Bioeconomy* (IETS BIO³, award 15/RP/2763). It was conducted on the framework of the “Waste Biorefinery” task group of the International Waste Working Group (IWWG). Fabiano Asunis and Marco Isipato gratefully acknowledges Sardinian Regional Government for the financial support of their PhD scholarship (P.O.R. Sardegna F.S.E. - Operational Programme of the Autonomous Region of Sardinia, European Social Fund 2014–2020 - Axis III Education and training, Thematic goal 10, Investment Priority 10ii, Specific goal 10.5).

References

- Ahmad, T., Aadil, R.M., Ahmed, H., ur Rahman, U., Soares, B.C.V., Souza, S.L.Q., Pimentel, T.C., Scudino, H., Guimaraes, J.T., Esmerino, E.A., Freitas, M.Q., Almada, R.B., Vendramel, S.M.R., Silva, M.C., Cruz, A.G., 2019. Treatment and utilization of dairy industrial waste: a review. *Trends Food Sci. Technol.* 88, 361–372. <https://doi.org/10.1016/j.tifs.2019.04.003>.
- Akhlaghi, M., Boni, M.R., De Gioannis, G., Muntoni, A., Poletti, A., Pomi, R., Rossi, A., Spiga, D., 2017. A parametric response surface study of fermentative hydrogen production from cheese whey. *Bioresour. Technol.* 244, 473–483. <https://doi.org/10.1016/j.biortech.2017.07.158>.
- Akhlaghi, M., Boni, M.R., Poletti, A., Pomi, R., Rossi, A., De Gioannis, G., Muntoni, A., Spiga, D., 2019. Fermentative H₂ production from food waste: parametric analysis of factor effects. *Bioresour. Technol.* 276, 349–360. <https://doi.org/10.1016/j.biortech.2019.01.012>.
- Antonopoulou, G., Stamatelou, K., Venetsaneas, N., Kornaros, M., Lyberatos, G., 2008. Biohydrogen and methane production from cheese whey in a two-stage anaerobic process. *Ind. Eng. Chem. Res.* 47, 5227–5233. <https://doi.org/10.1021/ie071622x>.
- Antonopoulou, G., Stamatelou, K., Bebelis, S., Lyberatos, G., 2010. Electricity generation from synthetic substrates and cheese whey using a two chamber microbial fuel cell. *Biochem. Eng. J.* 50, 10–15. <https://doi.org/10.1016/j.bej.2010.02.008>.
- Asunis, F., De Gioannis, G., Isipato, M., Muntoni, A., Poletti, A., Pomi, R., Rossi, A., Spiga, D., 2019. Control of fermentation duration and pH to orient biochemicals and biofuels production from cheese whey. *Bioresour. Technol.* 289, 121722. <https://doi.org/10.1016/j.biortech.2019.121722>.
- Azbar, N., Dokgöz, F.T., Keskin, T., Eltem, R., Korkmaz, K.S., Gezgin, Y., Akbal, Z., Öncel, S., Dalay, M.C., Gönen, Ç., Tutuk, F., 2009a. Comparative evaluation of biohydrogen production from cheese whey wastewater under thermophilic and mesophilic anaerobic conditions. *Int. J. Green Energy* 6, 192–200. <https://doi.org/10.1080/15435070902785027>.
- Azbar, N., Tuba Çetinkaya Dokgöz, F., Keskin, T., Korkmaz, K.S., Syed, H.M., 2009b. Continuous fermentative hydrogen production from cheese whey wastewater under thermophilic anaerobic conditions. *Int. J. Hydrogen Energy* 34, 7441–7447. <https://doi.org/10.1016/j.ijhydene.2009.04.032>.
- Battle-Vilanova, P., Puig, S., Gonzalez-Olmos, R., Balaguer, M.D., Colprim, J., 2016. Continuous acetate production through microbial electrosynthesis from CO₂ with microbial mixed culture. *J. Chem. Technol. Biotechnol.* 91, 921–927. <https://doi.org/10.1002/jctb.4657>.
- Bertin, L., Grilli, S., Spagni, A., Fava, F., 2013. Innovative two-stage anaerobic process for effective codigestion of cheese whey and cattle manure. *Bioresour. Technol.* 128, 779–783. <https://doi.org/10.1016/j.biortech.2012.10.118>.
- Blanco, V.M.C., Oliveira, G.H.D., Zaiat, M., 2019. Dark fermentative biohydrogen production from synthetic cheese whey in an anaerobic structured-bed reactor: performance evaluation and kinetic modeling. *Renew. Energy* 139, 1310–1319. <https://doi.org/10.1016/j.renene.2019.03.029>.
- Bosco, F., Carletto, R.A., Marmo, L., 2018. An integrated cheese whey valorization process. *Chem. Eng. Trans.* 64, 379–384. <https://doi.org/10.3303/CET1864064>.
- Carbery, 2018. A New Horizon, Our Journey of Sustainable Progress. <https://www.carbery.com/wp-content/uploads/Carbery-Sustainability-Report-2018.pdf>.
- Carriero, C., Di Pinto, A.C., Rozzi, A., Santori, M., 1993. Anaerobic co-digestion of sewage sludge and concentrated soluble wastewaters. *Water Sci. Technol.* 28, 187–197. <https://doi.org/10.2166/wst.1993.0102>.
- Carvalho, F., Prazeres, A.R., Rivas, J., 2013. Cheese whey wastewater: characterization and treatment. *Sci. Total Environ.* 445–446, 385–396. <https://doi.org/10.1016/j.scitotenv.2012.12.038>.
- Castelló, E., García y Santos, C., Iglesias, T., Paolino, G., Wenzel, J., Borzacconi, L., Etchebehere, C., 2009. Feasibility of biohydrogen production from cheese whey using a UASB reactor: links between microbial community and reactor performance. *Int. J. Hydrogen Energy* 34, 5674–5682. <https://doi.org/10.1016/j.ijhydene.2009.05.060>.
- Castelló, E., Braga, L., Fuentes, L., Etchebehere, C., 2018. Possible causes for the instability in the H₂ production from cheese whey in a CSTR. *Int. J. Hydrogen Energy* 43, 2654–2665. <https://doi.org/10.1016/j.ijhydene.2017.12.104>.
- Castillo Martínez, F.A., Balciunas, E.M., Salgado, J.M., Domínguez González, J.M., Converti, A., Oliveira, R.P. de S., 2013. Lactic acid properties, applications and production: a review. *Trends Food Sci. Technol.* 30, 70–83. <https://doi.org/10.1016/j.tifs.2012.11.007>.
- Chen, Z., Luo, J., Chen, X., Hang, X., Shen, F., Wan, Y., 2016. Fully recycling dairy wastewater by an integrated isoelectric precipitation – nanofiltration – anaerobic fermentation process. *Chem. Eng. J.* 283, 476–485. <https://doi.org/10.1016/j.cej.2015.07.086>.
- Cherubini, F., 2010. The biorefinery concept: using biomass instead of oil for producing energy and chemicals. *Energy Convers. Manag.* 51, 1412–1421. <https://doi.org/10.1016/j.enconman.2010.01.015>.
- Cherubini, F., Jungmeier, G., Wellisch, M., Willke, T., Skiadis, I., Van Ree, R., De Jong, E., 2009. Toward a common classification approach for biorefinery systems. *Biofuels, Bioprod. Biorefining* 3, 534–546. <https://doi.org/10.1002/bbb>.
- Christensen, A.D., Kádár, Z., Oleskovicz-Popiel, P., Thomsen, M.H., 2011. Production of bioethanol from organic whey using *Kluyveromyces marxianus*. *J. Ind. Microbiol. Biotechnol.* 38, 283–289. <https://doi.org/10.1007/s10295-010-0771-0>.
- Colombo, B., Sciarria, T.P., Reis, M., Scaglia, B., Adani, F., 2016. Polyhydroxyalkanoates (PHAs) production from fermented cheese whey by using a mixed microbial culture. *Bioresour. Technol.* 218, 692–699. <https://doi.org/10.1016/j.biortech.2016.07.024>.
- Colombo, A., Schievano, A., Trasatti, S.P., Morrone, R., D'Antona, N., Cristiani, P., 2017. Signal trends of microbial fuel cells fed with different food-industry residues. *Int. J. Hydrogen Energy* 42, 1841–1852. <https://doi.org/10.1016/j.ijhydene.2016.09.069>.
- Colombo, B., Villegas Calvo, M., Sciarria, T.P., Scaglia, B., Savio Kizito, S., D'Imporzano, G., Adani, F., 2019. Biohydrogen and polyhydroxyalkanoates (PHA) as products of a two-steps bioprocess from deproteinized dairy wastes. *Waste Manag.* 95, 22–31. <https://doi.org/10.1016/j.wasman.2019.05.052>.
- Comino, E., Riggio, V.A., Rosso, M., 2012. Biogas production by anaerobic co-digestion of cattle slurry and cheese whey. *Bioresour. Technol.* 114, 46–53. <https://doi.org/10.1016/j.biortech.2012.02.090>.
- Cota-Navarro, C.B., Carrillo-Reyes, J., Davila-Vazquez, G., Alatríste-Mondragón, F., Razo-Flores, E., 2011. Continuous hydrogen and methane production in a two-stage cheese whey fermentation system. *Water Sci. Technol.* 64, 367–374. <https://doi.org/10.2166/wst.2011.631>.
- Dahiya, M., Vij, S., 2012. Comparative analysis of bioethanol production from whey by different strains of immobilized thermotolerant yeast. *Int. J. Sci. Res. Publ.* 2, 3.
- Das, B., Sarkar, S., Maiti, S., Bhattacharjee, S., 2016. Studies on production of ethanol from cheese whey using *Kluyveromyces marxianus*. In: *Materials Today: Proceedings*, pp. 3253–3257. <https://doi.org/10.1016/j.matpr.2016.10.006>.
- Das, S., Majumder, A., Shukla, V., Suhazsini, P., Radha, P., 2018. Biosynthesis of Poly(3-hydroxybutyrate) from cheese whey by *Bacillus megaterium* NCIM 5472. *J. Polym. Environ.* 26, 4176–4187. <https://doi.org/10.1007/s10924-018-1288-2>.
- Davila-Vazquez, G., Alatríste-Mondragón, F., de León-Rodríguez, A., Razo-Flores, E., 2008. Fermentative hydrogen production in batch experiments using lactose, cheese whey and glucose: influence of initial substrate concentration and pH. *Int. J. Hydrogen Energy* 33, 4989–4997. <https://doi.org/10.1016/j.ijhydene.2008.06.065>.
- Davila-Vazquez, G., Cota-Navarro, C.B., Rosales-Colunga, L.M., de León-Rodríguez, A., Razo-Flores, E., 2009. Continuous biohydrogen production using cheese whey: improving the hydrogen production rate. *Int. J. Hydrogen Energy* 34, 4296–4304. <https://doi.org/10.1016/j.ijhydene.2009.02.063>.
- De Gioannis, G., Friargiu, M., Massi, E., Muntoni, A., Poletti, A., Pomi, R., Spiga, D., 2014. Biohydrogen production from dark fermentation of cheese whey: influence of pH. *Int. J. Hydrogen Energy* 39, 20930–20941. <https://doi.org/10.1016/j.ijhydene.2014.10.046>.
- De Gioannis, G., Muntoni, A., Poletti, A., Pomi, R., Spiga, D., 2017. Energy recovery from one- and two-stage anaerobic digestion of food waste. *Waste Manag.* 68, 595–602. <https://doi.org/10.1016/j.wasman.2017.06.013>.
- Demirel, B., Yenigun, O., Onay, T.T., 2005. Anaerobic treatment of dairy wastewaters: a review. *Process Biochem.* 40, 2583–2595. <https://doi.org/10.1016/j.procbio.2004.12.015>.
- Dessi, P., Asunis, F., Ravishankar, H., Cocco, F.G., De Gioannis, G., Muntoni, A., Lens, P. N.L., 2020. Fermentative hydrogen production from cheese whey with in-line, concentration gradient-driven butyric acid extraction. *Int. J. Hydrogen Energy.* <https://doi.org/10.1016/j.ijhydene.2020.06.081>.
- Diniz, R.H.S., Rodrigues, M.Q.R.B., Fietto, L.G., Passos, F.M.L., Silveira, W.B., 2014. Optimizing and validating the production of ethanol from cheese whey permeate by *Kluyveromyces marxianus* UVF-3. *Biocatal. Agric. Biotechnol.* 3, 111–117. <https://doi.org/10.1016/j.bcab.2013.09.002>.
- Domingos, J.M.B., Puccio, S., Martínez, G.A., Amaral, N., Reis, M.A.M., Bandini, S., Fava, F., Bertin, L., 2018. Cheese whey integrated valorisation: production, concentration and exploitation of carboxylic acids for the production of polyhydroxyalkanoates by a fed-batch culture. *Chem. Eng. J.* 336, 47–53. <https://doi.org/10.1016/j.cej.2017.11.024>.
- Dragone, G., Mussatto, S.I., Almeida e Silva, J.B., Teixeira, J.A., 2011. Optimal fermentation conditions for maximizing the ethanol production by *Kluyveromyces*

- fragilis* from cheese whey powder. *Biomass Bioenergy* 35, 1977–1982. <https://doi.org/10.1016/j.biombioe.2011.01.045>.
- Duque, A.F., Oliveira, C.S.S., Carmo, I.T.D., Gouveia, A.R., Pardelha, F., Ramos, A.M., Reis, M.A.M., 2014. Response of a three-stage process for PHA production by mixed microbial cultures to feedstock shift: impact on polymer composition. *N. Biotech.* 31, 276–288. <https://doi.org/10.1016/j.nbt.2013.10.010>.
- Dąbrowski, W., Żyłka, R., Malinowski, P., 2017. Evaluation of energy consumption during aerobic sewage sludge treatment in dairy wastewater treatment plant. *Environ. Res.* 153, 135–139. <https://doi.org/10.1016/j.envres.2016.12.001>.
- Erguder, T.H., Tezel, U., Guven, E., Demirel, G.N., 2001. Anaerobic biotransformation and methane generation potential of cheese whey in batch and UASB reactors. *Waste Manag.* 21, 643–650. <https://doi.org/10.1016/j.wasman.2005.05.017>.
- Escalante, H., Castro, L., Amaya, M.P., Jaimés, L., Jaimés-Estévez, J., 2018. Anaerobic digestion of cheese whey: energetic and nutritional potential for the dairy sector in developing countries. *Waste Manag.* 71, 711–718. <https://doi.org/10.1016/j.wasman.2017.09.026>.
- Escalante-Hernández, H., Besson, V., Casro-Molano, L., Jaimés-Estévez, J., 2017. Feasibility of the anaerobic digestion of cheese whey in a plug flow reactor (PFR) under local conditions. *Ing. Investig. y Technol.* 18, 265–277.
- European Bioplastics, 2018. Global Production Capacities of Bioplastics 2018–2023. http://www.european-bioplastics.org/wp-content/uploads/2016/02/Report_Bioplastics-Market-Data_2018.pdf.
- European Commission, 2018. A Sustainable Bioeconomy for Europe: Strengthening the Connection between Economy, Society and the Environment Updated Bioeconomy Strategy. <https://doi.org/10.2777/478385>.
- European Commission, 2020. Financing the Green Transition: the European Green Deal Investment Plan and Just Transition Mechanism. https://ec.europa.eu/commission/presscorner/detail/en/ip_20_17.
- European Dairy Association, 2019. The Dairy Sector & the Green Deal. <https://doi.org/10.1787/9789264015906-en>.
- Eurostat, 2018. *Agriculture, Forestry and Fishery Statistics, 2018 Edition*.
- Faria, A., Gonçalves, L., Peixoto, J.M., Peixoto, L., Brito, A.G., Martins, G., 2017. Resources recovery in the dairy industry: bioelectricity production using a continuous microbial fuel cell. *J. Clean. Prod.* 140, 971–976. <https://doi.org/10.1016/j.jclepro.2016.04.027>.
- Ferchichi, M., Crabbe, E., Gil, G.H., Hintz, W., Almadidy, A., 2005. Influence of initial pH on hydrogen production from cheese whey. *J. Biotechnol.* 120, 402–409. <https://doi.org/10.1016/j.jbiotec.2005.05.017>.
- Fernández, C., Cuetos, M.J., Martínez, E.J., Gómez, X., 2015. Thermophilic anaerobic digestion of cheese whey: coupling H₂ and CH₄ production. *Biomass Bioenergy* 81, 55–62. <https://doi.org/10.1016/j.biombioe.2015.05.024>.
- Ferreira Rosa, P.R., Santos, S.C., Sakamoto, I.K., Varesche, M.B.A., Silva, E.L., 2014a. Hydrogen production from cheese whey with ethanol-type fermentation: effect of hydraulic retention time on the microbial community composition. *Bioresour. Technol.* 161, 10–19. <https://doi.org/10.1016/j.biortech.2014.03.020>.
- Ferreira Rosa, P.R., Santos, S.C., Silva, E.L., 2014b. Different ratios of carbon sources in the fermentation of cheese whey and glucose as substrates for hydrogen and ethanol production in continuous reactors. *Int. J. Hydrogen Energy* 39, 1288–1296. <https://doi.org/10.1016/j.ijhydene.2013.11.011>.
- Fradinho, J.C., Oehmen, A., Reis, M.A.M., 2019. Improving polyhydroxyalkanoates production in phototrophic mixed cultures by optimizing accumulator reactor operating conditions. *Int. J. Biol. Macromol.* 126, 1085–1092. <https://doi.org/10.1016/j.jbiomac.2018.12.270>.
- Gabardo, S., Rech, R., Rosa, C.A., Ayub, M.A.Z., 2014. Dynamics of ethanol production from whey and whey permeate by immobilized strains of *Kluyveromyces marxianus* in batch and continuous bioreactors. *Renew. Energy* 69, 89–96. <https://doi.org/10.1016/j.renene.2014.03.023>.
- Gajda, I., Greenman, J., Ieropoulos, I.A., 2018. Recent advancements in real-world microbial fuel cell applications. *Curr. Opin. Electrochem.* 11, 78–83. <https://doi.org/10.1016/j.coelec.2018.09.006>.
- Gelegenis, J., Georgakakis, D., Angelidaki, I., Mavris, V., 2007. Optimization of biogas production by co-digesting whey with diluted poultry manure. *Renew. Energy* 32, 2147–2160. <https://doi.org/10.1016/j.renene.2006.11.015>.
- Ghasemi, M., Ahmad, A., Jafary, T., Azad, A.K., Kakooei, S., Wan Daud, W.R., Sedighi, M., 2017. Assessment of immobilized cell reactor and microbial fuel cell for simultaneous cheese whey treatment and lactic acid/electricity production. *Int. J. Hydrogen Energy* 42, 9107–9115. <https://doi.org/10.1016/j.ijhydene.2016.04.136>.
- Ghimire, A., Luongo, V., Frunzo, L., Pirozzi, F., Lens, P.N.L., Esposito, G., 2017. Continuous biohydrogen production by thermophilic dark fermentation of cheese whey: use of buffalo manure as buffering agent. *Int. J. Hydrogen Energy* 42, 4861–4869. <https://doi.org/10.1016/j.ijhydene.2016.11.185>.
- Gouveia, A.R., Freitas, E.B., Galinha, C.F., Carvalho, G., Duque, A.F., Reis, M.A.M., 2017. Dynamic change of pH in acidogenic fermentation of cheese whey towards polyhydroxyalkanoates production: impact on performance and microbial population. *N. Biotech.* 37, 108–116. <https://doi.org/10.1016/j.nbt.2016.07.001>.
- Guimarães, P.M.R., Teixeira, J.A., Domingues, L., 2010. Fermentation of lactose to bioethanol by yeasts as part of integrated solutions for the valorisation of cheese whey. *Biotechnol. Adv.* 28, 375–384. <https://doi.org/10.1016/j.biotechadv.2010.02.002>.
- Hadiyanto, Ariyanti, D., Aini, A.P., Pinundi, D.S., 2014. Optimization of ethanol production from whey through fed-batch fermentation using *Kluyveromyces marxianus*. *Energy Procedia* 47, 108–112. <https://doi.org/10.1016/j.egypro.2014.01.203>.
- Hagen, L.H., Vivekanand, V., Linjordet, R., Pope, P.B., Eijsink, V.G.H., Horn, S.J., 2014. Microbial community structure and dynamics during co-digestion of whey permeate and cow manure in continuous stirred tank reactor systems. *Bioresour. Technol.* 171, 350–359. <https://doi.org/10.1016/j.biortech.2014.08.095>.
- Hublin, A., Zelić, B., 2013. Modelling of the whey and cow manure co-digestion process. *Waste Management and Research*, pp. 353–360. <https://doi.org/10.1177/0734242X12455088>.
- IEA Bioenergy Task42, 2012. Bio-based Chemicals - Value Added Products from Biorefineries. https://doi.org/10.1007/978-3-319-07593-8_30.
- International Energy Agency, 2017. Energy Statistics. <https://doi.org/10.1017/CBO9781107415324.004>.
- IPCC, 2014. Climate Change 2014: Mitigation of Climate Change, Working Group III Contribution to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. <https://doi.org/10.1017/CBO9781107415416>.
- Joshi, Y., Senatore, B., Poletto, M., 2011. *Kluyveromyces marxianus* biofilm in cheese whey fermentation for bioethanol production. *Chem. Eng. Trans.* 24, 493–498. <https://doi.org/10.3303/CET1124083>.
- Kavacik, B., Topaloglu, B., 2010. Biogas production from co-digestion of a mixture of cheese whey and dairy manure. *Biomass Bioenergy* 34, 1321–1329. <https://doi.org/10.1016/j.biombioe.2010.04.006>.
- Kelly, P.T., He, Z., 2014. Understanding the application niche of microbial fuel cells in a cheese wastewater treatment process. *Bioresour. Technol.* 157, 154–160. <https://doi.org/10.1016/j.biortech.2014.01.085>.
- Koller, M., Bona, R., Chiellini, E., Fernandes, E.G., Horvat, P., Kutschera, C., Hesse, P., Braunnegg, G., 2008. Polyhydroxyalkanoate production from whey by *Pseudomonas hydrogenovora*. *Bioresour. Technol.* 99, 4854–4863. <https://doi.org/10.1016/j.biortech.2007.09.049>.
- Kothari, R., Kumar, V., Pathak, V.V., Tyagi, V.V., 2017. Sequential hydrogen and methane production with simultaneous treatment of dairy industry wastewater: bioenergy profit approach. *Int. J. Hydrogen Energy* 42, 4870–4879. <https://doi.org/10.1016/j.ijhydene.2016.11.163>.
- Labatut, R.A., Angenent, L.T., Scott, N.R., 2011. Biochemical methane potential and biodegradability of complex organic substrates. *Bioresour. Technol.* 102, 2255–2264. <https://doi.org/10.1016/j.biortech.2010.10.035>.
- Lane, M.M., Morrissey, J.P., 2010. *Kluyveromyces marxianus*: a yeast emerging from its sister's shadow. *Fungal Biol. Rev.* 24, 17–26. <https://doi.org/10.1016/j.fbr.2010.01.001>.
- Logan, B.E., Hamelers, B., Rozendal, R., Schröder, U., Keller, J., Freguia, S., Aelterman, P., Verstraete, W., Rabaey, K., 2006. Microbial fuel cells: methodology and technology. *Environ. Sci. Technol.* 40, 5181–5192. <https://doi.org/10.1021/es0605016>.
- Marone, A., Ayala-Campos, O.R., Trably, E., Carmona-Martínez, A.A., Moscoviz, R., Latrille, E., Steyer, J.-P., Alcaraz-Gonzalez, V., Bernet, N., 2017. Coupling dark fermentation and microbial electrolysis to enhance bio-hydrogen production from agro-industrial wastewaters and by-products in a bio-refinery framework. *Int. J. Hydrogen Energy* 42, 1609–1621. <https://doi.org/10.1016/j.ijhydene.2016.09.166>.
- Mazzoli, R., Bosco, F., Mizrahi, I., Bayer, E.A., Pessione, E., 2014. Towards lactic acid bacteria-based biorefineries. *Biotechnol. Adv.* 32, 1216–1236. <https://doi.org/10.1016/j.biotechadv.2014.07.005>.
- Miller, C., Fosmer, A., Rush, B., McMullin, T., Beacom, D., Suominen, P., 2011. Industrial production of lactic acid. *Compr. Biotechnol.* 3, 179–188. <https://doi.org/10.1016/B978-0-08-088504-9.00177-X>.
- Mohan, S.V., Butti, S.K., Amulya, K., Dahiya, S., Modestra, J.A., 2016. Waste Biorefinery: a new paradigm for a sustainable bioelectro economy. *Trends Biotechnol.* 34, 852–855. <https://doi.org/10.1016/j.tibtech.2016.06.006>.
- Montecchio, D., Yuan, Y., Malpei, F., 2018. Hydrogen production dynamic during cheese whey dark fermentation: new insights from modelization. *Int. J. Hydrogen Energy* 43, 17588–17601. <https://doi.org/10.1016/j.ijhydene.2018.07.146>.
- Morais, A.R., Bogel-Lukasik, R., 2013. Green chemistry and the biorefinery concept. *Sustain. Chem. Process.* 1, 18. <https://doi.org/10.1186/2043-7129-1-18>.
- Moreno, R., Escapa, A., Cara, J., Carracedo, B., Gómez, X., 2015. A two-stage process for hydrogen production from cheese whey: integration of dark fermentation and biocatalyzed electrolysis. *Int. J. Hydrogen Energy* 40, 168–175. <https://doi.org/10.1016/j.ijhydene.2014.10.120>.
- Moscoviz, R., Trably, E., Bernet, N., Carrère, H., 2018. The environmental biorefinery: state-of-the-art on the production of hydrogen and value-added biomolecules in mixed-culture fermentation. *Green Chem.* 20, 3159–3179. <https://doi.org/10.1039/c8gc00572a>.
- Oliveira, C.S.S., Silva, M.O.D., Silva, C.E., Carvalho, G., Reis, M.A.M., 2018. Assessment of protein-rich cheese whey waste stream as a nutrients source for low-cost mixed microbial PHA production. *Appl. Sci.* 8, 1817. <https://doi.org/10.3390/app8101817>.
- Ottaviano, L.M., Ramos, L.R., Botta, L.S., Amâncio Varesche, M.B., Silva, E.L., 2017. Continuous thermophilic hydrogen production from cheese whey powder solution in an anaerobic fluidized bed reactor: effect of hydraulic retention time and initial substrate concentration. *Int. J. Hydrogen Energy* 42, 4848–4860. <https://doi.org/10.1016/j.ijhydene.2016.11.168>.
- Pantazaki, A.A., Papaneophytou, C.P., Pritsa, A.G., Liakopoulou-Kyriakides, M., Kyriakidis, D.A., 2009. Production of polyhydroxyalkanoates from whey by *Thermus thermophilus* HB8. *Process Biochem.* 44, 847–853. <https://doi.org/10.1016/j.procbio.2009.04.002>.
- Perna, V., Castelló, E., Wenzel, J., Zampol, C., Fontes Lima, D.M., Borzacconi, L., Varesche, M.B., Zaiat, M., Etchebehere, C., 2013. Hydrogen production in an upflow anaerobic packed bed reactor used to treat cheese whey. *Int. J. Hydrogen Energy* 38, 54–62. <https://doi.org/10.1016/j.ijhydene.2012.10.022>.
- Povolo, S., Toffano, P., Basaglia, M., Casella, S., 2010. Polyhydroxyalkanoates production by engineered *Cupriavidus necator* from waste material containing lactose. *Bioresour. Technol.* 101, 7902–7907. <https://doi.org/10.1016/j.biortech.2010.05.029>.

- Prazeres, A.R., Carvalho, F., Rivas, J., 2012. Cheese whey management: a review. *J. Environ. Manag.* 110, 48–68. <https://doi.org/10.1016/j.jenvman.2012.05.018>.
- Rago, L., Baeza, J.A., Guisasaola, A., 2016. Bioelectrochemical hydrogen production with cheese whey as sole substrate. *J. Chem. Technol. Biotechnol.* 92, 173–179. <https://doi.org/10.1002/jctb.4987>.
- Rai, P.K., Singh, S.P., Asthana, R.K., 2012. Biohydrogen production from cheese whey wastewater in a two-step anaerobic process. *Appl. Biochem. Biotechnol.* 167, 1540–1549. <https://doi.org/10.1007/s12010-011-9488-4>.
- Reis, M.A.M., Serafim, L.S., Lemos, P.C., Ramos, A.M., Aguiar, F.R., Van Loosdrecht, M.C. M., 2003. Production of polyhydroxyalkanoates by mixed microbial cultures. *Bioproc. Biosyst. Eng.* 25, 377–385. <https://doi.org/10.1007/s00449-003-0322-4>.
- Remón, J., García, L., Arauzo, J., 2016. Cheese whey management by catalytic steam reforming and aqueous phase reforming. *Fuel Process. Technol.* 154, 66–81. <https://doi.org/10.1016/j.fuproc.2016.08.012>.
- Rico, C., Muñoz, N., Rico, J.L., 2015. Anaerobic co-digestion of cheese whey and the screened liquid fraction of dairy manure in a single continuously stirred tank reactor process: limits in co-substrate ratios and organic loading rate. *Bioresour. Technol.* 189, 327–333. <https://doi.org/10.1016/j.biortech.2015.04.032>.
- Rivera, I., Bakonyi, P., Cuautle-Marín, M.A., Buitrón, G., 2017. Evaluation of various cheese whey treatment scenarios in single-chamber microbial electrolysis cells for improved biohydrogen production. *Chemosphere* 174, 253–259. <https://doi.org/10.1016/j.chemosphere.2017.01.128>.
- Rosales-Colunga, L.M., Razo-Flores, E., Ordoñez, L.G., Alariste-Mondragón, F., De León-Rodríguez, A., 2010. Hydrogen production by *Escherichia coli* ΔhycA ΔlacI using cheese whey as substrate. *Int. J. Hydrogen Energy* 35, 491–499. <https://doi.org/10.1016/j.ijhydene.2009.10.097>.
- Ryan, M.P., Walsh, G., 2016. The biotechnological potential of whey. *Rev. Environ. Sci. Biotechnol.* 15, 479–498. <https://doi.org/10.1007/s11157-016-9402-1>.
- Saddoud, A., Hassaïri, I., Sayadi, S., 2007. Anaerobic membrane reactor with phase separation for the treatment of cheese whey. *Bioresour. Technol.* 98, 2102–2108. <https://doi.org/10.1016/J.BIORTECH.2006.08.013>.
- Santos, S.C., de Sousa, A.S., Dionísio, S.R., Tramontina, R., Ruller, R., Squina, F.M., Vaz Rossell, C.E., da Costa, A.C., Ienczak, J.L., 2016. Bioethanol production by recycled *Scheffersomyces stipitis* in sequential batch fermentations with high cell density using xylose and glucose mixture. *Bioresour. Technol.* 219, 319–329. <https://doi.org/10.1016/J.BIORTECH.2016.07.102>.
- Shete, B.S., Shinkar, N.P., 2013. Dairy industry wastewater sources, characteristics & its effects on environment. *Int. J. Curr. Eng. Technol.* 3, 1611–1615.
- Sikora, A., Błaszczak, M., Jurkowski, M., Zielenkiewicz, U., 2013. Lactic acid bacteria in hydrogen-producing consortia: on purpose or by coincidence? *Lact. Acid Bact. Food, Heal. Livest. Purp.* 487–514. <https://doi.org/10.5772/50364>.
- Slavov, A.K., 2017. General characteristics and treatment possibilities of dairy wastewater – a review. *Food Technol. Biotechnol.* 55, 14–28. <https://doi.org/10.17113/ft.b.55.01.17.4520>.
- Stamatelatou, K., Antonopoulou, G., Tremouli, A., Lyberatos, G., 2011. Production of gaseous biofuels and electricity from cheese whey. *Ind. Eng. Chem. Res.* 50, 639–644. <https://doi.org/10.1021/ie1002262>.
- Staniszewski, M., Kujawski, W., Lewandowska, M., 2007. Ethanol production from whey in bioreactor with co-immobilized enzyme and yeast cells followed by pervaporative recovery of product – kinetic model predictions. *J. Food Eng.* 82, 618–625. <https://doi.org/10.1016/J.JFOODENG.2007.03.031>.
- Taleghani, H.G., Ghoreyshi, A.A., Najafpour, G.D., 2018. Thin film composite nanofiltration membrane for lactic acid production in membrane bioreactor. *Biochem. Eng. J.* 132, 152–160. <https://doi.org/10.1016/j.bej.2018.01.020>.
- Traversi, D., Bonetta, S., Degan, R., Villa, S., Porfido, A., Bellero, M., Carraro, E., Gilli, G., 2013. Environmental advances due to the integration of food industries and anaerobic digestion for biogas production: perspectives of the Italian milk and dairy product sector. *Bioenerg. Res.* 6, 851–863. <https://doi.org/10.1007/s12155-013-9341-4>.
- Tremouli, A., Antonopoulou, G., Bebelis, S., Lyberatos, G., 2013. Operation and characterization of a microbial fuel cell fed with pretreated cheese whey at different organic loads. *Bioresour. Technol.* 131, 380–389. <https://doi.org/10.1016/j.biortech.2012.12.173>.
- Tribst, A.A.L., Falcade, L.T.P., de Oliveira, M.M., 2019. Strategies for raw sheep milk storage in smallholdings: effect of freezing or long-term refrigerated storage on microbial growth. *J. Dairy Sci.* 102, 4960–4971. <https://doi.org/10.3168/jds.2018-15715>.
- Ubando, A.T., Felix, C.B., Chen, W.-H., 2020. Biorefineries in circular bioeconomy: a comprehensive review. *Bioresour. Technol.* 299, 122585. <https://doi.org/10.1016/j.biortech.2019.122585>.
- Valentino, F., Karabegovic, L., Majone, M., Morgan-Sagastume, F., Werker, A., 2015. Polyhydroxyalkanoate (PHA) storage within a mixed-culture biomass with simultaneous growth as a function of accumulation substrate nitrogen and phosphorus levels. *Water Res.* 77, 49–63. <https://doi.org/10.1016/j.watres.2015.03.016>.
- Vassilev, I., Hernandez, P.A., Batlle-Vilanova, P., Freguia, S., Krömer, J.O., Keller, J., Ledezma, P., Virdis, B., 2018. Microbial electrosynthesis of isobutyric, butyric, caproic acids, and corresponding alcohols from carbon dioxide. *ACS Sustain. Chem. Eng.* 6, 8485–8493. <https://doi.org/10.1021/acsuschemeng.8b00739>.
- Venetsaneas, N., Antonopoulou, G., Stamatelatou, K., Kornaros, M., Lyberatos, G., 2009. Using cheese whey for hydrogen and methane generation in a two-stage continuous process with alternative pH controlling approaches. *Bioresour. Technol.* 100, 3713–3717. <https://doi.org/10.1016/j.biortech.2009.01.025>.
- Venkata Mohan, S., Mohanakrishna, G., Velvizhi, G., Babu, V.L., Sarma, P.N., 2010. Bio-catalyzed electrochemical treatment of real field dairy wastewater with simultaneous power generation. *Biochem. Eng. J.* 51, 32–39. <https://doi.org/10.1016/j.bej.2010.04.012>.
- Vivekanand, V., Mulat, D.G., Eijsink, V.G.H., Horn, S.J., 2018. Synergistic effects of anaerobic co-digestion of whey, manure and fish ensilage. *Bioresour. Technol.* 249, 35–41. <https://doi.org/10.1016/J.BIORTECH.2017.09.169>.
- Wei, P., Cheng, L.H., Zhang, L., Xu, X.H., Chen, H.L., Gao, C.J., 2014. A review of membrane technology for bioethanol production. *Renew. Sustain. Energy Rev.* 30, 388–400. <https://doi.org/10.1016/j.rser.2013.10.017>.
- Wenzel, J., Fuentes, L., Cabezas, A., Etchebehere, C., 2017. Microbial fuel cell coupled to biohydrogen reactor: a feasible technology to increase energy yield from cheese whey. *Bioproc. Biosyst. Eng.* 40, 807–819. <https://doi.org/10.1007/s00449-017-1746-6>.
- Xu, Q., Ouyang, J., Liu, P., Liu, J., Qian, Z., Zheng, Z., 2018. Valorization of dairy waste for enhanced D-lactic acid production at low cost. *Process Biochem.* 71, 18–22. <https://doi.org/10.1016/j.procbio.2018.05.014>.
- Yang, P., Zhang, R., McGarvey, J.A., Benemann, J.R., 2007. Biohydrogen production from cheese processing wastewater by anaerobic fermentation using mixed microbial communities. *Int. J. Hydrogen Energy* 32, 4761–4771. <https://doi.org/10.1016/j.ijhydene.2007.07.038>.
- Yazar, E., Cirik, K., Ozdemir, S., Akman, D., Cuci, Y., Cinar, O., 2016. Optimization of two-stage and single-stage anaerobic reactors treating cheese whey. *J. Eng. Sci.* 19, 1. <https://doi.org/10.17780/ksujes.18309>.
- Yilmazer, G., Yenigün, O., 1999. Two-phase anaerobic treatment of cheese whey. *Water Sci. Technol.* 40, 289–295. [https://doi.org/10.1016/S0273-1223\(99\)00397-2](https://doi.org/10.1016/S0273-1223(99)00397-2).
- Zoppellari, F., Bardi, L., 2013. Production of bioethanol from effluents of the dairy industry by *Kluyveromyces marxianus*. *N. Biotech.* 30, 607–613. <https://doi.org/10.1016/j.nbt.2012.11.017>.