

Daily submicron particle doses received by populations living in different low- and middle-income countries

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Abstract

In the present study, the daily dose in terms of particle surface area received by citizens living in different low- and middle-income countries, characterized by different lifestyles, habits, and climates, was evaluated. The level of exposure to submicron particles and the dose received by the populations of Accra (Ghana), Cairo (Egypt), Florianópolis (Brazil), and Nur-Sultan (Kazakhstan) were analyzed. A direct exposure assessment approach was adopted to measure the submicron particle concentration levels of volunteers at a personal scale during their daily activities. Non-smoking adult volunteers performing non-industrial jobs were considered. Exposure data were combined with time-activity pattern data (characteristic of each population) and the inhalation rates to estimate the daily dose in terms of particle surface area.

43 The received dose of the populations under investigation varied from 450 mm² (Florianopolis,
44 Brazil) to 1300 mm² (Cairo, Egypt). This work highlights the different contributions of the
45 microenvironments to the daily dose with respect to high-income western populations. It was
46 evident that the contribution of the Cooking & Eating microenvironment to the total exposure
47 (which was previously proven to be one of the main exposure routes for western populations)
48 was only 8%-14% for low- and middle-income populations. In contrast, significant
49 contributions were estimated for Outdoor day and Transport microenvironments (up to 20% for
50 Cairo, Egypt) and the Sleeping & Resting microenvironment (up to 28% for Accra, Ghana),
51 highlighting the effects of different site-specific lifestyles (e.g. time-activity patterns), habits,
52 socioeconomic conditions, climates, and outdoor air quality.

53
54 **Keywords:** airborne particles; dose; lung-deposited surface area; personal monitoring;
55 submicron particles; ultrafine particles.

56
57 **Capsule:** A direct exposure assessment approach was adopted to measure the particle
58 concentrations whose people is exposed to and to estimate the dose of sub-micron particles
59 received by people living in low and middle-income countries. The contribution of the
60 microenvironments and lifestyles on the total daily dose was assessed.

62 1 Introduction

63 Several scientific studies reported a relationship between inhalation (and consequent deposition
64 in the respiratory tract) of airborne particles and adverse effects for human health, including
65 respiratory diseases and inflammation (Schmid et al., 2009), cardiovascular diseases (Buteau &
66 Goldberg, 2016), diabetes (Brook et al., 2008), higher systolic blood pressure and pulse pressure
67 (Auchincloss et al., 2008), decreased cognitive function (Power et al., 2011), heart rate increase
68 (Rizza et al., 2019), changes in the brain wave pattern (Naseri et al., 2019), and lung cancer
69 (Beelen et al., 2014; International Agency for Research on Cancer, 2013). Recently, the
70 scientific community has shifted its interest from super-micrometric particles (characterized in
71 terms of the mass concentrations of particles smaller than 10 µm and 2.5 µm, i.e. PM₁₀ and
72 PM_{2.5}, respectively) to submicron and ultrafine particles (UFPs, particles smaller than 100 nm),
73 which are usually characterized in terms of particle number and surface area concentrations
74 instead of mass concentration (Franck et al., 2011; Giechaskiel et al., 2009; Kumar et al., 2014).
75 In particular, the surface area of the particles deposited in the lungs was recognized the most
76 appropriate aerosol metrics when evaluating the effect of UFPs on human health (Sager &
77 Castranova, 2009; Schmid & Stoeger, 2016; Stoeger et al., 2006).

78 Nonetheless, particle-phase pollutants and air quality standards worldwide are still defined in
79 terms of outdoor PM₁₀ and PM_{2.5} levels; moreover, the outdoor sampling sites are limited to a
80 certain number of fixed sampling points (FSPs), then not representing the effective citizen
81 exposure to PM fractions due to the lacking of proper links to climate or population lifestyle
82 (Buonanno et al., 2010; European Parliament and Council of the European Union, 2008; Rizza
83 et al., 2017). Moreover, due to the different dynamics (e.g. dilution, deposition) and sources
84 among supermicron and submicron particles, the measurements of PM fractions through FSPs
85 cannot be considered proxies for the exposure to submicron and ultrafine particles (Kaur et al.,
86 2005; Kumar et al., 2011; Neft et al., 2016). Additionally, the outdoor concentration levels are
87 often uncorrelated with the indoor ones; thus, exposure assessment based on outdoor FSP
88 measurements (i.e. city scale or outdoor scale approaches) could significantly underestimate
89 the exposure in an indoor environment, which represents the environment where the population
90 spends most of the day (Scungio et al., 2020). Therefore, to obtain a better representativeness
91 of the complete human exposure to submicron and ultrafine particle measurements at a personal
92 scale, personal monitoring approaches should be adopted (Pacitto et al., 2020; Scungio et al.,
93 2020).

94 Along with exposure data, information on the sort of activity is additionally needed to estimate
95 the inhalation and consequent deposition of particles in several regions of the lungs
96 (International Commission on Radiological Protection, 1994) as well as the time-activity
97 pattern data (i.e. time spent in each indoor or outdoor environment). Indeed, as shown in
98 previous studies, people's different lifestyles lead to significantly different exposures to
99 airborne particles (Bekö et al., 2015; Pacitto et al., 2018; Schweizer et al., 2006). As an example,
100 in a previous study the daily dose received by different high-income western populations was
101 estimated applying a direct exposure assessment approach, i.e. characterizing the exposure to
102 airborne particles through personal monitoring, then evaluating the daily dose received by the
103 typical population whose time-activity pattern is known. The study highlighted significant
104 differences among the five western world populations investigated in terms of the daily dose
105 received, highlighting the influence of the lifestyle (e.g. massive use of particle sources indoors
106 as well as outdoors) and ventilation of the indoor environment (e.g. widespread use of
107 mechanical ventilation systems) (Pacitto et al., 2018). Thus, previous studies opened questions
108 on the effect of lifestyle and geographical location (e.g. climate, outdoor concentration) on
109 airborne particle doses received by the populations. Nonetheless, the amount of studies on this
110 subject are still limited and, typically, they consider populations in developed and high-income
111 countries characterized by a stronger sensitivity towards air quality topics (Fernández-Iriarte et

112 al., 2020; Moreno et al., 2019; Pacitto et al., 2019; Scungio, Rizza, Stabile, Morawska, &
113 Buonanno, 2020; Tomassetti et al., 2020). In contrast, in low- and middle-income countries the
114 race for industrialization tends to overshadow air quality-related aspects, such as the utilization
115 of highly emitting cooking fuels and fewer efficient vehicles (Mehta et al., 2013; Puzzolo et al.,
116 2019). Therefore, the evaluation of the exposure to submicron particles and the daily dose in
117 these countries is crucial because it could help in detecting the principal activities and
118 environments contributing to the daily dose of the populations under investigation. This might
119 lead to more detailed knowledge of the factors that contribute to possible adverse health effects
120 and outcomes and of related prevention or mitigation strategies that may be adopted.

121 To this end, in the present work we carried out personal exposure field campaigns in four major
122 cities in low- and middle-income countries: Greater Cairo (Egypt), Greater Accra (Ghana),
123 Florianopolis (Brazil), and Nur-Sultan (Kazakhstan). In particular, a direct UFP exposure
124 assessment approach was applied to estimate the daily dose received by the typical populations
125 of those cities in relation to the activity patterns and inhalation rates typical of those populations.
126 The obtained data were compared to those of a previous work focused on the UFP exposure of
127 the populations of western and high-income countries. The main influence parameters affecting
128 the dose of the different populations were determined and discussed.

129 **2 Methodology**

130 **2.1 Study area**

131 Measurements were carried out in Greater Cairo (Egypt), Greater Accra (Ghana), Florianopolis
132 (Brazil), and Nur-Sultan (Kazakhstan). The choice of the cities was based on the existing
133 research collaboration network capable of supporting the study. A brief description of the cities
134 under investigation is provided below to characterize their climatic conditions and main indoor
135 and outdoor airborne particle sources. Indeed, the climate and location of the cities investigated
136 affect the outdoor particle concentration levels and the resulting contribution to the populations'
137 daily dose. Because much of the daily dose is typically contributed to by indoor environments,
138 information on possible indoor sources (Morawska et al., 2013; Scungio et al., 2017) and
139 building ventilation (Bakó-Biró et al., 2012; Stabile et al., 2017; Stephens & Siegel, 2012) are
140 useful as they can affect the exposures and doses of the people living/residing therein
141 (Buonanno et al., 2015b).

142 Greater Cairo (Cairo, Giza and Shoubra El-Kheima and their urban agglomeration) is the largest
143 urban concentration in Africa and the Middle-East (surface area of about 1700 km²) with

144 approximately 20 million permanent residents; it hosts various industrial complexes and is
145 characterized by intensive vehicular traffic with almost 3.2 million cars (CAPMAS, 2018).
146 Similar to most of the Egyptian terrains, Greater Cairo's climate is classified as a hot desert
147 climate (Köppen-Geiger classification) and is thus characterized by warm winters (14–22 °C
148 on average), hot summers (20–35 °C on average), and sparse rainfall (Peel et al., 2007). The
149 prevailing wind direction is northwesterly, although southern wind storms loaded with Saharan
150 dust are frequent during spring. Both natural and anthropogenic sources contribute to the very
151 high airborne particulate matter concentrations in Greater Cairo, with annual average
152 concentrations of 161 $\mu\text{g m}^{-3}$ and 71 $\mu\text{g m}^{-3}$ for PM₁₀ and PM_{2.5}, respectively (EEAA, 2016).
153 The main indoor particle source is fossil-fuel combustion (including coal, natural gas, propane,
154 and kerosene) for cooking and heating.

155 The Greater Accra region is located in the Republic of Ghana on the West African coast, with
156 a surface area of about 3245 km² and a population of approximately 4 million permanent
157 residents. The climate is classified as a tropical savanna climate (Köppen-Geiger classification)
158 with an annual average rainfall of 730 mm and very little variation in temperature throughout
159 the year (25.9–29.6 °C). The main anthropogenic sources of atmospheric air pollution in the
160 city include vehicle emissions and combustion activities, which account for over 70% of
161 outdoor air pollution in Accra (Arku et al., 2015), and the high percentage of old vehicles,
162 characterized by less fuel-efficient engines, used in the city. A further outdoor source is the
163 widespread use of biomass fuels (firewood and charcoal) for cooking in underprivileged
164 communities. For indoor air quality, cooking and other combustion activities, such as burning
165 of anti-mosquito coils in bedrooms at night, represent the main indoor sources; indeed, several
166 people living in low-income countries burn coils at night to expel mosquitos (Nyarku et al.,
167 2019).

168 Florianopolis is the second largest urban concentration in Santa Catarina State, located in the
169 South of Brazil, with a surface area of 675 km² and a population of about 500,000 people
170 (IBGE, 2018). The climate of Florianopolis is humid subtropical (Köppen-Geiger
171 classification), with annual average temperature and precipitation of 20.1 °C and 1462 mm,
172 respectively. The air quality in Florianopolis complies with the permissible limits for PM₁₀; the
173 average concentration of PM₁₀ during the period 2011–2016 was 24 $\mu\text{g m}^{-3}$ (Vasques et al.,
174 2017). The main outdoor PM source is vehicular traffic since Santa Catarina has the highest
175 vehicle density in Brazil (49.8 cars km⁻²).

176 Nur-Sultan city is the capital of the Republic of Kazakhstan and is located in the center of the
177 country and at the center of Central Asia; it has a surface area of 722 km² and 1 million citizens.

178 The climate of Nur-Sultan is humid continental (Köppen climate classification) with an average
179 annual temperature of 3.5 °C. Indeed, Nur-Sultan has an extreme continental climate with warm
180 summers (featuring occasional brief rain showers) and long, very cold, dry winters. The main
181 outdoor PM sources are road traffic and two power plants, due to the widespread consumption
182 of coal as a fuel and energy source.

183 **2.2 Study design**

184 To characterize and compare the submicron particle daily doses received by the populations in
185 four low- and middle-income countries, we carried out an experimental campaign following a
186 direct exposure assessment approach, i.e. measuring the exposure to airborne particle
187 concentrations at a personal scale by means of handheld particle counters (Pacitto et al., 2018;
188 Pacitto et al., 2020; Wallace & Ott, 2011). The study was limited to the non-smoking population
189 and non-industrial working environments; the smoking population should be analyzed
190 separately due to the large dose they typically receive, as should people working in industrial
191 jobs characterized by specific particle sources (Fuoco et al., 2017). Thus, the volunteers
192 considered in this study, to characterize the different environments in terms of submicron
193 particle concentrations, were all adults (both male and female), office workers, and non-
194 smokers.

195 To estimate the daily dose of submicron airborne particles received by the populations living
196 in the cities under investigation, the following steps were applied:

- 197 – identification of the typical time-activity patterns characteristic of citizens living in each
198 country;
- 199 – measurement of the concentrations of particle numbers and lung-deposited particle
200 surface area to which citizens are exposed in different environments and during different
201 activities;
- 202 – estimation of the daily dose received by the typical population living in the investigated
203 cities.

204 **2.2.1 Population time-activity patterns under investigation**

205 The typical time-activity patterns characteristic of citizens living in the investigated cities were
206 obtained from national human activity pattern surveys performed by national statistics institutes
207 and/or research studies taking into account people living both in urban areas and country sites
208 (Central Agency for Public Mobilization and Statistics of Egyptian Arabic Republic, 2016;
209 2017; Ghana Statistical Service, 2012; 2014; 2015) except for the Brazilian population. In

210 particular, time-activity patterns of people from 19 to 65 years old were considered. A
211 descriptive exploratory study was performed for Brazilian people due to the lack of information
212 about the time-activity pattern characteristics of this population. Data were collected from a
213 non-probabilistic sample by quotas and convenience in people living in the metropolitan region
214 of Florianopolis. It was not possible to develop an epidemiological study or a probabilistic
215 sample of the entire population of the area due to the COVID-19 pandemic. However, the
216 significant number of participating subjects ($n = 435$) can be considered acceptable as a sample
217 of the general population. The questionnaire was filled out through the Google form platform.
218 The data were analyzed using IBM SPSS Statistics version 20.0. Data were analyzed with
219 descriptive statistics such as mean and standard deviation.

220 ***2.2.2 Measurement of the exposure to submicron particle concentrations***

221 The exposure to airborne particles in each environment of the four cities was obtained by
222 performing mobile measurements of particle number (PN) concentrations, average particle
223 sizes (D_p), and lung-deposited surface area (LDSA) concentrations through direct
224 measurements carried out with hand-held diffusion charger particle counters. The experimental
225 campaigns were carried out from October 2017 to December 2017 in Accra (Ghana), from
226 December 2018 to March 2019 in Cairo (Egypt), from December 2019 to March 2020 in
227 Florianopolis (Brazil), and from March 2020 to June 2020 in Nur-Sultan (Kazakhstan). Details
228 of the meteorological conditions in the four cities during the experimental campaigns are
229 reported in Table 1.

230

231 To obtain a significant amount of data for each environment or activity, 60 volunteers (15 for
232 each city) were selected. The volunteers, both males and females aged from 25 to 55 years old,
233 lived and worked in the urban area of the city. Personal particle counters were worn by
234 volunteers for 3 days; therefore, each volunteer contributed continuous measurements for
235 3 days. When volunteers were at home/office (e.g. working, studying, sleeping, etc.), the
236 instrument was left on a close desk. In addition, the volunteers had to fill out an activity diary
237 to take note about (i) place, (ii) time, and (iii) kind of activity, to relate the particle exposure
238 values to the activity and the specific environments. The authors highlight that the 15 volunteers
239 for each city were chosen in order to cover a wide range of activities and to better characterize
240 the concentrations in the different environments, following the methodology already used in
241 the previous study by Pacitto et al., (2018). For example, different measurements during
242 cooking activities allow the effects of different cooking practices on the exposure levels to
243 airborne particles to be included (such as the type of food, cooking, and stove). Additionally,

244 several measurements in the transport environments allow different transport modes (such as
245 metro, train, car) to be taken into account. The authors point out, that in the present study we
246 were not interested in the volunteer per se, the volunteers just acted as mobile platforms
247 allowing moving the instruments across the different environments. Thus, we are not really
248 interested in completely describe the volunteers, we just avoided to enroll volunteers living in
249 the same house or belonging to the same groups (i.e. family, friend, age, job), in order to avoid
250 obtaining exposure data only for a very specific sub-population.

251 **2.2.3 The daily dose of the population**

252 The doses of submicron particles received by the population were calculated from the typical
253 exposure to lung-deposited surface area concentrations (obtained from the experimental
254 analysis described in section 2.2.2) and time-activity patterns. In particular, the total surface
255 area dose ($\delta_{SA,j}$) for the j -th activity was obtained by applying the following equation (Buonanno
256 et al., 2011; Buonanno et al., 2012):

$$257 \quad \delta_{SA,j} = IR_j \cdot LDSA_j \cdot T_j \quad (1)$$

258 where IR_j ($m^3 h^{-1}$) is the inhalation rate of the j -th activity (which depends on the age and
259 activity level) (Buonanno et al., 2012), T_j is the time spent on each activity (as a function of the
260 age and sex), and $LDSA_j$ is the concentration in terms of lung-deposited surface area of
261 submicron particles during the j -activity. The information on time-activity patterns obtained
262 from the activity diary was not used to calculate the dose because this was specific to the
263 volunteers under investigation. The time-activity diary was simply used to relate the particle
264 exposure values to the environments/activities in order to characterize them in terms of the
265 particle concentration of each environment they resided in and the activities they performed.
266 The dose $\delta_{SA,j}$ estimated through eq. (1) represents the median dose received during that activity
267 as it was calculated considering the median $LDSA$ concentrations and the average time.

268 The activities performed by the population, obtained on the basis of the typical time-activity
269 patterns, were grouped into six main microenvironments, as reported in Table 2. The dose for
270 each microenvironment was obtained by summing up the median doses evaluated for each
271 activity; similarly, the daily dose was obtained as the sum of the doses received in each
272 microenvironment. To compare the doses in different microenvironments, the dose intensity
273 ratio (i.e. the ratio between the daily dose fraction and the daily time fraction characteristics of
274 each activity) was also evaluated (Buonanno et al., 2014; Pacitto et al., 2018).

275 **2.2.4 Comparison with high-income western countries**

276 In order to compare the dose received by the low- and middle-income populations to high-
277 income western countries we used the data reported in our abovementioned previous paper
278 (Pacitto et al., 2018). In that paper, we applied the same methodology and instrumentation here
279 shown to gather exposure data for five different western populations. In particular, we
280 performed the experimental analyses in Barcelona (Spain), Cassino (Italy), Guilford (United
281 Kingdom), Brisbane (Australia), and Lund (Sweden). For detailed information regarding the
282 climate, location, and anthropogenic sources of these five cities please refer to Pacitto et al.,
283 (2018). For the sake of brevity, here we just summarize the measurement periods of the
284 experimental campaigns, as they may affect the people exposure. In particular, measurements
285 in Barcelona, Cassino, Guilford, Brisbane, and Lund were performed on Oct-Dec 2015 (average
286 temperature and relative humidity of 18 °C and 73%), Apr-June 2016 (17 °C, 74%), Aug-Nov
287 2016 (17 °C, 81%), Jun-Aug 2016 (27 °C, 79%), and Sept-Dec 2016 (10 °C, 78%),
288 respectively. A further aspect to be highlighted, as it may help in explaining the indoor exposure
289 levels, is the widespread use of mechanical ventilation systems in Lund.

290 The experimental analyses in the five high income countries were carried out with the support
291 of 15 volunteers (for each city) as we did in the current study adopting the same criteria for
292 their enrollment. Analogously, also in the previous experimental analysis, the volunteers were
293 asked to take note about place, time, and kind of activity in an activity diary.

294 The exposure data post-processing and dose estimates were carried out on the basis of the
295 methodology reported in the present paper.

296 **2.3 Instrumentation and quality assurance**

297 Particle number (PN) concentrations, lung-deposited surface area (LDSA) concentrations and
298 average particle sizes (D_p) were measured through three diffusion charger particle counters:

- 299 – the Aerasense NanoTracer XP (Oxility, partner of Royal Philips Electronics), which
300 measures PN and LDSA concentrations and D_p in the range 10–300 nm with 10 seconds
301 sampling time, was used in the experimental analyses performed in Ghana;
- 302 – the DiscMini (Testo), which measures PN and LDSA concentrations and D_p in the range 10–
303 700 nm with 1 second sampling time, was used in the experimental analyses performed in
304 Egypt and Brazil;
- 305 – the Partector 2 (Naneos Particle Solutions GmbH), which measures PN and LDSA
306 concentrations and the D_p in the range 10–300 nm with 1 second sampling time, was adopted
307 in the experimental analyses performed in Kazakhstan.

308 The operating principle of the instruments is based on the diffusion charging technique.
309 Specifically, the sampled aerosol is charged in a positive unipolar diffusion charger imparting
310 an average known charge on the particles that is approximately proportional to the particle
311 diameter of the aerosol. The number of charges, and thus the number of particles, is then
312 detected by an electrometer (Buonanno et al., 2014; Fierz et al., 2011; Marra et al., 2010). Since
313 over 99% of total PN concentration in urban environments is contributed by particles below
314 300 nm in diameter (Goel & Kumar, 2014; Kumar et al., 2013), all the instruments were able
315 to measure total particle numbers despite their different ranges for particle diameters. On the
316 basis of the measured values of PN concentration and D_p , the LDSA concentration was
317 calculated by means of built-in semi-empirical correlation as reported and discussed in previous
318 papers (Marra et al., 2010; Pacitto et al., 2018; Stabile et al., 2015).

319 **2.3.1 Instrument intercomparison**

320 Calibrations were performed before and after the experimental campaigns. In particular, the
321 NanoTracer, DiscMini, and Naneos Partector 2 were compared to a CPC (model 3775, TSI
322 Inc.) and SMPS system (model 3936, TSI Inc.) in terms of particle number concentration and
323 particle size, respectively, before and after each experimental campaign. Calibrations were
324 carried out at the European Accredited Laboratory of Industrial Measurements (LaMI) of the
325 University of Cassino and Southern Lazio (Italy) while placed in a 150 m³ room, with an
326 ordinary mechanical ventilation system guaranteeing constant thermo-hygrometric conditions
327 (20 ± 2 °C and $50 \pm 5\%$ relative humidity). Comparisons were performed for two different
328 aerosols: aged indoor aerosol and freshly emitted aerosol produced by incense burning. Tests
329 were performed by measuring for 2 h simultaneously with the diffusion charger monitors
330 (Nanotracer, DiscMini or Naneos Partector 2), CPC 3775, and SMPS 3936. The average
331 correction factors, in terms of particle number concentration, obtained by averaging the two
332 aerosols investigated before and after the experimental campaign, were applied as correction
333 factors for the entire campaigns.

334 **2.3.2 Statistical analysis**

335 The exposure levels to particle number and lung deposited surface area concentrations in the
336 different microenvironments were checked in order to determine whether they were statistically
337 different or not. In particular, we compared the nine cities, microenvironment by
338 microenvironment; thus, nine distinct statistical analyses (one for each microenvironment) were
339 carried out for particle number and lung deposited surface area concentrations, respectively. To
340 this end, preliminary normality tests (Shapiro-Wilk test) were performed to check the statistical

341 distribution of the data. Since the data did not meet Gaussian distributions, non-parametric tests
342 and further post-hoc tests, the Kruskal-Wallis test (Kruskal & Wallis, 1952), were conducted
343 for each microenvironment. The statistically significant result was referred to a significance
344 level of 99% in order to reduce the probability of type I error (i.e. the data were assessed as
345 statistical different for p -value lower than $\alpha=0.01$). The authors point out that the population
346 were tested just for one parameter (particle number or lung deposited surface area
347 concentrations separately, thus no adjustments of statistical inference (type I error) were needed
348 (Chen et al., 2017).

349 **3 Results and Discussions**

350 **3.1 Time-activity pattern**

351 Data on the time-activity patterns of the populations under investigation, and the contribution
352 of the individual microenvironment for the whole day as a function of sex for the four countries,
353 are shown in Table 3. The data clearly demonstrate that all populations, both males and females,
354 spend the largest time fraction indoors (in a range from 84% to 96%). This finding matches
355 with the results found in our previous studies on high-income western populations where,
356 whatever the population (Italian, Australian, Spanish, English and Swedish) and the sex, indoor
357 environments were those where people spend the largest daily time fraction, ranging from 88%
358 to 95% of their time. Cooking & Eating, recognized as the most significant activities in terms
359 of exposure (Buonanno et al., 2015b; Buonanno et al., 2017), presented here jointly a
360 contribution ranging from 4.0% to 10.7% – lower than the range found in Pacitto et al. (2018)
361 (from 13% to 26%) for western populations. In fact, as shown in Table 3, western populations
362 spend more time in the Cooking & Eating microenvironment than the Egyptian, Brazilian and
363 Ghanaian populations (but not the Kazakhstan population); as an example, the time spent in the
364 Cooking & Eating microenvironment by Spanish people was almost three-fold that spent by
365 Egyptian people.

366 Another microenvironment typically recognized as critical in terms of people exposure is the
367 Transport microenvironment (Scungio et al., 2015; Scungio et al., 2013; Stabile et al., 2015);
368 nonetheless, the exposure times in such microenvironments were quite short (as already
369 recognized for western populations in our previous paper) and did not vary greatly amongst the
370 populations investigated: they ranged from 4.3% to 6.9% (except for Kazakhstan females who
371 spend 2.8% of the day in such microenvironments).

372 The Sleeping & Resting microenvironment, as expected, accounted for a third of the day; in
373 particular, the time fraction ranges roughly from 30% (Kazakhstan females) to 40% (Egyptian
374 females), which is quite in line with the western population. The Indoor day microenvironment,
375 as seen in Table 3, displays huge variability as a function of nationality; in fact, the contribution
376 of this microenvironment to the whole day varies from 11.4% to 22.1%. Comparing the Indoor
377 day time fractions presented here to those obtained for western countries by Pacitto et al. (2018)
378 (and not reported here for the sake of brevity), it can clearly be seen that high-income western
379 populations spend more time in indoor microenvironments than the low- and middle-income
380 populations analyzed here.

381 Finally, in the Outdoor day microenvironment, there is an evident difference between the
382 Kazakhstan population and the other populations. In fact, the contribution to the total daily time
383 for the Kazakhstan population varies between 1.7% and 2.6%, likely due to the severe
384 meteorological conditions characteristics of the country; whereas for the other populations the
385 contribution is much higher (up to 10%) and also higher than those characteristic of western
386 populations (4%-8%).

387 **3.2 Concentration levels of submicron particles in the investigated cities**

388 Figure 1 and Table 4 report statistics of submicron particles in terms of (i) PN concentration,
389 (ii) LDSA concentration, and (iii) average particle size as a function of the microenvironments
390 for the investigated populations. Moreover, data of the five western countries considered in our
391 previous study were also reported (and are re-analyzed here accordingly) in order to perform
392 comparisons amongst the populations.

393 The data from western countries, discussed in detail in our previous paper and here just
394 summarized for the sake of brevity, highlighted that: (i) the highest PN concentrations are
395 typically measured in the Cooking & Eating microenvironment (median values in the range
396 $0.16\text{--}3.05 \times 10^5$ part. cm^{-3}), (ii) the exposure in Traffic and Outdoor day microenvironments are
397 mostly lower than that experienced in other indoor activities, including the Indoor day
398 microenvironment, (iii) the lowest concentrations occur in the Sleeping microenvironment
399 (median values in the range $0.8\text{--}1.2 \times 10^4$ part. cm^{-3}), and (iv) the widespread use of mechanical
400 ventilation systems indoors, including homes (see Lund exposure data, Table 4), can strongly
401 reduce (by roughly one order of magnitude) the exposure to submicron particle concentrations
402 in indoor environments and activities.

403 Exposure data characteristic of the populations investigated in the present study reveal
404 significant differences with respect to the high-income western countries. The first clear finding

405 concerns the comparison between the Cooking & Eating and Transportation
406 microenvironments. Indeed, in the four cities analyzed, the exposure levels in these two
407 microenvironments were roughly comparable; for Egyptian and Kazakhstani people, the PN
408 and LDSA concentrations in the Transport microenvironment were even higher than in Cooking
409 & Eating. As an example, median values of 1.73×10^5 part. cm^{-3} and 2.88×10^2 $\mu\text{m}^2 \text{cm}^{-3}$ were
410 measured for Egyptians, likely due to the intensive vehicular traffic and the particular
411 meteorological condition (e.g. no rain). The exposure levels (in terms of both PN and LDSA) in
412 the Transport microenvironments of the Egyptian and Kazakhstani populations were also larger
413 than those in the western populations. Similarly, in these two cities, the exposure levels in the
414 Outdoor day microenvironment (in terms of both PN and LDSA) were significantly larger than
415 those experienced in western countries, with median values of up to 5×10^4 part. cm^{-3} /
416 1×10^2 $\mu\text{m}^2 \text{cm}^{-3}$ for the Egyptian population. These data emphasize the important role of the
417 outdoor air quality in the overall exposure of a population. Moreover, poor outdoor air quality
418 can also affect indoor microenvironments where no other particle sources are typically in use.
419 As an example, the data in Figure 1 and Table 4 clearly highlight that the exposure to submicron
420 particles in the Working and Sleeping & Resting microenvironments in Cairo and Nur-Sultan
421 reflects the outdoor conditions (likely due to particle penetration), thus resulting in
422 concentrations typically higher than those received in the western countries previously
423 investigated.

424 For the Sleeping & Resting microenvironment, high particle concentrations were measured for
425 the population living in Accra (Ghana) (median value of 4.18×10^4 part. cm^{-3} and
426 9.45×10^1 $\mu\text{m}^2 \text{cm}^{-3}$). This is due to the common use of anti-mosquito products (e.g. mosquito
427 coils) during the night in Ghana (Nyarku et al., 2019), as they have gained widespread
428 popularity in malaria-endemic countries (Hogarh et al., 2016). This could also partly explain
429 the higher concentrations in the Indoor day microenvironment of Ghana in relation to Egypt
430 and Kazakhstan.

431 As mentioned above, the exposure in the Cooking & Eating microenvironment, which was the
432 most critical for western countries, was mostly lower or similar to those measured in other
433 microenvironments for the four populations analyzed in the present study. Several reasons
434 could explain this finding, including the (unknown) emission factors of cooking activities
435 typically performed in such countries (likely related to the type of food, cooking activity, fuel,
436 and stoves used). Nonetheless, a further reason could be strictly related to the climatic
437 conditions; indeed, higher outdoor temperatures would lead to better airing of the homes than

438 in the western populations (whose data were mostly collected in periods of lower temperature),
439 thus leading to faster particle concentration decay.

440 The results for Florianopolis were somewhat different from the other three populations
441 investigated. Indeed, the concentrations were generally lower than those measured in the other
442 three cities; indeed, the highest exposure level was measured in the Cooking & Eating
443 microenvironment, which was actually lower than those characteristics of the western cities
444 previously analyzed (median PN and LDSA concentrations were 1.94×10^4 part. cm^{-3} and
445 $4.10 \times 10^1 \mu\text{m}^2 \text{cm}^{-3}$). This could be partly due to climatic conditions (once again the average
446 outdoor temperature during the campaign was significantly higher than those occurring during
447 the western population characterization) but also to the socioeconomic conditions of Santa
448 Catarina. Although Brazil is not considered a high-income country, the state of Santa Catarina
449 is amongst the richest and most developed areas of the country (www.atlasbrasil.org.br);
450 therefore, no site-specific air quality issues typical of the other low- and middle-income
451 countries typically occur. These factors could explain the lower concentrations in both indoor
452 and outdoor environments reported in Figure 1 and Table 4.

453 **3.3 Total daily dose received by the population**

454 Table 5 shows the values of particle surface area doses received by the populations of the cities
455 investigated, and of the western cities reported in our previous paper, as obtained by combining
456 (i) exposure data (summarized in section 3.2), (ii) time-activity patterns of the typical
457 population (summarized in section 3.1), and (iii) inhalation rates characteristic of the age and
458 activity performed. In particular, Table 5 reports the particle surface area doses (mm^2) as a
459 function of population, sex, and microenvironment. Additionally, the contributions to the daily
460 dose (%) and dose intensity ratios ($\text{mm}^2 \text{min}^{-1}$) for each microenvironment are also reported.

461 The dose evaluated for western populations, as discussed in detail in our previous paper (Pacitto
462 et al., 2018) and summarized here for the sake of brevity, highlighted that: (i) particle surface
463 area doses were significantly larger than 1000mm^2 for people living in Cassino (Italy) and
464 Guilford (UK), around 700mm^2 for those residing in Barcelona (Spain) and Brisbane
465 (Australia), and around 100mm^2 for Lund inhabitants (Sweden), (ii) the higher doses estimated
466 for Cassino (Italy) and Guilford (UK) are due to the higher contribution of the Cooking &
467 Eating microenvironment; whereas the lower doses estimated for the Lund population are due
468 to the widespread use of mechanical ventilation systems, (iii) in general, the main contributions
469 to the daily dose are due to the Cooking & Eating and Indoor day microenvironments (from
470 58% to 90% of the total daily dose) for all the investigated western populations, and (iv) the

471 total daily dose of the western populations is not greatly affected by the outdoor air quality,
472 indeed the contribution of the Outdoor day and Transport microenvironments is generally
473 negligible.

474 For the low- and middle-income populations investigated, the total doses received vary over a
475 wide range – approximately 1300 mm² for Cairo (Egypt), 1100 mm² for Accra (Ghana),
476 750 mm² for Nur-Sultan (Kazakhstan), and 450 mm² for Florianopolis (Brazil). No significant
477 differences were found between males and females in terms of total daily dose, although some
478 differences were recognized in some specific microenvironments due to different time-activity
479 patterns characteristic of females and males. As an example, as was also recognized for western
480 populations, women typically receive higher doses in the Cooking & Eating microenvironment
481 and lower doses in the Transport and Working microenvironments.

482 Apart from the absolute dose values, an interesting aspect is the different contribution of each
483 microenvironment to the total daily dose with respect to the western populations previously
484 analyzed. The most relevant difference is related to the Cooking & Eating microenvironment,
485 for which the contribution to the total daily dose ranges from 8% to 14%, which is much lower
486 than those evaluated for western populations (up to > 50) due to the shorter time spent in this
487 microenvironment (see section 3.1) and, in some cases, to the lower concentrations measured
488 therein as discussed in section 3.2 (e.g. the possible effect of different cooking habits and of the
489 favorable climatic conditions supporting home airing). In fact, the contribution of the Cooking
490 & Eating microenvironment was quite similar to that of the Transportation microenvironment.
491 In particular, the median dose received by residents of Cairo during transport activities was
492 even larger than 200 mm² (contributing up to 20% of the daily dose), mostly due to the larger
493 concentrations of people who are exposed during Transportation (Figure 1), as confirmed by
494 the highest dose intensity ratio amongst the population analyzed (2.96 mm² min⁻¹). A high dose
495 intensity ratio in the Transportation microenvironment (> 1 mm² min⁻¹) was also recognized in
496 Nur-Sultan (Kazakhstan) due to the abovementioned high concentrations, although the shorter
497 time spent performing such activities led to a limited dose contribution.

498 In section 3.2 we highlighted the role of the outdoor air quality in the overall exposure of
499 residents of Cairo and Nur-Sultan, which also affects some indoor microenvironments where
500 no other particle sources are typically adopted. This can also be partly recognized in terms of
501 doses: people in Cairo also received significant doses in the Sleeping & Resting
502 microenvironment (about 200 mm²), the Outdoor day microenvironment (> 200 mm²), and the
503 Working microenvironment (up to > 300 mm²). A similar situation was recognized for people
504 in Nur-Sultan, except for the Outdoor day contribution which was reduced by the short time

505 spent outdoors. The absolute doses received in the Indoor day microenvironment were lower
506 than 200 mm² (with the exception of Accra) and lower than the typical doses received by
507 western populations (once again the authors point out that Lund should be dealt with separately
508 due to the widespread adoption of ventilation systems). This is mainly due to the shorter time
509 spent indoors in low- and middle-income countries in comparison with the high-income western
510 populations (section 3.1). As mentioned above, the Accra (Ghana) situation is quite different;
511 indeed, the doses received both in the Sleeping & Resting (about 300 mm²) and Indoor day
512 (> 400 mm²) microenvironments are significantly larger than those received by the other low-
513 and middle-income populations. This is due to high exposure related to the use of anti-mosquito
514 products indoors. The dose fraction received by the Accra population in these two
515 microenvironments is roughly 70%, whereas the contribution of each of the other
516 microenvironments is lower than 10%.

517 Finally, for the Florianopolis population, the above-discussed low exposure in all the
518 microenvironments (likely related to both favorable climatic conditions and good outdoor air
519 quality) resulted in very low absolute dose values both in indoor- and outdoor-related
520 microenvironments.

521 Summarizing, the data shown provide an important insight into the exposure to submicron
522 particles and the related dose received by non-western populations that are characterized by
523 different lifestyles, habits, socioeconomic conditions, climates, and outdoor air quality. The
524 data revealed that, despite their higher air quality awareness and socioeconomic status,
525 populations in some western countries receive particle doses larger than non-western
526 populations. This is strongly related to the time spent indoors and to the adoption of inadequate
527 ventilation practices. In contrast, in the warmer low- and middle-income countries analyzed in
528 this study, where the populations spend less time indoors and home airing is likely allowed by
529 favorable climatic conditions, the indoor contribution is reduced. Low- and middle-income
530 countries present their own site-specific concerns; for example, the dose of some populations
531 is significantly affected by outdoor air quality and transport microenvironments, and the use of
532 anti-mosquito products in Accra (Ghana) highlighted an unpredicted huge contribution to the
533 night time exposure dose. Thus, the dose data provided here importantly highlight some critical
534 exposures not easily recognizable through typical outdoor air quality measurements, and thus
535 provide the opportunity to search for solutions. Additionally, dose data could be combined with
536 particle chemical composition data to perform a-priori estimates of the lung cancer risk of the
537 population as performed in our previous paper on the Italian population (Buonanno et al.,
538 2015a).

539 In order to properly use the data here shown the authors want to point out some limitations of
540 the study that should be consider when generalizing the results or transferring them to other
541 contexts. Since the study aims at providing a population-based dose (i.e. a dose characteristics
542 of the entire population), a key question is “how the concentrations measured in the
543 experimental campaigns are representative of the typical exposures of the populations?” We
544 tried to cope with this problem enrolling volunteers with different habits in order to obtain
545 exposure data in very different situations, microenvironment by microenvironment;
546 nonetheless, no one knows if all the possible exposure situations were covered. This critical
547 aspect could have minimized enrolling a huge number of volunteers, but this would have been
548 not workable for both practical and economic reasons. Further aspects to be contemplated are
549 the possible effects on the exposures of (i) the urban development (e.g. the outdoor exposure in
550 small towns could be different from main cities), and (ii) the season (colder or warmer periods
551 would have increased or reduced the exposure in some microenvironments). These critical
552 aspects related to the selection of a representative sample should be handled in future studies
553 along with further technical issues, e.g. the uncertainty of the portable counters, in order to
554 provide the uncertainty budget of the dose values. Nonetheless, such limitations do not
555 undermine the importance and significance of the findings here obtained, since for the very first
556 time the particle dose received by several populations were determined (including non-western
557 world countries) and the contribution of each microenvironment, habit, lifestyle, and climate
558 were highlighted.

559 **Conclusions**

560 In the present study, the daily doses in terms of particle surface area received by four low- and
561 middle-income populations were assessed and compared to those characteristics of high-
562 income western populations obtained in our previous study. In particular, the total daily doses
563 received by populations living in Greater Cairo (Egypt), Greater Accra (Ghana), Florianopolis
564 (Brazil), and Nur-Sultan (Kazakhstan) were estimated based on submicron particle
565 concentration measurements at a personal scale (i.e. direct exposure assessment method), time-
566 activity patterns (obtained from national statistics institutes) and inhalation rates. The
567 contribution of the main microenvironments was also investigated.

568 The total doses received by the low- and middle-income populations here investigated were
569 roughly 1300 mm² for Egyptians, 1100 mm² for Ghanaians, 750 mm² for Kazakhstanis, and
570 450 mm² for Brazilians. The different concentration levels with respect to the western

571 populations, along with different time-activity patterns (e.g. generally low- and middle-income
572 populations spend less time indoors), resulted in different contributions of the
573 microenvironments to the daily dose. Indeed, the Cooking & Eating microenvironment
574 contributed just 8%–14% to the daily dose for the four populations analyzed (for western
575 populations it contributed up to > 50%), whereas the outdoor-related microenvironments
576 increased their contributions (e.g. the Transportation microenvironment reached up to 20% for
577 Egyptians). Finally, in contrast with the other populations, the main contribution to the total
578 daily dose for Ghanaians came from the Sleeping & Resting and Indoor day
579 microenvironments, possibly because of the use of anti-mosquito products indoors.
580 To summarize, the different dose values and relative contributions amongst the populations
581 analyzed clearly highlight the effect of the lifestyle of the populations; indeed, the different
582 time-activity patterns and site-specific habits strongly affected the dose data. The study also
583 showed that, in contrast with the western population study, outdoor air quality (and
584 consequently the climate) can also play a role in the total daily dose received by low- and
585 middle-income populations.

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598 **References**

- 599 Auchincloss, A. H., Diez Roux, A. V., Dvorchak, J. T., Brown, P. L., Barr, R. G., Daviglus, M. L., . . .
600 O'Neill, M. S. (2008). Associations between Recent Exposure to Ambient Fine Particulate
601 Matter and Blood Pressure in the Multi-Ethnic Study of Atherosclerosis (MESA).
602 *Environmental Health Perspectives*, 116(4), 486-491.
- 603 Bakó-Biró, Z., Clements-Croome, D. J., Kochhar, N., Awbi, H. B., & Williams, M. J. (2012).
604 Ventilation rates in schools and pupils' performance. *Building and Environment*, 48(0), 215-
605 223.
- 606 Beelen, R., Stafoggia, M., Raaschou-Nielsen, O., Andersen, Z. J., Xun, W. W., Katsouyanni, K., . . .
607 Hoek, G. (2014). Long-term exposure to air pollution and cardiovascular mortality: An analysis
608 of 22 European cohorts. *Epidemiology*.
- 609 Bekö, G., Kjeldsen, B. U., Olsen, Y., Schipperijn, J., Wierzbicka, A., Karotki, D. G., . . . Clausen, G.
610 (2015). Contribution of various microenvironments to the daily personal exposure to ultrafine
611 particles: Personal monitoring coupled with GPS tracking. *Atmospheric Environment*, 110, 122-
612 129.
- 613 Brook, R. D., Jerrett, M., Brook, J. R., Bard, R. L., & Finkelstein, M. M. (2008). The relationship
614 between diabetes mellitus and traffic-related air pollution. *J Occup Environ Med*, 50(1), 32-38.
- 615 Buonanno, G., Dell'Isola, M., Stabile, L., & Viola, A. (2010). Critical Aspects of the Uncertainty Budget
616 in the Gravimetric PM Measurements. *Measurement*, 44, 139-147.
- 617 Buonanno, G., Giovinco, G., Morawska, L., & Stabile, L. (2011). Tracheobronchial and alveolar dose
618 of submicrometer particles for different population age groups in Italy. *Atmospheric*
619 *Environment*, 45(34), 6216-6224.
- 620 Buonanno, G., Giovinco, G., Morawska, L., & Stabile, L. (2015). Lung cancer risk of airborne particles
621 for Italian population. *Environmental Research*, 142, 443-451.
- 622 Buonanno, G., Jayaratne, E. R., Morawska, L., & Stabile, L. (2014). Metrological Performances of a
623 Diffusion Charger Particle Counter for Personal Monitoring. *Aerosol and Air Quality Research*,
624 14, 156-167.
- 625 Buonanno, G., Morawska, L., Stabile, L., Wang, L., & Giovinco, G. (2012). A comparison of
626 submicrometer particle dose between Australian and Italian people. *Environmental Pollution*,
627 169, 183-189.
- 628 Buonanno, G., Stabile, L., & Morawska, L. (2014). Personal exposure to ultrafine particles: the influence
629 of time-activity patterns. *Science of the Total Environment*, 468-469, 903-907.
- 630 Buonanno, G., Stabile, L., Morawska, L., Giovinco, G., & Querol, X. (2017). Do air quality targets
631 really represent safe limits for lung cancer risk? *Science of the Total Environment*, 580, 74-82.
- 632 Buteau, S., & Goldberg, M. S. (2016). A structured review of panel studies used to investigate
633 associations between ambient air pollution and heart rate variability. *Environ Res*, 148, 207-
634 247.
- 635 CAPMAS. (2018). Annual Report of Licensed Vehicles. Central Agency for Public Mobilization and
636 Statistics.

- 637 Central Agency for Public Mobilization and Statistics of Egyptian Arabic Republic (2017). International
638 classification of active time use statistics.
- 639 Central Agency for Public Mobilization and Statistics of Egyptian Arabic Republic (2016). Time use
640 survey In the Arab Republic of Egypt 2015.
- 641 Chen, S.Y., Feng, Z., and Yi, X. (2017). A general introduction to adjustment for multiple comparisons.
642 *Journal of Thoracic Disease*, 9(6), 1725–1729.
- 643 EEAA. (2016). Egyptian State of the Environment Report. Egyptian Environmental Affairs Agency.
- 644 EU Directive 2008/50/EC of the European Parliament and of the Council of 21 May 2008 on ambient
645 air quality and cleaner air for Europe, 2008, L 152/1 C.F.R. (2008).
- 646 Fernández-Iriarte, A., Amato, F., Moreno, N., Pacitto, A., Reche, C., Marco, E., . . . Moreno, T. (2020).
647 Chemistry and sources of PM_{2.5} and volatile organic compounds breathed inside urban
648 commuting and tourist buses. *Atmospheric Environment*, 223, 117234.
- 649 Fierz, M., Houle, C., Steigmeier, P., & Burtscher, H. (2011). Design, Calibration, and Field Performance
650 of a Miniature Diffusion Size Classifier. *Aerosol Science and Technology*, 45(1), 1-10.
- 651 Franck, U., Odeh, S., Wiedensohler, A., Wehner, B., & Herbarth, O. (2011). The effect of particle size
652 on cardiovascular disorders - The smaller the worse. *Science of the Total Environment*, 409(20),
653 4217-4221.
- 654 Fuoco, F., Stabile, L., Buonanno, G., Scungio, M., Manigrasso, M., & Frattolillo, A. (2017).
655 Tracheobronchial and Alveolar Particle Surface Area Doses in Smokers. *Atmosphere*, 8(1), 19.
- 656 Giechaskiel, B., Alföldy, B., & Drossinos, Y. (2009). A metric for health effects studies of diesel exhaust
657 particles. *Journal of Aerosol Science*, 40(8), 639-651.
- 658 Goel, A., & Kumar, P. (2014). A review of fundamental drivers governing the emissions, dispersion and
659 exposure to vehicle-emitted nanoparticles at signalised traffic intersections. *Atmospheric
660 Environment*, 97, 316-331.
- 661 Hogarh, J. N., Antwi-Agyei, P., & Obiri-Danso, K. (2016). Application of mosquito repellent coils and
662 associated self-reported health issues in Ghana. *Malaria Journal*, 15(1), 61.
663 doi:10.1186/s12936-016-1126-8
- 664 IBGE. (2018). *Instituto Brasileiro De Geografia E Estatística - 2018*.
- 665 International Agency for Research on Cancer. (2013). *IARC: Outdoor air pollution a leading
666 environmental cause of cancer deaths*. Retrieved from Lyon/Geneva, 17 October 2013:
- 667 International Commission on Radiological Protection. (1994). Human respiratory tract model for
668 radiological protection. A report of a Task Group of the International Commission on
669 Radiological Protection. *Annals of the ICRP*, 24(1–3), 1-482.
- 670 Kaur, S., Nieuwenhuijsen, M. J., & Colvile, R. N. (2005). Pedestrian exposure to air pollution along a
671 major road in Central London, UK. *Atmospheric Environment*, 39(38), 7307-7320.
- 672 Kruskal, W. H., & Wallis, W. A. (1952). Use of Ranks in One-Criterion Variance Analysis. *Journal of
673 the American Statistical Association*, 47(260).

- 674 Kumar, P., Ketzel, M., Vardoulakis, S., Pirjola, L., & Britter, R. (2011). Dynamics and dispersion
675 modelling of nanoparticles from road traffic in the urban atmospheric environment—A review.
676 *Journal of Aerosol Science*, 42(9), 580-603.
- 677 Kumar, P., Morawska, L., Birmili, W., Paasonen, P., Hu, M., Kulmala, M., . . . Britter, R. (2014).
678 Ultrafine particles in cities. *Environment International*, 66, 1-10.
- 679 Kumar, P., Pirjola, L., Ketzel, M., & Harrison, R. M. (2013). Nanoparticle emissions from 11 non-
680 vehicle exhaust sources – A review. *Atmospheric Environment*, 67, 252-277.
- 681 Marra, J., Voetz, M., & Kiesling, H.-J. (2010). Monitor for detecting and assessing exposure to airborne
682 nanoparticles. *Journal of Nanoparticle Research*, 12(1), 21-37.
- 683 Mehta, S., Wolf, J., Bruce, N. G., Smith, K. R., Adair-Rohani, H., Bonjour, S., . . . Rehfuess, E. A.
684 (2013). Solid fuel use for household cooking: country and regional estimates for 1980-2010.
685 *Environmental Health Perspectives*, 121(7).
- 686 Meteorology, N. I. o. (2019). Brazil Annual Historical Data.
- 687 Morawska, L., Afshari, A., Bae, G. N., Buonanno, G., Chao, C. Y., Hanninen, O., . . . Wierzbicka, A.
688 (2013). Indoor aerosols: from personal exposure to risk assessment. *Indoor Air*, 23(6), 462-487.
- 689 Moreno, T., Pacitto, A., Fernández, A., Amato, F., Marco, E., Grimalt, J., . . . Querol, X. (2019). Vehicle
690 interior air quality conditions when travelling by taxi. *Environmental Research*, 172, 529-542.
- 691 Naseri, M., Jouzizadeh, M., Tabesh, M., Malekipirbazari, M., Gabdrashova, R., Nurzhan, S., . . . Amouei
692 Torkmahalleh, M. (2019). The impact of frying aerosol on human brain activity.
693 *NeuroToxicology*, 74, 149-161.
- 694 Neft, I., Scungio, M., Culver, N., & Singh, S. (2016). Simulations of aerosol filtration by vegetation:
695 Validation of existing models with available lab data and application to near-roadway scenario.
696 *Aerosol Science and Technology*, 50(9), 937-946.
- 697 Nyarku, M., Buonanno, G., Ofosu, F., Jayaratne, R., Mazaheri, M., & Morawska, L. (2019).
698 Schoolchildren's personal exposure to ultrafine particles in and near Accra, Ghana.
699 *Environment International*, 133, 105223.
- 700 Pacitto, A., Amato, F., Moreno, T., Pandolfi, M., Fonseca, A., Mazaheri, M., . . . Querol, X. (2019).
701 Effect of ventilation strategies and air purifiers on the children's exposure to airborne particles
702 and gaseous pollutants in school gyms. *Science of the Total Environment*.
- 703 Pacitto, A., Stabile, L., Moreno, T., Kumar, P., Wierzbicka, A., Morawska, L., & Buonanno, G. (2018).
704 The influence of lifestyle on airborne particle surface area doses received by different Western
705 populations. *Environmental Pollution*, 232, 113-122.
- 706 Pacitto, A., Stabile, L., Russo, S., & Buonanno, G. (2020). Exposure to submicron particles and
707 estimation of the dose received by children in school and non-school environments.
708 *Atmosphere*, 11(5). doi:10.3390/ATMOS11050485
- 709 Peel, M. C., Finlayson, B. L., & McMahon, T. A. (2007). Updated world map of the Köppen-Geiger
710 climate classification. *Hydrol. Earth Syst. Sci.*, 11(5), 1633-1644.
- 711 Power, M. C., Weisskopf, M. G., Alexeeff, S. E., Coull, B. A., Spiro, A., 3rd, & Schwartz, J. (2011).
712 Traffic-related air pollution and cognitive function in a cohort of older men. *Environ Health*
713 *Perspect*, 119(5), 682-687.

- 714 Puzzolo, E., Zerriffi, H., Carter, E., Clemens, H., Stokes, H., Jagger, P., . . . Petach, H. (2019). Supply
715 Considerations for Scaling Up Clean Cooking Fuels for Household Energy in Low- and Middle-
716 Income Countries. *GeoHealth*, 3(12), 370-390.
- 717 Rizza, V., Stabile, L., Buonanno, G., & Morawska, L. (2017). Variability of airborne particle metrics in
718 an urban area. *Environmental Pollution*, 220, 625-635.
- 719 Rizza, V., Stabile, L., Vistocco, D., Russi, A., Pardi, S., & Buonanno, G. (2019). Effects of the exposure
720 to ultrafine particles on heart rate in a healthy population. *Science of the Total Environment*,
721 650, 2403-2410. doi:10.1016/j.scitotenv.2018.09.385
- 722 Sager, T. M., & Castranova, V. (2009). Surface area of particle administered versus mass in determining
723 the pulmonary toxicity of ultrafine and fine carbon black: comparison to ultrafine titanium
724 dioxide. *Particle and Fibre Toxicology*, 6, 15-15.
- 725 Schmid, O., Möller, W., Semmler-Behnke, M., A. Ferron, G., Karg, E., Lipka, J., . . . Stoeger, T. (2009).
726 Dosimetry and toxicology of inhaled ultrafine particles. *Biomarkers*, 14(SUPPL.1), 67-73.
- 727 Schmid, O., & Stoeger, T. (2016). Surface area is the biologically most effective dose metric for acute
728 nanoparticle toxicity in the lung. *Journal of Aerosol Science*, 99, 133-143.
- 729 Schweizer, C., Edwards, R. D., Bayer-Oglesby, L., Gauderman, W. J., Ilacqua, V., Juhani Jantunen, M.,
730 . . . Kunzli, N. (2006). Indoor time-microenvironment-activity patterns in seven regions of
731 Europe. *J Expos Sci Environ Epidemiol*, 17(2), 170-181.
- 732 Scungio, M., Arpino, F., Cortellessa, G., & Buonanno, G. (2015). Detached eddy simulation of turbulent
733 flow in isolated street canyons of different aspect ratios. *Atmospheric Pollution Research*, 6(2),
734 351-364.
- 735 Scungio, M., Arpino, F., Stabile, L., & Buonanno, G. (2013). Numerical Simulation of Ultrafine Particle
736 Dispersion in Urban Street Canyons with the Spalart-Allmaras Turbulence Model. *Aerosol and
737 Air Quality Research*, 13, 1423-1437.
- 738 Scungio, M., Rizza, V., Stabile, L., Morawska, L., & Buonanno, G. (2020). Influence of methodology
739 on the estimation of the particle surface area dose received by a population in all-day activities.
740 *Environmental Pollution*, 266, 115209.
- 741 Scungio, M., Vitanza, T., Stabile, L., Buonanno, G., & Morawska, L. (2017). Characterization of particle
742 emission from laser printers. *Science of the Total Environment*, 586, 623-630.
- 743 Ghana Statistical Service (2012). Ghana Time-Use Survey 2009.
- 744 Ghana Statistical Service (2014). Ghana Living Standards Survey Round 6.
- 745 Ghana Statistical Service (2015). Labour Force Report.
- 746 Stabile, L., Arpino, F., Buonanno, G., Russi, A., & Frattolillo, A. (2015). A simplified benchmark of
747 ultrafine particle dispersion in idealized urban street canyons: A wind tunnel study. *Building
748 and Environment*, 93, Part 2, 186-198
- 749 Stabile, L., Dell'Isola, M., Russi, A., Massimo, A., & Buonanno, G. (2017). The effect of natural
750 ventilation strategy on indoor air quality in schools. *Science of the Total Environment*, 595, 894-
751 902.
- 752 Stabile, L., Fuoco, F. C., Marini, S., & Buonanno, G. (2015). Effects of the exposure to indoor cooking-
753 generated particles on nitric oxide exhaled by women. *Atmospheric Environment*, 103, 238-246.

- 754 Stephens, B., & Siegel, J. A. (2012). Penetration of ambient submicron particles into single-family
755 residences and associations with building characteristics. *Indoor Air*, 22(6), 501-513.
- 756 Stoeger, T., Reinhard, C., Takenaka, S., Schroepfel, A., Karg, E., Ritter, B., . . . Schulz, H. (2006).
757 Instillation of six different ultrafine carbon particles indicates a surface area threshold dose for
758 acute lung inflammation in mice. *Environmental Health Perspectives*, 114(3), 328-333.
- 759 Tomassetti, L., Torre, M., Tratzi, P., Paolini, V., Rizza, V., Segreto, M., & Petracchini, F. (2020).
760 Evaluation of air quality and mobility policies in 14 large Italian cities from 2006 to 2016.
761 *Journal of Environmental Science and Health - Part A Toxic/Hazardous Substances and*
762 *Environmental Engineering*, 55(7), 886-902.
- 763 Vasques, T., Brancher, M., Hoinaski, L., & Lisboa, H. (2017). *Assessment of daily and annual*
764 *concentrations of PM10 in the city of Florianópolis*.
- 765 Wallace, L., & Ott, W. (2011). Personal exposure to ultrafine particles. *Journal of Exposure Science*
766 *and Environmental Epidemiology*, 21(1), 20-30.
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Tables

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773**Table 1** Main meteorological parameters characteristic of the cities under investigation during the experimental campaign: data are expressed as maximum (max) average (avg) and minimum (min) values (Meteorology, 2019).

City	Temperature (°C)			Humidity (%)			Wind speed (m/s)	Precipitation (mm)
	Max	Avg	Min	Max	Avg	Min	Avg	Total
Cairo (Egypt)	20	16	12	73	53	33	7	0
Accra (Ghana)	32	29	24	95	78	23	1	54
Florianopolis (Brazil)	33	25	16	92	74	30	2	24
Nur-Sultan (Kazakhstan)	39	13	-8	93	53	13	2	Not available

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Table 2 Classification of the activities performed by citizens in the six main microenvironments.

Microenvironment	Activities
Transportation	Trip and use of time not specified, round-trip to work
Working	Non-industrial workplaces, Profitable work, Main and secondary job, Working-connected activities, Studying, School, Institute or University, Voluntary job and meetings, Voluntary work in an organization
Cooking & Eating	Eating and drinking (including home and restaurant), Cooking
Outdoor day	Gardening and animal care, Construction and restoration, Sport and outdoor activities, Physical workout, Productive exercise, Sports-connected activities, Collecting firewood, Fetching, Running errands, Collecting food from the garden
Indoor day	Personal care, Other personal care, Studying not specified, Studying in free time, Activities for home and family not specified, Housework, Clothes care and folding, Purchasing goods and services, Home maintenance, Baby care, Helping adult family members, Helping other family members, Active activities, Social activities and entertainment, Social life, Entertainment and culture, Inactivity, Hobbies and computer science, Art and hobbies, Computing, Playing, Media, Reading, Watching TV, DVD or videos, Listening to the radio or recording
Sleeping & Resting	Sleeping and resting

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777**Table 3** Time-activity pattern data (daily time spent, in minutes, for each microenvironment expressed as mean values) and the contribution of single microenvironments to the total daily time (expressed in percentages) of the investigated populations as a function of sex.

Parameter	Microenvironment	Greater Cairo (Egypt)		Greater Accra (Ghana)		Nur-Sultan (Kazakhstan)		Florianopolis (Brazil)	
		Male	Female	Male	Female	Male	Female	Male	Female
Time (min)	Sleeping & Resting	541	576	490	524	517	499	476	444
	Indoor day	177	188	318	311	163	295	253	287
	Outdoor day	116	145	107	120	37	24	132	112
	Transport	87	62	73	70	80	40	98	99
	Working	462	396	379	318	505	427	396	392
	Cooking & Eating	57	73	70	93	137	154	85	107
Time contribution (%)	Sleeping & Resting	37.6%	40.0%	34.1%	36.5%	35.9%	34.7%	33.1%	30.8%
	Indoor day	12.3%	13.1%	22.1%	21.7%	11.3%	20.5%	17.6%	19.9%
	Outdoor day	8.1%	10.1%	7.4%	8.4%	2.6%	1.7%	9.2%	7.7%
	Transport	6.0%	4.3%	5.1%	4.9%	5.6%	2.8%	6.8%	6.8%
	Working	32.1%	27.5%	26.4%	22.1%	35.1%	29.7%	27.5%	27.2%
	Cooking & Eating	4.0%	5.1%	4.9%	6.5%	9.5%	10.7%	5.9%	7.4%

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Table 4 Median values and 5th-95th percentile ranges of particle number (PN) and lung-deposited surface area (LDSA) concentrations, and average particle size (D_p) to which the nine populations are exposed in each microenvironment. Data of the four cities investigated in the present paper (Cairo, Accra, Florianopolis, and Nur-Sultan) and of the five western cities (Barcelona, Cassino, Guilford, Brisbane, and Lund) investigated in our previous paper (Pacitto et al., 2018), and here re-adapted accordingly, are summarized.

Population	City	Parameter	Microenvironments					
			Sleeping & Resting	Indoor day	Outdoor day	Transport	Working	Cooking & Eating
Low- and middle-income populations; this study	Cairo (Egypt)	PN (part. cm ⁻³)	1.83×10 ⁴ (0.33-5.75×10 ⁴)	2.61×10 ⁴ (0.45-13.7×10 ⁴)	5.14×10 ⁴ (0.47-48.8×10 ⁴)	17.3×10 ⁴ (1.96-162×10 ⁴)	4.31×10 ⁴ (1.30-37.6×10 ⁴)	14.5×10 ⁴ (2.16-174×10 ⁴)
		D_p (nm)	53 (29-81)	43 (22-70)	32 (15-52)	30 (18-49)	39 (26-56)	33 (16-64)
		LDSA (μm ² cm ⁻³)	5.87×10 ¹ (0.61-18.2×10 ¹)	6.83×10 ¹ (0.68-39.7×10 ¹)	10.9×10 ¹ (0.51-85.9×10 ¹)	28.8×10 ¹ (0.41-216×10 ¹)	9.23×10 ¹ (0.23-88.6×10 ¹)	25.9×10 ¹ (0.49-270×10 ¹)
	Accra (Ghana)	PN (part. cm ⁻³)	4.18×10 ⁴ (1.36-10.2×10 ⁴)	2.94×10 ⁴ (1.18-18.5×10 ⁴)	1.04×10 ⁴ (0.57-10.9×10 ⁴)	2.06×10 ⁴ (0.68-16.6×10 ⁴)	0.99×10 ⁴ (0.53-10.1×10 ⁴)	4.53×10 ⁴ (1.65-40.4×10 ⁴)
		D_p (nm)	43 (29-56)	48 (38-75)	55 (42-66)	49 (39-67)	55 (39-73)	47 (34-60)
		LDSA (μm ² cm ⁻³)	9.45×10 ¹ (3.74-25.3×10 ¹)	8.90×10 ¹ (3.50-39.7×10 ¹)	3.80×10 ¹ (1.67-37.6×10 ¹)	8.14×10 ¹ (2.09-58.1×10 ¹)	3.68×10 ¹ (1.35-31.7×10 ¹)	14.4×10 ¹ (4.69-99.7×10 ¹)
	Florianopolis (Brazil)	PN (part. cm ⁻³)	0.983×10 ⁴ (0.78-1.82×10 ⁴)	1.21×10 ⁴ (0.89-2.60×10 ⁴)	1.31×10 ⁴ (0.89-2.71×10 ⁴)	1.33×10 ⁴ (0.86-7.21×10 ⁴)	1.18×10 ⁴ (0.9-2.78×10 ⁴)	1.94×10 ⁴ (0.98-38.5×10 ⁴)
		D_p (nm)	53 (29-56)	42 (29-56)	44 (29-56)	39 (29-56)	44 (29-56)	37 (29-56)
		LDSA (μm ² cm ⁻³)	3.11×10 ¹ (1.9-4.82×10 ¹)	2.94×10 ¹ (2.0-5.19×10 ¹)	3.26×10 ¹ (1.81-6.33×10 ¹)	3.25×10 ¹ (1.72-10.6×10 ¹)	3.26×10 ¹ (2.17-6.63×10 ¹)	4.10×10 ¹ (1.98-50.7×10 ¹)
	Nur-Sultan (Kazakhstan)	PN (part. cm ⁻³)	1.10×10 ⁴ (0.27-51.0×10 ⁴)	2.13×10 ⁴ (0.31-17.6×10 ⁴)	1.98×10 ⁴ (0.31-20.9×10 ⁴)	5.93×10 ⁴ (0.43-52.9×10 ⁴)	3.55×10 ⁴ (0.30-28.8×10 ⁴)	4.13×10 ⁴ (0.58-47.4×10 ⁴)
		D_p (nm)	55 (20-123)	50 (19-86)	41 (18-74)	37 (19-64)	45 (13-73)	39 (16-78)
		LDSA (μm ² cm ⁻³)	3.52×10 ¹ (0.91-151×10 ¹)	5.01×10 ¹ (1.10-28.5×10 ¹)	3.94×10 ¹ (1.01-43.7×10 ¹)	11.8×10 ¹ (1.35-70.9×10 ¹)	6.94×10 ¹ (0.92-31.6×10 ¹)	8.31×10 ¹ (1.47-89.9×10 ¹)
Western populations; data adapted from Pacitto et al. (2018)	Barcelona (Spain)	PN (part. cm ⁻³)	0.66×10 ⁴ (0.38-2.25×10 ⁴)	1.51×10 ⁴ (0.65-7.75×10 ⁴)	0.93×10 ⁴ (0.47-2.20×10 ⁴)	2.79×10 ⁴ (1.43-9.46×10 ⁴)	1.42×10 ⁴ (0.32-1.84×10 ⁴)	3.37×10 ⁴ (0.99-28.2×10 ⁴)
		D_p (nm)	50 (40-57)	48 (40-57)	59 (40-79)	65 (44-108)	49 (46-59)	44 (36-53)
		LDSA (μm ² cm ⁻³)	1.70×10 ¹ (1.05-6.54×10 ¹)	4.10×10 ¹ (1.05-6.54×10 ¹)	3.05×10 ¹ (1.62-6.63×10 ¹)	11.2×10 ¹ (5.30-30.4×10 ¹)	3.87×10 ¹ (1.06-4.94×10 ¹)	8.19×10 ¹ (3.49-6.54×10 ¹)
	Cassino (Italy)	PN (part. cm ⁻³)	1.24×10 ⁴ (0.95-1.91×10 ⁴)	1.04×10 ⁴ (0.35-8.33×10 ⁴)	0.51×10 ⁴ (0.23-1.03×10 ⁴)	1.02×10 ⁴ (0.35-5.38×10 ⁴)	0.57×10 ⁴ (0.29-1.59×10 ⁴)	35.6×10 ⁴ (1.64-107×10 ⁴)
		D_p (nm)	61 (49-81)	62 (49-81)	110 (78-189)	88 (56-136)	97 (79-119)	49 (31-129)
		LDSA (μm ² cm ⁻³)	4.30×10 ¹ (3.91-5.25×10 ¹)	3.42×10 ¹ (3.91-5.25×10 ¹)	3.10×10 ¹ (2.22-5.23×10 ¹)	5.18×10 ¹ (1.79-18.2×10 ¹)	2.90×10 ¹ (1.76-10.2×10 ¹)	69.1×10 ¹ (7.85-367×10 ¹)
	Guilford (UK)	PN (part. cm ⁻³)	0.38×10 ⁴ (0.16-1.71×10 ⁴)	3.09×10 ⁴ (0.62-35.9×10 ⁴)	1.29×10 ⁴ (0.52-2.20×10 ⁴)	1.42×10 ⁴ (0.31-5.55×10 ⁴)	6.05×10 ³ (0.28-0.89×10 ⁴)	6.01×10 ⁴ (0.41-56.5×10 ⁴)
		D_p (nm)	82 (63-113)	54 (30-113)	63 (33-96)	56 (27-99)	85 (65-125)	73 (27-121)
		LDSA (μm ² cm ⁻³)	1.51×10 ¹ (0.93-6.74×10 ¹)	9.25×10 ¹ (1.97-79.1×10 ¹)	4.38×10 ¹ (2.03-7.21×10 ¹)	4.24×10 ¹ (1.10-16.1×10 ¹)	2.51×10 ¹ (1.31-5.59×10 ¹)	18.9×10 ¹ (1.63-387×10 ¹)
	Brisbane (Australia)	PN (part. cm ⁻³)	0.49×10 ⁴ (0.12-1.20×10 ⁴)	1.36×10 ⁴ (0.46-4.73×10 ⁴)	1.51×10 ⁴ (0.57-4.55×10 ⁴)	1.88×10 ⁴ (0.61-7.28×10 ⁴)	0.97×10 ⁴ (0.41-3.21×10 ⁴)	7.07×10 ⁴ (0.79-74.8×10 ⁴)
		D_p (nm)	59 (43-91)	43 (15-97)	38 (26-47)	40 (29-63)	54 (34-69)	66 (34-95)
		LDSA (μm ² cm ⁻³)	1.35×10 ¹ (0.46-5.53×10 ¹)	3.14×10 ¹ (0.93-6.74×10 ¹)	3.14×10 ¹ (1.23-8.83×10 ¹)	4.49×10 ¹ (1.42-15.0×10 ¹)	2.68×10 ¹ (1.18-10.1×10 ¹)	22.7×10 ¹ (3.41-135×10 ¹)
Lund (Sweden)	PN (part. cm ⁻³)	0.77×10 ³ (0.44-1.37×10 ³)	0.86×10 ³ (0.42-5.35×10 ³)	2.88×10 ³ (1.38-18.9×10 ³)	3.13×10 ³ (0.83-6.92×10 ³)	1.48×10 ³ (0.69-8.96×10 ³)	22.6×10 ³ (1.44-112×10 ³)	
	D_p (nm)	72 (49-103)	62 (39-88)	47 (33-70)	55 (36-85)	58 (29-83)	35 (25-80)	
	LDSA (μm ² cm ⁻³)	0.32×10 ¹ (0.15-0.58×10 ¹)	0.33×10 ¹ (0.16-1.21×10 ¹)	0.79×10 ¹ (0.40-4.31×10 ¹)	0.98×10 ¹ (0.23-2.27×10 ¹)	0.47×10 ¹ (0.28-1.72×10 ¹)	4.56×10 ¹ (0.53-24.4×10 ¹)	

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Table 5 Dose, dose intensity, and contribution to the daily dose as a function of the population, sex (female/male, F/M) and microenvironments for the four cities investigated in the present paper (Cairo, Accra, Florianopolis, and Nur-Sultan) and for the five western cities (Barcelona, Cassino, Guilford, Lund, Brisbane) investigated in our previous paper (Pacitto et al., 2018) (and here re-adapted accordingly).

City	Parameter	Microenvironment						Total daily dose (mm ²)
		Sleeping & Resting	Indoor day	Outdoor day	Transport	Working	Cooking & Eating	
Cairo (Egypt)	Dose, δ_{SA} (mm ²)	217 / 204	209 / 196	257 / 206	182 / 257	286 / 334	154 / 112	1306 / 1309
	Dose intensity ratio (mm ² min ⁻¹)	0.38 / 0.38	1.11 / 1.11	1.77 / 1.77	2.96 / 2.96	0.72 / 0.72	2.11 / 1.96	
	Contribution to the daily dose (%)	17 / 16	16 / 15	20 / 16	14 / 20	22 / 25	11 / 8	
Accra (Ghana)	Dose, δ_{SA} (mm ²)	318 / 297	450 / 460	74 / 66	58 / 61	92 / 109	116 / 87	1108 / 1081
	Dose intensity ratio (mm ² min ⁻¹)	0.61 / 0.61	1.45 / 1.45	0.62 / 0.62	0.83 / 0.83	0.29 / 0.29	1.25 / 1.25	
	Contribution to the daily dose (%)	29 / 28	41 / 43	7 / 6	5 / 6	8 / 10	10 / 8	
Florianopolis (Brazil)	Dose, δ_{SA} (mm ²)	89 / 95	137 / 121	59 / 70	33 / 33	100 / 101	38 / 30	456 / 450
	Dose intensity ratio (mm ² min ⁻¹)	0.2 / 0.2	0.48 / 0.48	0.53 / 0.53	0.33 / 0.33	0.26 / 0.26	0.35 / 0.35	
	Contribution to the daily dose (%)	19 / 20	30 / 27	13 / 16	7 / 7	22 / 23	8 / 7	
Nur-Sultan (Kazakhstan)	Dose, δ_{SA} (mm ²)	113 / 117	241 / 133	16 / 24	48 / 97	232 / 275	105 / 94	755 / 739
	Dose intensity ratio (mm ² min ⁻¹)	0.23 / 0.23	0.81 / 0.81	0.64 / 0.64	1.21 / 1.21	0.54 / 0.54	0.69 / 0.69	
	Contribution to the daily dose (%)	15 / 16	32 / 18	2 / 3	6 / 13	31 / 37	14 / 13	
Barcelona (Spain)	Dose, δ_{SA} (mm ²)	60 / 61	337 / 290	24 / 40	97 / 114	47 / 68	151 / 88	715 / 658
	Dose intensity ratio (mm ² min ⁻¹)	0.12 / 0.12	0.75 / 0.75	0.53 / 0.53	1.24 / 1.25	0.30 / 0.30	0.79 / 0.63	
	Contribution to the daily dose (%)	8 / 9	47 / 44	3 / 6	14 / 17	7 / 10	21 / 14	
Cassino (Italy)	Dose, δ_{SA} (mm ²)	127 / 123	269 / 239	39 / 40	34 / 35	69 / 77	779 / 587	1317 / 1099
	Dose intensity ratio (mm ² min ⁻¹)	0.29 / 0.28	0.59 / 0.56	0.50 / 0.48	0.55 / 0.53	0.26 / 0.25	5.88 / 5.21	
	Contribution to the daily dose (%)	10 / 11	20 / 22	3 / 4	3 / 3	5 / 7	59 / 53	
Guilford (UK)	Dose, δ_{SA} (mm ²)	61 / 59	585 / 452	27 / 45	30 / 36	24 / 30	683 / 497	1408 / 1188
	Dose intensity ratio (mm ² min ⁻¹)	0.12 / 0.13	1.39 / 1.03	0.45 / 0.56	0.35 / 0.56	0.12 / 0.10	4.12 / 4.05	
	Contribution to the daily dose (%)	4 / 5	42 / 40	2 / 4	2 / 3	2 / 3	48 / 45	
Brisbane (Australia)	Dose, δ_{SA} (mm ²)	35 / 42	325 / 283	21 / 24	6 / 12	40 / 88	224 / 173	650 / 620
	Dose intensity ratio (mm ² min ⁻¹)	0.09 / 0.10	0.46 / 0.50	0.43 / 0.47	0.44 / 0.44	0.31 / 0.34	1.64 / 1.66	
	Contribution to the daily dose (%)	5 / 7	50 / 46	3 / 4	1 / 2	6 / 14	35 / 28	
Lund (Sweden)	Dose, δ_{SA} (mm ²)	11 / 11	22 / 23	14 / 13	6 / 5	9 / 12	41 / 32	102 / 93
	Dose intensity ratio (mm ² min ⁻¹)	0.02 / 0.02	0.06 / 0.06	0.14 / 0.13	0.10 / 0.09	0.04 / 0.04	0.34 / 0.30	
	Contribution to the daily dose (%)	10 / 11	22 / 23	14 / 14	5 / 5	9 / 12	40 / 35	

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Figure Captions

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Figure 1 Box-plots of a) particle number concentrations, and b) lung deposited surface area concentrations experienced in each microenvironment by the populations of the four cities investigated in the present paper (Cairo, Accra, Florianopolis, and Nur-Sultan, reported as yellow boxes) and of the five western cities (Barcelona, Cassino, Guilford, Lund, and Brisbane, reported as white boxes) investigated in our previous paper (Pacitto et al.,2018) (and here re-adapted accordingly). Data not statistically different amongst the different populations are also indicated ($p > 0.01$) microenvironment by microenvironment as resulting from the statistical analysis performed.

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